NEAR THREATENED SPECIES

DWARF CASSOWARY *Casuarius bennetti* occurs in New Guinea (Papua, formerly Irian Jaya, **Indonesia** and **Papua New Guinea**) and, presumably as a long-established introduction, on New Britain, where it is a forest species occurring into the mountains and occasionally to the treeline at 3,600 m (Coates 1985, Beehler *et al.* 1986). Although probably tolerant of moderate habitat degradation, logging opens up previously inaccessible areas to hunters; despite heavy hunting pressure, it remains relatively common over a wide altitudinal range (Coates 1985, Beehler *et al.* 1986, B. M. Beehler *in litt.* 2000, A. Mack *in litt.* 1999). It is judged to have a substantial population and to be declining more slowly than the other larger and more lowland cassowaries *Casuarius*. **Criteria nearly met:** A1b,d; A2b,d.

PYGMY CORMORANT *Phalacrocorax pygmeus* breeds in (all data for pairs, and unless otherwise indicated all information from Crivelli *et al.* 1996) **Bulgaria** (20–180), **Greece** (1,250–1,310) (Kazantzidis and Nazirides 1999), **Italy** (30–50) (M. Passarella *in litt.* 1999), **Moldova** (30–500), **Hungary**, **Romania** (4,000–7,000), **Turkey** (1,000–1,500) (Eken and Magnin in press), **Slovakia**, **Yugoslavia** (1,000–1,200), **F.Y.R.O. Macedonia**, **Croatia**, **Ukraine** (20–320), **Russia**, **Iran** (20–30), **Azerbaijan** (14,749 estimated in 1986), **Kazakhstan**, **Tajikistan**, **Turkmenistan**, and **Uzebekistan**, plus, possibly, south-east **Iraq**, and it winters primarily in **Albania**, Greece, Yugoslavia, F.Y.R.O. Macedonia, Turkey, **Cyprus**, Iraq, Iran, Azerbaijan and also **Israel**, Bulgaria, Romania and Syria. There is one record from **Pakistan** (Grimmett *et al.* 1998). The estimated population is 22,345–27,055 pairs (Crivelli *et al.* in press), considerably higher than the 13,000 pairs estimated in 1996 (Crivelli *et al.* 1996). Trends are unclear because significant declines may be occurring in the Azerbaijan population, given the potential threats, but numbers there are uncertain (Crivelli *et al.* 1996). In south-east Europe, conservation measures have reduced the most important threats and resulted in relatively stable populations with some local increases (Crivelli *et al.* in press). It occurs in reedbeds, transition zones between reedbeds and open waters, extensively grazed or mowed shores and wet meadows (Willems and de Vries 1998) and, in winter, in coastal wetlands, along rivers, and sometimes on inland lakes. The preferred nesting habitat is willow *Salix* trees but, in Azerbaijan, birds breed mainly in *Tamarix* (Crivelli *et al.* in press). Key threats are wetland drainage and degradation, persecution by fishermen, and changes to hydrological regimes (Crivelli *et al.* in press, Eken and Magnin in press, Kazantzidis and Nazirides 1999). A European action plan was published in 1996 (Crivelli *et al.* 1996). **Criteria nearly met:** A1c; A2c.

ORIENTAL DARTER *Anhinga melanogaster* occurs in **Pakistan** (fairly widespread irregular year-round visitor to Sind and Punjab, locally resident: Roberts 1991–1992), **India** (widespread resident, locally common in Assam, current status poorly known but apparently declining: Grimmett *et al.* 1998), **Nepal** (uncommon resident and non-breeding visitor: Grimmett *et al.* 1998), **Sri Lanka** (common resident in dry lowlands, scarce visitor elsewhere: Grimmett *et al.* 1998), **Bangladesh** (local resident in northern and coastal regions: Grimmett *et al.* 1998), **Myanmar** (previously a widespread resident, now scarce to locally fairly common in the south, status uncertain elsewhere: Robson 2000), **Thailand** (formerly widespread, now very rare and possibly no longer breeds: Lekagul and Round 1991, Robson 2000), **Laos** (previously widespread and numerous but numbers have plummeted with only a few sporadic recent records: Duckworth *et al.* 1999), **Vietnam** (previously widespread and common breeder, but now rare after a major decline: Buckton *et al.* 1999, Eames and Tordoff in prep.), **Cambodia** (abundant in early 1960s with flocks "totalling several thousand" observed on the Mekong: Thomas 1964; currently a local resident in small numbers: Mundkur *et al.* 1995a, Robson 2000), Peninsular **Malaysia** (vagrant in west, possibly former resident: Wells 1999, Robson 2000), **Brunei** (Mann 1988), **Indonesia** (locally common breeder on Borneo, Java and Sulawesi, vagrant to other islands in the Lesser Sundas and Moluccas: MacKinnon and Phillipps 1993, Coates and Bishop 1997). The species is generally uncommon and declining throughout Asia,

although an estimate of 4,000 for South Asia (Rose and Scott 1997) may be too low. It inhabits shallow inland wetlands including lakes, rivers, swamps and reservoirs, as well as estuaries, tidal inlets, mangroves and coastal lagoons (del Hoyo *et al.* 1992), ascending to 1,400 m, at least in India and Java (MacKinnon and Phillipps 1993, Grimmett *et al.* 1998). In common with many other Asian waterbirds, it is primarily threatened by habitat loss, disturbance (at feeding grounds and colonies), hunting and pollution (del Hoyo *et al.* 1992). **Criteria nearly met:** A1b,c,d; A2b,c,d.

FOREST BITTERN *Zonerodius heliosylus* occurs throughout New Guinea (Papua, formerly Irian Jaya, **Indonesia** and **Papua New Guinea**) and on the adjacent islands of Salawati and Aru, where it frequents streams, pools and swamps in forest to 1,430 m; there are very few recent records but it is such a reclusive species that it is impossible to assess its true population (Coates 1985, Beehler *et al.* 1986, K. D. Bishop *in litt.* 1994, D. Gibbs *in litt.* 1994, Eastwood 1998). Although it is not hunted (B. M. Beehler *in litt.* 1994), and there are still huge areas of suitable habitat left, it may be threatened by the extensive logging of lowland forest (I. Burrows *in litt.* 1994, R. Burrows *in litt.* 1994), especially as it occurs along watercourses (A. Mack *in litt.* 1999). **Criteria nearly met:** A1c; A2c; C1.

PAINTED STORK *Mycteria leucocephala* occurs in **Pakistan** (scarce; mainly confined to the Indus delta region: Roberts 1991–1992), **Nepal** (rare in terai; mainly a summer visitor: Grimmett *et al.* 1998), **India** (widespread and locally common resident: del Hoyo *et al.* 1992, Grimmett *et al.* 1998), **Bangladesh** (former resident, now a straggler to coastal regions: Grimmett *et al.* 1998), **Sri Lanka** (locally abundant, particularly in the dry zone: del Hoyo *et al.* 1992, Grimmett *et al.* 1998), mainland **China** (previously a common summer visitor in the south, probably breeding, but now rare and possibly extinct: Cheng Tso-hsin 1987, MacKinnon and Phillipps 2000), **Myanmar** (former resident in central region and visitor throughout; current status unknown but clearly rare: Robson 2000), **Thailand** (previously common breeder in the south, now on the verge of extinction, small numbers recorded sporadically elsewhere: Lekagul and Round 1991, Wells 1999, Robson 2000), **Laos** (previously widespread, now rare: Duckworth *et al.* 1999), **Vietnam** (formerly widespread resident, now a rare non-breeding visitor: Robson 2000, Eames and Tordoff in prep.), **Cambodia** (local resident, minimum of several hundred pairs breeding at Tonle Sap: Mundkur *et al.* 1995a, Robson 2000) and Peninsular **Malaysia** (previously regular, now a vagrant: Wells 1999, Robson 2000). It frequents freshwater marshes, lakes and reservoirs, flooded fields, rice paddies, freshwater swamp forest, riverbanks, intertidal mudflats and saltpans (Grimmett *et al.* 1998, Robson 2000). There are an estimated 15,000 individuals in South Asia and fewer than 10,000 in South-East Asia, with populations declining throughout (Rose and Scott 1997). Although it is thus considered "one of the most numerous and secure of Asian storks" (del Hoyo *et al.* 1992), this is more a reflection of the rarity and endangerment of most other storks in the region than of the security of this species. The increasing impact of habitat loss, disturbance, pollution and hunting of adults and collection of eggs and nestlings from colonies is cause for concern (del Hoyo *et al.* 1992, Mundkur *et al.* 1995a). **Criteria nearly met:** A1b,c,d; A2b,c,d.

BLACK-NECKED STORK *Ephippiorhynchus asiaticus* occurs in **Pakistan** (previously frequent in lower Sind, breeding in the Indus delta until the 1970s, now a straggler: Roberts 1991–1992, del Hoyo *et al.* 1992), **Nepal** (rare resident and winter visitor to the terai: Grimmett *et al.* 1998), **India** (still a widespread resident, but now generally rare and local: Grimmett *et al.* 1998), **Bangladesh** (former resident, now a vagrant: Grimmett *et al.* 1998), **Sri Lanka** (fewer than 50 mature individuals resident, principally in the dry lowlands: Grimmett *et al.* 1998, Hoffmann 1998), **Myanmar** (formerly a widespread resident, current status unclear

but certainly scarce: Robson 2000), **Thailand** (formerly quite widespread, now a rare resident in southern Thailand, almost extinct: Lekagul and Round 1991, Wells 1999, Robson 2000), **Laos** (previously a widespread non-breeding visitor, probably breeding in the south, but now extremely rare: Duckworth *et al*. 1999), **Cambodia** (previously fairly common: Delacour and Jabouille 1931; very few recent records, with very small numbers breeding around Tonle Sap: Mundkur *et al*. 1995a), **Indonesia** (apparently once present in the Sundaic region, but now extinct there: MacKinnon and Phillipps 1993; population >650 in south Papua [formerly Irian Jaya]: del Hoyo *et al*. 1992), **Papua New Guinea** (very local, but occasionally not uncommon: Coates 1985), **Australia** (relatively large population in the north, possibly spreading southwards: Marchant and Higgins 1990). The combined populations of South and South-East Asia are thought not to exceed 400 individuals (Rose and Scott 1997), while the former is in steep decline (Rahmani 1987d, 1989a,b) and the latter has dwindled disastrously to the brink of extinction (Duckworth *et al*. 1999, Robson 2000). It is probably stable or even increasing in Papua and Australia (Luthin 1987, Marchant and Higgins 1990, Rose and Scott 1997), although the situation needs further review (E. Dorfman verbally 2000). It frequents freshwater marshes, lakes, pools in open forest, large rivers, occasionally mangroves and rarely coastal mudflats, up to 1,200 m (Grimmett *et al*. 1998, Robson 2000), being highly susceptible to disturbance and tending to prefer areas least visited by humans (del Hoyo *et al*. 1992). It is threatened by a variety of factors across its range, including drainage of wetlands, felling of nest trees, development, encroachment of agriculture or aquaculture, overfishing, overgrazing, hunting and excessive capture for zoos (Rahmani 1987d, 1989a,b, del Hoyo *et al*. 1992, Hoffmann 1992, 1998). **Criterion nearly met:** C1.

BLACK-HEADED IBIS *Threskiornis melanocephalus* occurs in **Japan** (scarce non-breeding visitor: del Hoyo *et al*. 1992), **South Korea** (Won 1993), mainland **China** (probably breeds in Heilongjiang, but this is not confirmed; non-breeding visitors are rare along the east and south coasts, occasionally inland to Sichuan and Yunnan: MacKinnon and Phillipps 2000), **Hong Kong** (regular winter visitor in small numbers with occasional summer records: Chalmers 1986), **Pakistan** (scarce resident, principally in the Indus delta region: Roberts 1991–1992), **Nepal** (frequent resident and summer visitor to the south-east: Baral 1993a, Grimmett *et al*. 1998), **India** (widespread and locally common in the west, scarce in the east: Grimmett *et al*. 1998; possibly increasing locally due to spread of man-made wetlands: Rahmani 1995a), **Sri Lanka** (common resident in the lowlands, particularly the dry zone: Grimmett *et al*. 1998), **Bangladesh** (local visitor to coastal regions and the north-east: Grimmett *et al*. 1998), **Philippines** (rare non-breeding visitor to the south: del Hoyo *et al*. 1992), **Myanmar** (uncommon but widespread non-breeding visitor, 730 counted in 1991: del Hoyo *et al*. 1992, Robson 2000), **Thailand** (formerly common resident, now uncommon winter visitor: Lekagul and Round 1991, del Hoyo *et al*. 1992, Robson 2000), **Laos** (only one record, a single bird prior to 1950: David-Beaulieu 1949–1950), **Vietnam** (previously an abundant breeder, now a few large colonies remaining: del Hoyo *et al*. 1992, Eames and Tordoff in prep.), **Cambodia** (fairly common resident in early 1960s: Thomas 1964; now scarce and local with small numbers breeding around Tonle Sap: Mundkur *et al*. 1995a, Robson 2000), Peninsular **Malaysia** (formerly occurred and probably bred in the west, but few recent records: del Hoyo *et al*. 1992, Wells 1999), **Indonesia** (scarce non-breeding visitor to Sumatra and north Borneo, possibly breeding in Sumatra with c.2,000 birds estimated, numerous breeding colonies in Java early in the twentieth century, but now local and declining: del Hoyo *et al*. 1992, MacKinnon and Phillipps 1993). While the East Asian population is extremely small (<100), those in South-East Asia and South Asia probably number between 10,000 and 25,000 individuals each (Rose and Scott 1997). It inhabits freshwater marshes, lakes, rivers, flooded grasslands, paddyfields, tidal creeks, mudflats, saltmarshes and coastal lagoons, usually in extreme lowlands, but occasionally up to 950 m, tending to be nomadic in response to water

levels and feeding conditions (Grimmett *et al*. 1998). It is vulnerable to drainage, disturbance, pollution, agricultural conversion, hunting and collection of eggs and nestlings from colonies (del Hoyo *et al*. 1992). **Criteria nearly met:** A1b,c,d; A2b,c,d.

LESSER FLAMINGO *Phoenicopterus minor* breeds mainly in the Rift Valley lakes of East Africa in **Ethiopia**, **Kenya** and **Tanzania**, with three smaller breeding congregations in West Africa, southern Africa, and, in Asia, **India** (breeding in the Great and Little Ranns of Kutch, Gujarat); when not breeding, it ranges occasionally into virtually every sub-Saharan country and from the Arabian Peninsula to **Pakistan** (del Hoyo *et al.* 1992, Grimmett *et al.* 1998). The global population is c.5,000,000, including c.150,000 (stable or even increasing) in Asia, but declines have been suggested for much of Africa (Simmons 1996, Rose and Scott 1997; see *Waterbirds* 23 [2000], special publication 1). It breeds in huge colonies on large, undisturbed alkaline and saline lakes, and is adapted to respond to local environmental changes in sites by moving elsewhere, and thus depends on a network of suitable areas; however, proposed soda-ash mining and hydroelectric power schemes in what is in effect the only breeding site in the Rift Valley, Lake Natron in Tanzania, could cause rapid overall population declines and permanently alter the ecosystem on which the species depends (L. Bennun *in litt.* 1999, Mari and Collar 2000). Other threats include land claim, water pollution, and disturbance (R. E. Simmons *in litt.* 1998, S. J. Tyler *in litt.* 1999). CMS Appendix II. **Criterion nearly met:** A2c.

EMPEROR GOOSE *Anser canagica* is restricted to the Bering Sea, breeding in coastal saltmarshes in Arctic and subarctic Alaska, **USA**, and extreme north-east coastal **Russia**, and wintering principally along ice-free coasts of the Aleutian Islands and, in smaller numbers, in the Alaska Peninsula, with very few reaching as far south as California (Petersen *et al.* 1994). Its population in Alaska declined precipitously from 139,000 in 1964 to 42,000 in 1986, but subsequently recovered to c.68,000 in 1993; the factors causing such fluctuations are poorly understood, but subsistence hunting in Alaska and coastal oil pollution are considered to be contributory (Petersen *et al*. 1994). **Criterion nearly met:** A1a.

FERRUGINOUS DUCK *Aythya nyroca* breeds in Europe, Asia (east to mainland **China** and south to **India**) and North Africa, the wintering range overlapping with the breeding range but extending to the Middle East, West and north-east Africa and South-East Asia (del Hoyo *et al.* 1992). The main part of the population occurs in Asia, although quantitative data are lacking. An estimate for North Africa and Asia of 10,000 individuals in 1991 appears too low (Perennou *et al.* 1994, Callaghan in press). Recent surveys have found high numbers, perhaps into the tens of thousands, in Inner Mongolia (Xing Lianlian verbally 1998), and it is apparently common on the Tibetan Plateau, mainland **China** (Scott 1993), and either scarce or locally common in winter in **Pakistan**, **India**, **Nepal**, **Bhutan**, **Bangladesh**, **Myanmar** and **Thailand** (Grimmett *et al.* 1998, Robson 2000). It is also recorded from **Mongolia** (A. Bräunlich *in litt.* 1999) and western **Russia** (Dement'ev and Gladkov 1951–1954). Other large winter counts have been made in **Azerbaijan** (9,000 birds), **Turkmenistan** (20,833 birds) (del Hoyo *et al*. 1992) and **Uzbekistan** (7,000 birds) (Kashkarov and Mukhina 1997), but numbers of breeding birds have declined in **Kazakhstan** (Berezovikov and Samusev 1998) and Uzbekistan (Kashkarov and Mukhina 1997). In Europe, there have been some declines in breeding populations, notably in **Ukraine** where numbers have fallen from a possible 70,000 to an estimated 1,500 pairs since the 1950s, and also in **Moldova**, **Poland**, **Hungary**, and **Spain** (Callaghan in press). The European population is now an estimated 13,000–24,000 pairs, and the key threat is the loss of its wetland habitat, of well vegetated shallow pools, including extensively managed fishponds, although hunting is also a serious threat (Callaghan in press). **Criteria nearly met:** A1c; A2c.

WHITE-TAILED EAGLE *Haliaeetus albicilla* has its strongholds in **Norway** (1,600–1,800 pairs) (I. J. Øien *in litt.* 1999) and **Russia**, and important populations in south-west Greenland (to **Denmark**), **Sweden**, **Poland** and **Germany** (301 pairs) (Hauff 1998). Smaller numbers breed in **Iceland**, **United Kingdom**, **Finland**, **Estonia**, **Latvia**, **Lithuania**, **Belarus**, **Austria**, **Czech Republic**, **Slovakia**, **Slovenia**, **Yugoslavia**, **Bulgaria**, **Romania**, **Hungary**, **Moldova**, **Greece**, **Turkey**, **Iran**, **Armenia**, **Georgia**, **Azerbaijan**, **Ukraine**, **Kazakhstan**, **Turkmenistan**, **Mongolia**, mainland **China** and **Japan** (Collar *et al.* 1994; also Cheng Tso-hsin 1987, Brazil 1991). Small numbers winter south to **North Korea** and **South Korea** (Gore and Won 1971, Tomek 1999), **Taiwan** (China) (Wang Chia-hsiong *et al.* 1991), **Pakistan** (where it was once fairly common: see Hume 1872–1873), **India**, especially Assam (Ali and Ripley 1968–1998, Choudhury 2000c), and **Nepal**, while it is a vagrant to **Bangladesh** (Grimmett *et al.* 1998) and **Thailand** (Robson 2000). It is mainly migratory in the north and east of its breeding range but sedentary elsewhere. The population is estimated at 5,000–7,000 pairs but the size of the Russian population is poorly known. During the nineteenth and first half of the twentieth centuries numbers declined dramatically and its range contracted in western Europe. This trend has been reversed in the north-west of its breeding range since the 1970s. Numbers are thought to be stable or slightly increasing in countries of the former Soviet Union but are still declining in south-east Europe. Loss and degradation of wetlands, increasing human disturbance, and the indiscriminate use of poisons are continuing threats. Modern forestry methods have reduced the availability of suitable habitat. It is susceptible to environmental pollution with the accumulation of mercury, organochlorine and other pesticide residues leading to reduced breeding success. (Thiollay 1996). **Criterion nearly met:** C2a.

LESSER FISH-EAGLE *Ichthyophaga humilis* occurs in **India** (restricted to Himalayan foothills and north-east; declining in range and population: Samant *et al.* 1995, Grimmett *et al.* 1999), **Nepal** (rare and local in lowlands: Grimmett *et al.* 1999), **Bhutan** (very rare at lower altitudes: Inskipp *et al.* 1999a), mainland **China** (rare visitor to Hainan: MacKinnon and Phillipps 2000), **Myanmar** (widespread, scarce to locally fairly common: Robson 2000), **Thailand** (rare in west and south: Round 1988a, Lekagul and Round 1991, Robson 2000), **Laos** (small numbers persist in several catchments, although fragmentation of populations and their small size renders them vulnerable to local extinction: Thewlis *et al.* 1998, Duckworth *et al.* 1999), **Cambodia** (four recent records: C. M. Poole *in litt.* 2001), **Vietnam** (rare to locally fairly common in west Tonkin and south Annam: Robson 2000), Peninsular **Malaysia** (previously common: Chasen 1939a; now scarce to locally fairly common but declining: Wells 1999, Robson 2000) and East Malaysia (MacKinnon and Phillipps 1993), **Brunei** (Mann 1988), **Indonesia** (uncommon in Sumatra and Borneo: MacKinnon and Phillipps 1993; locally common in south-east Sulawesi, uncommon to rare elsewhere, and in the Sula islands and Buru: Coates and Bishop 1997). It frequents large forested rivers and wetlands in the lowlands and foothills up to 2,400 m, but usually below 1,000 m (MacKinnon and Phillipps 1993, Grimmett *et al.* 1998, Robson 2000). Loss of forest habitat along rivers, siltation, overfishing and increasing human disturbance of waterways are important threats that are causing widespread declines (Round 1988a, del Hoyo *et al.* 1994). It is also declining in Uttar Pradesh, India, partly because of pesticide use (Naoroji 1995, 1997) and this is presumably relevant throughout much of its range. **Criteria nearly met:** A1b,c,e; A2b,c,e; C1; C2a.

GREY-HEADED FISH-EAGLE *Ichthyophaga ichthyaet* occurs in **India** (widespread and locally frequent in the north-east, scarce and local in the peninsula: Samant *et al.* 1995, Grimmett *et al.* 1998), **Nepal** (now rare and local, mainly below 250 m: Grimmett *et al.* 1998), **Sri Lanka** (rare in the dry lowlands: Grimmett *et al.* 1998), **Bangladesh** (widely

distributed but uncommon and local: Grimmett *et al*. 1998), **Myanmar** (rare to scarce resident: Robson 2000), **Philippines** (formerly quite common in the north and east, now rare and apparently declining: Collar *et al*. 1999), **Thailand** (formerly a widespread resident, now absent from the north and centre, rare and local in the south: Round 1988a, Lekagul and Round 1991, Wells 1999, Robson 2000), **Laos** (now rare: Duckworth *et al*. 1999), **Vietnam** (scarce in the south, disappearing from the north: Robson 2000), **Cambodia** (scarce: Robson 2000), Peninsular **Malaysia** (previously common, now uncommon and sparse, perhaps 40 pairs remaining: Wells 1999), East Malaysia (MacKinnon and Phillipps 1993), **Singapore** (scarce: Robson 2000), **Brunei** (Mann 1988), and the Greater Sundas and Sulawesi, **Indonesia** (widely distributed but uncommon in Sumatra and Borneo, but now very rare in Java: MacKinnon and Phillipps 1993). Although widely distributed, historical and even recent records are difficult to interpret, in South-East Asia at least, owing to identification difficulties between this species and Lesser Fish-eagle *Ichthyophaga humilis* (Duckworth *et al*. 1999). It is found near slow-moving rivers and streams, lakes, reservoirs and tidal lagoons in wooded country, usually in lowlands but ascending locally to 1,525 m (Grimmett *et al*. 1998, Robson 2000). The most pertinent threats are the loss of undisturbed wetlands, overfishing, siltation, pollution and persecution (del Hoyo *et al*. 1992). **Criteria nearly met:** A1b,c; A2b,c; C1.

CINEREOUS VULTURE *Aegypius monachus* breeds in **Spain**, **Bulgaria**, **Greece**, **Turkey**, **Armenia**, **Azerbaijan**, **Georgia**, **Ukraine**, **Russia**, **Uzbekistan**, **Kazakhstan**, **Tajikistan**, **Turkmenistan**, **Kyrgyzstan**, **Iran**, **Afghanistan**, north **India**, **Russia** (Dement'ev and Gladkov 1951–1954), **Mongolia** and mainland **China**, with a small re-introduced population in **France**, with possible occasional breeding in **Portugal**, F.Y.R.O. Macedonia and Albania, and wintering areas in **Sudan**, **Pakistan**, north-west India, **Nepal**, **Bangladesh** (vagrant: Grimmett *et al*. 1998), **Myanmar**, **North Korea** and **South Korea** (del Hoyo *et al.* 1994, Heredia 1996). There are a few records from South-East Asia in **Thailand** (Lekagul and Round 1991), **Cambodia** (O'Sullivan 1994), **Vietnam** (Vo Quy 1971), and Peninsular **Malaysia** (Wells 1999). It has a small total population (probably over 10,000 birds) with declines occurring in some parts of its range (Heredia 1996). In Europe, populations are declining in Ukraine and the Caucasus (180–200 pairs), but are increasing in Spain (1,050–1,150 pairs), and are stable in Greece (20–21 pairs) (Tucker and Heath 1994, Heredia 1996, V. Galushin *in litt.* 1999, WWF Greece 1999). The population in Turkey (100–500 pairs) was considered stable but recent surveys in western Turkey suggest that there may be fewer breeding pairs than previously thought (Heredia *et al.* 1997). There are probably over 1,000 pairs in the Asian part of the former Soviet Union and a further 1,760 pairs in mainland China (Ye Xiaoti 1991). It is threatened by habitat alteration in its breeding areas, particularly the destruction of native forests, forest fires, poisoning and a shortage of food as a result of changes in traditional farming practices, and many birds are trapped or shot in China for trade in their feathers (Heredia 1996, V. Galushin *in litt.*1999, WWF Greece 1999). A European action plan was published in 1996. (Heredia 1996). **Criterion nearly met:** C1.

RED-HEADED VULTURE *Sarcogyps calvus* occurs in **Pakistan** (previously regular, now a rare straggler: Roberts 1991–1992), **Nepal** (uncommon: Grimmett *et al*. 1998), **India** (sparsely distributed and declining, now rare or absent from some areas, e.g. parts of Gujarat and the north-eastern states, but still fairly common in the Western Himalayan foothills: Grimmett *et al*. 1998), **Bangladesh** (rare in the north-west: Grimmett *et al*. 1998), **Myanmar** (former resident, current status unknown: Robson 2000), mainland **China** (very rare in south-west Yunnan and possibly occurs in south-east Tibet: Cheng Tso-hsin 1987, MacKinnon and Phillipps 2000), **Thailand** (rare resident in remote portions of the west, now absent elsewhere: Lekagul and Round 1991, Robson 2000), **Laos** (previously widespread and common, but

now rare and restricted to the south: Duckworth *et al.* 1999), **Vietnam** (previously regular in central regions, now rare after a major decline: Robson 2000, Eames and Tordoff in prep.), **Cambodia** (previously common, now uncommon and restricted to the north-east: Robson 2000), and Peninsular **Malaysia** (previously locally common in north, now absent: Wells 1999, Robson 2000). It frequents open country (often near human habitation), well-wooded hills and dry deciduous forest with rivers, usually below 2,500 m (Grimmett *et al.* 1998, Robson 2000). Historical reports indicate that it was widespread and generally abundant, but it has undergone a serious diminution in numbers and range in recent decades: in Thailand, for example, it was once "common everywhere" in the north, particularly in the plains, but also up to the highest summits (Deignan 1945) before virtually disappearing from the country after an extraordinary decline (Scott 1989). The disappearance of vultures from Asia is linked to a suite of factors, amongst which can be listed the demise of wild ungulates, the intensification of agriculture, increased sophistication of waste disposal techniques, direct persecution and the spread of avian diseases (del Hoyo *et al.* 1992; see Threats under White-rumped Vulture *Gyps bengalensis*). **Criteria nearly met:** A1a,b,c,d,e; A2a,b,c,d,e.

SOUTH NICOBAR SERPENT-EAGLE *Spilornis klossi* (here treated as separate from "Nicobar [or Small] Serpent-eagle *S. minimus*", which is instead provisionally placed with *S. cheela*) is endemic to the islands of Great Nicobar (including Pulo Kunji), Little Nicobar and Menchal in the South Nicobar island group, Nicobar islands, **India**, where it is most frequently found in the canopy of forest (Richmond 1903, Abdulali 1967, 1978, Sankaran 1998, K. Sivakumar verbally 1999). There is some confusion over records in 1993, when it was reported to be "probably one of the rarest raptors in the country" and "rarely seen in the Great Nicobar island" (Samant *et al.* 1995), but this has not been the impression of other fieldworkers (Sankaran 1995d, 1998, K. Sivakumar verbally 1999). Increased settlement of the islands has led to increased pressure on natural resources, and planned development projects could severely affect the habitat of this species (Stattersfield *et al.* 1998). **Criterion nearly met:** C1.

ANDAMAN SERPENT-EAGLE *Spilornis elgini* is endemic to South Andaman island, **India**, where it is a common resident in inland forest clearings and hillsides with scattered trees, appearing to be ecologically separated from Crested Serpent-eagle *S. cheela*, which inhabits coastal forests on the same island (Davidar *et al.* 1996, Grimmett *et al.* 1998). Although forest remains extensive on the Andamans, loss and fragmentation of cover continues and is perhaps accelerating (Pande *et al.* 1991). The human population on larger islands is rising rapidly and habitat is consequently under mounting pressure from agriculture, grazing and logging (Whitaker 1985, Curson, 1989, Sinha 1992, Stattersfield *et al.* 1998). Hunting is also apparently common on the islands (Stattersfield *et al.* 1998) and may affect this species. **Criteria nearly met:** B1+2a,b,c,d,e; C1; C2b.

PALLID HARRIER *Circus macrourus* breeds primarily in the steppes of Asiatic **Russia**, **Kazakhstan** and north-west mainland **China**; small populations breed in **Moldova**, **Ukraine** and **Turkey** (del Hoyo *et al.* 1994, Tucker and Heath 1994). A few pairs breed in taiga and forest-tundra, north of its main breeding range (Kuznetsov 1994, V. V. Morozov *in litt.* 1999). The population is estimated at 20,000 pairs, having shown marked declines and range contractions: it has declined in southern Ukraine, south-west Russia and south-central Siberia (del Hoyo *et al.* 1994, Tucker and Heath 1994). A minority winter in south-east Europe, North Africa and the Middle East but most migrate to the Afrotropics (**Sudan**, **Eritrea**, **Djibouti**, **Ethiopia**, **Somalia**, **Kenya**, **Uganda**, **Rwanda**, **Burundi**, **Tanzania**, **Malawi**, **Zambia**, **Zimbabwe**, **Mozambique**, **Chad**, **Niger**, **Mali**, **Senegal**, **Gambia**, **Sierra**

Leone, **Liberia**, **Ivory Coast**, **Ghana**, **Togo**, **Benin**, **Nigeria**, **Cameroon**, **Central African Republic**, **Democratic Republic of Congo**, **Angola**, **Namibia**, **Botswana** and **South Africa**) and the Indian subcontinent (**Afghanistan**, **Pakistan**, **India**, **Sri Lanka**, **Nepal**, **Bangladesh** and **Myanmar**) (del Hoyo *et al.* 1994). It has occurred as a migrant in **Mongolia** (Vaurie 1964), and there is apparently an old sight record from **Vietnam** (Robson 2000). Scrub, savanna and wetlands are used in winter (J. Brouwer *in litt.*). It is primarily threatened by the destruction and degradation of steppe grasslands through conversion to arable agriculture, intensive grazing of wet pastures and the clearance of shrubs and tall weeds (Tucker and Heath 1994). It may be affected by pesticides and rodenticides but this requires further research (R. Simmons *in litt.* 1999). In South Africa, wintering birds may be affected by poisoning, grassland destruction and persecution (Barnes 2000). CMS Appendix II. **Criteria nearly met:** A1c,d,e; A2c,d,e.

SMALL SPARROWHAWK *Accipiter nanus* is restricted to the Sulawesi Endemic Bird Area (but not the East Peninsula), **Indonesia**, where it inhabits montane forest at 550–2,000 m (Stattersfield *et al*. 1998; also Coates and Bishop 1997). It is rare on Buton (Caterall undated) and uncommon to rare on Sulawesi (del Hoyo *et al*. 1994), with very few localities recorded (six mentioned in Coates and Bishop 1997). Forest loss in the lower-lying areas of Sulawesi has been rampant in recent decades (see Threats under Lombopatang Flycatcher *Ficedula bonthaina*) and this species must be in a commensurately steep decline in the areas in question. **Criterion nearly met:** C1.

DORIA'S GOSHAWK *Megatriorchis doriae* is a little-known New Guinea endemic (Papua, formerly Irian Jaya, **Indonesia** and **Papua New Guinea**) (Beehler *et al.* 1986, including Batanta island (Bishop 1986). It is a forest species, sometimes found in mangrove and semi-deciduous forest to 1,400 m (Coates 1985, Beehler *et al.* 1986). It is rarely recorded, partly because of its unobtrusive habits, with for instance only one record in seven years observation at Tabubil (P. Gregory *in litt.* 1999). Its general tolerance of logging is poorly known but it has been seen repeatedly in logged forest near Port Moresby (P. Gregory *in litt.* 1999). Although it appears to be widespread and not threatened (K. D. Bishop *in litt.* 1994, J. M. Diamond *in litt.* 1987), it may be suffering a rapid population decline and more data are required on its population size and tolerance of logged forest. **Criteria nearly met:** A1c; A2c; C1.

GURNEY'S EAGLE *Aquila gurneyi* is a wide-ranging species of the Moluccas and New Guinea (Papua, formerly Irian Jaya, **Indonesia** and **Papua New Guinea**). It appears to be widespread in a variety of forested habitats to 1,000 m, occasionally to 1,500 m (Coates 1985, Beehler *et al.*1986). There are no estimates of population sizes or trends but in Lakekamu Basin it is sparsely distributed in lowland alluvial forest (A. Mack *in litt.*1999). It clearly occurs at low population densities and is likely to be declining slowly through habitat loss and degradation. **Criterion nearly met:** C1.

WHITE-RUMPED FALCON *Polihierax insignis* occurs in **Myanmar** (previously widespread and locally abundant: Oates 1882; it now appears scarce or uncommon, although the large quantity of suitable habitat remaining suggests that healthy populations might survive: C. R. Robson verbally 1998, Robson 2000), **Thailand** (distributed through north, north-east and western provinces south to Ratchaburi, once widespread and fairly common but now scarce throughout after an apparent decline due to clearance of open deciduous forest habitat: Deignan 1963, Lekagul and Round 1991, P. D. Round *in litt*. 1998, Robson 2000), **Laos** (historically "very common" and "locally widespread" in the south, but now apparently local and scarce: Engelbach 1932, David-Beaulieu 1949–1950, Round 1998, Duckworth *et al*. 1999), **Cambodia** (fairly widespread, chiefly in north, with large areas of suitable habitat

remaining: F. Goes *in litt*. 1998, Timmins and Soriyun 1998) and **Vietnam** (previously common locally in the south, now scarce: Robson 2000; only present in any numbers in Dak Lak province: Le Xuan Canh *et al*. 1997, Brickle *et al*. 1998). Populations in Myanmar, Laos and Cambodia are potentially large, but few data are available owing to lack of fieldwork in suitable habitat (JAT). It is resident in savannas and forest, chiefly primary deciduous dipterocarp and mixed deciduous forest of the plains and foothills up to 915 m, where it uses holes in trees for nesting and roosting (del Hoyo *et al*. 1994, Robson 2000, Eames and Tordoff in prep.). Although dry dipterocarp forest has generally suffered less degradation than evergreen forest in many areas (C. R. Robson *in litt*. 1998, C. M. Poole *in litt*. 1999), it is increasingly being cleared and disturbed, through a process of wood collection and burning (Thewlis *et al*. 1998). Given the high levels of hunting in much of its range, and the ease with which this species is shot (David-Beaulieu 1949–1950), persecution presumably poses an additional threat. **Criteria nearly met:** C1; C2a. .

WHITE-FRONTED FALCONET *Microhierax latifrons* is restricted to north Borneo in Sabah, **Malaysia**, where it occurs in forest-edge habitats and cultivation with scattered trees up to 1,200 m (MacKinnon and Phillipps 1993, del Hoyo *et al.* 1994). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysian Borneo (see Threats under Crestless Fireback *Lophura erythrophthalma*). However, the species has probably not declined very rapidly because it is able to survive in human-modified habitats. **Criteria nearly met:** A1c; A2c; C1; C2b.

SULA MEGAPODE *Megapodius bernsteini* is restricted to the Banggai and Sula Islands Endemic Bird Area, **Indonesia**, where it inhabits lowland forest, particularly in coastal areas, and dense lowland scrub fringing farmland (Stattersfield *et al*. 1998; also Coates and Bishop 1997). There are some 7,000 birds in the Banggai Islands, mostly on Peleng, and as many as 38,000 (22,500–54,000) on Taliabu, but these populations are thought to be experiencing declines and local extinctions owing to habitat loss (through logging and clearance), exploitation (collection of eggs, hunting of adults) and introduced animals (cats and dogs as predators, feral domestic chickens as competitors) (Indrawan *et al*. 1992, 1993b, Davidson *et al*. 1995). **Criteria nearly met:** A1c,d,e; A2c,d,e.

SIBERIAN GROUSE *Dendragapus falcipennis* occurs in far eastern **Russia**, from c.120°E to the shores of the Sea of Okhotsk and Sakhalin Island, north to c.60°N and south to the lower Amur region and the Sikhote-Alin mountains, and to the Xiao Hinggan Ling mountains in Heilongjiang, mainland **China** (del Hoyo *et al.* 1994, Storch 2000). It inhabits coniferous forest, mainly of spruce, e.g. *Picea jezoensis* and fir *Abies nephrolepis* (del Hoyo *et al.* 1994), and although it uses secondary forest, it avoids open areas and the youngest stages of forest succession (Storch 2000). It has been found to occur at low densities of between six and 25 birds per 100 km^2 within this relatively small range, although these densities may be underestimated because of its elusive behaviour (Potapov and Flint 1989 in Storch 2000; see Threats under Blakiston's Fish-owl *Ketupa blakistoni*). It is assumed to be declining because of large-scale clear-cutting for timber, forest fires and hunting for food (Storch 2000), and has been reported to disappear rapidly from colonised areas, apparently because it does not fear man (del Hoyo *et al.* 1994). **Criteria nearly met:** A1c,d; A2c,d; C1; C2a.

CHINESE GROUSE *Bonasa sewerzowi* is found in the mountains of south-west mainland **China**, in eastern Tibet, Qinghai, Gansu and western Sichuan, where it occurs in birch and coniferous forest, generally above 1,000 m and at up to 4,000 m in Tibet (Cheng Tso-hsin 1987, del Hoyo *et al.* 1994). Its known range was recently extended to the western edge of the

forest zone in Tibet (at c.93°30′E), suggesting that it may occur much more widely in the vast and continuous forests of south-east Tibet than was previously documented (Lu Xin 1997). It is common in suitable forest, and recent studies in Gansu found extraordinarily high densities of up to 15 occupied territories per km^2 (Klaus *et al.* 1998). Its habitat has been greatly reduced and fragmented by large-scale forest clearance, and illegal hunting and egg-collecting may be a problem in parts of its range (Storch 2000; see Smil 1984, 1993, MacKinnon *et al.* 1996). It is believed to have disappeared from eastern Qinghai and central Gansu because of deforestation, and it has also suffered intensive habitat loss in south-west Gansu (del Hoyo *et al.* 1994). **Criteria nearly met:** A1cd; A2cd.

LONG-BILLED PARTRIDGE *Rhizothera longirostris* is confined to the Sundaic lowlands of southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sarawak and Peninsular **Malaysia**, **Brunei**, and Kalimantan and Sumatra, **Indonesia**, where it is generally uncommon in lowland and hill forest and bamboo up to 1,500 m, with an apparent preference for limestone hills on Borneo (Smythies 1981, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000, R. S. R. Williams *in litt*. 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*). However, this species has probably not declined very rapidly because it ranges up to elevations where forest loss is less severe. **Criteria nearly met:** A1c,d; A2c,d.

SNOW MOUNTAIN QUAIL *Anurophasis monorthonyx* is endemic to the highest peaks of the Snow and Star mountains of Papua, **Indonesia**. (Beehler *et al*. 1986, K. D. Bishop and J. M. Diamond *in litt*. 2000). It has long been known only from high grassland plateaus and the edge of heavy alpine scrub with trees and scrub, between c.3,000 and 4,200 m on the northern slopes of Mounts Jaya (Carstensz) and Trikora (Wilhelmina), where it appears to be uncommon (Beehler *et al*. 1986, Gibbs 1993, M. van Beirs *in litt*. 2000). There are recent records from Lake Habbema, below Mount Trikora (Gibbs 1993, Eastwood 1996, M. van Beirs *in litt*. 2000), and from close to the Freeport mine (SvB). However, its recent discovery in the Star mountains (K. D. Bishop and J. M. Diamond *in litt*. 2000) suggests that it may be much more widespread in the little-known high mountains of Papua. It has a small and naturally fragmented range, and its population appears to be naturally low; its relatively inaccessible habitats are generally not threatened but some habitat has been lost around the Freeport mine (SvB). However, it is actively hunted in accessible areas and the Trans-Irian Highway from Jayapura to Wamena will open up a large area of previously inaccessible habitat to hunters (N. Bostock *in litt*. 1993, D. Gibbs *in litt*. 2000). **Criteria nearly met:** B1+2c,e; C2a; D2.

WHITE-CHEEKED PARTRIDGE *Arborophila atrogularis* is resident in north-east **India** (locally common in Arunachal Pradesh, Assam, Manipur, Meghalaya, Mizoram and Tripura: Grimmett *et al*. 1998), **Bangladesh** (very local in the north-east, but could still occur in the Chittagong Hill Tracts: Grimmett *et al*. 1998) and **Myanmar** (widespread, uncommon to common resident: Smythies 1986, Robson 2000) and mainland **China** (local in Yingjiang area of west Yunnan to west of the Salween river: MacKinnon and Phillipps 2000). It inhabits dense undergrowth of broadleaved primary and secondary evergreen forest (Grimmett *et al*. 1998), sometimes adjacent scrub, bamboo, grassland and cultivation (del Hoyo *et al*. 1994), most frequently below 750 m in India (Grimmett *et al*. 1998), but usually between 610 and 1,220 m in South-East Asia (Robson 2000). It is principally threatened by habitat loss and persecution: within its range, hill forests are diminishing rapidly in extent and becoming fragmented because of intense shifting agriculture and logging (Katti *et al*. 1990, 1992, del Hoyo *et al*. 1994, *Oriental Bird Club Bull*. 21 [1995]: 15–20, P. M. Thompson *in litt*. 1997,

Stattersfield *et al*. 1998, Singh 1999) and hunting and snaring of galliforms is rife (U Tun Yin 1954, Choudhury 1991, B. F. King verbally 1998). However, given the size of its range and the paucity of fieldwork conducted within it, the species is currently likely to be more abundant than records suggest and to exceed the threshold for classification as Vulnerable. **Criteria nearly met:** C1; C2a.

TAIWAN PARTRIDGE *Arborophila crudigularis* is confined to the mountains of central **Taiwan** (China), where it occurs in broadleaf forest at 700–2,300 m (Stattersfield *et al.* 1998). It was formerly widespread but there have been no surveys and few records in recent years (L. L. Severinghaus in McGowan *et al.* 1995), and its population could now be below 10,000 individuals (del Hoyo *et al.* 1994, McGowan *et al.* 1995). Some 11% of Taiwan is protected, in six national parks and in nature reserves and wildlife sanctuaries (P. K. D. Perng *in litt.* 1996), and the species is well represented in these protected areas (L. L. Severinghaus in McGowan *et al.* 1995), where its population is believed to be stable (del Hoyo *et al.* 1994). Elsewhere, it could be declining because of forest loss for timber production and conversion to agricultural land and possibly the use of pesticides (del Hoyo *et al.* 1994, McGowan *et al.* 1995). **Criteria nearly met:** B1+2c; C1; C2a.

CHESTNUT-NECKLACED PARTRIDGE *Arborophila charltoni* has a disjunct, mainly Sundaic distribution but with an outlying population (race *tonkinensis*, formerly treated under Green-legged Hill-partridge *A. chloropus*) in east Tonkin and north Annam, **Vietnam**, ranging (race *charltoni*) through lowland rainforest from southern **Myanmar** and southern **Thailand** into Peninsular **Malaysia**, present (race *atjenensis*) in two isolated areas (Aceh and South Sumatra) of Sumatra, **Indonesia**, and occupying a wide area of Sabah (Borneo), Malaysia, with a specimen (in BMNH) from Mt Dulit, Sarawak (del Hoyo *et al*. 1994, Robson 2000). Estimates of numbers are of possibly under 1,000 (100–10,000) (*charltoni*), 10–100 (*atjenensis*), >100 (*graydoni*), with *tonkinensis* "locally quite common" and in three protected areas (del Hoyo *et al*. 1994, McGowan *et al*. 1995), but *graydoni* is now known to be very common in Danum Valley Conservation Area (Collar *et al*. 1994, D. Yong verbally 1999). As a bird of lowland rainforest it will have suffered extensive losses but its numerical strength in at least two main areas suggests that its decline rate is not as steep as in other species. **Criteria nearly met:** A1c,d; A2c,d.

FERRUGINOUS PARTRIDGE *Caloperdix oculea* ranges from south-east **Myanmar** and south-west **Thailand** through Peninsular **Malaysia** onto Sumatra, **Indonesia**, with isolated areas in Borneo in northern Sarawak and eastern Sabah, Malaysia (Smythies 1981, 1986, Lekagul and Round 1991, del Hoyo *et al*. 1994, Wells 1999). It is generally scarce throughout its range and is likely to have been adversely affected by forest clearance throughout the Sundaic lowlands (see Threats under Crestless Fireback *Lophura erythrophthalma*), but it tolerates secondary forest and drier formations up to 1,000 m (del Hoyo *et al*. 1994). **Criteria nearly met:** A1c,d; A2c,d.

CRESTED PARTRIDGE *Rollulus rouloul* is confined to the Sundaic lowlands, where it is known from southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, **Brunei** and Kalimantan and Sumatra, **Indonesia**, where it occurs in broadleaved evergreen and dense primary lowland and hill forests and bamboo up to 1,550 m (Smythies 1981, 1986, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*). The species remains common in several other areas, is able to persist in

selectively logged forest and can utilise early-stage regenerating forest (Wells 1999), and it is therefore unlikely to be under any immediate threat. **Criteria nearly met:** A1c,d; A2c,d.

SATYR TRAGOPAN *Tragopan satyra* occurs in the Himalayas of **Nepal** (uncommon: Grimmett *et al*. 1998), **India** (uncommon: Grimmett *et al*. 1998), **Bhutan** (fairly common and apparently stable: Inskipp *et al*. 1999a) and mainland **China** (local, with a limited range in south and south-east Tibet: Cheng Tso-hsin 1987, MacKinnon and Phillipps 2000). The total population is not thought to exceed 20,000 individuals (McGowan and Garson 1995). It is resident in moist oak and rhododendron forest with dense undergrowth and bamboo clumps, mixed forest, shrubberies and densely vegetated ravines, usually between 2,200 m and 4,250 m in the breeding season, sometimes moving down to 1,800 m in winter (Grimmett *et al*. 1998, MacKinnon and Phillipps 2000). Major threats include excessive hunting as well as habitat clearance and degradation due to timber harvesting, livestock-grazing, fuelwood and fodder collection (McGowan and Garson 1995), such that its distribution is now fragmented in the Indian subcontinent (Grimmett *et al*. 1998). It occurs in several protected areas throughout its range (McGowan and Garson 1995). **Criterion nearly met:** C1.

SWINHOE'S PHEASANT *Lophura swinhoii* is found in the mountains of central **Taiwan** (China), where it inhabits primary broadleaf forest and mature secondary forest at 200–2,300 m (Severinghaus and Blackshaw 1976, del Hoyo *et al.* 1994, Stattersfield *et al.* 1998). Intensive fieldwork in the early 1970s suggested that there might be 5,000–10,000 individuals, although a recent estimate of c.6,500 in Yushan National Park alone (del Hoyo *et al.* 1994) indicates that its total population is likely to exceed 10,000 birds. Heavy hunting pressure was a problem for it in the past, but is not a serious threat nowadays (L. L. Severinghaus *in litt.* 1996). The species became extinct at several localities in the 1960s and 1970s, but it apparently remains common in suitable habitat, and it has populations in several protected areas; its numbers are probably stable where protected, but may be declining elsewhere because of a variety of pressures on its habitat (del Hoyo *et al.* 1994). **Criteria nearly met:** B1+2c; C1; C2a.

CRESTED FIREBACK *Lophura ignita* is known from extreme southern Tenasserim, **Myanmar**, peninsular **Thailand**, Peninsular and East **Malaysia**, Kalimantan, Sumatra and Bangka, **Indonesia**, and **Brunei** (Smythies and Davison 1999, Wells 1999). There are no recent records from Myanmar, it is now rare in peninsular Thailand, and despite there being recent records from Peninsular Malaysian protected areas, its current distribution there is rather poorly known (Smythies 1986, Round 1988a, Wells 1999). It is apparently still locally common in Sabah, Kalimantan and on Sumatra (Smythies and Davison 1999, R. Sözer verbally 1999), and the total population may be more than 100,000 (McGowan and Garson 1995). It occurs in lowland forest, but ranges up to 1,000 m, and survives in logged, disturbed and secondary forest, but the rates of forest loss in the Sundaic lowlands are still considerable (see Threats under Crestless Fireback *L. erythrophthalma*), and it must be declining in lowland Indonesia, and probably also outside well protected areas in the Thai-Malay peninsula. ***Taxonomy*** The nominate race *ignita*, present on Borneo and, curiously, Bangka, is very different from the Peninsular and Sumatran form *rufa*, and the two are probably better considered as separate species, although the highly variable southern Sumatran population called "*macartneyi*" probably represents a hybrid swarm between the two (van Marle and Voous 1988, del Hoyo *et al.* 1994); but even if *rufa* and *ignita* were so separated, it is not certain that either would qualify for threatened status. **Criteria nearly met:** A1c; A2c.

SIAMESE FIREBACK *Lophura diardi* is found in **Thailand** (uncommon to locally common resident, principally in the north-east and south-east, c.5,000 individuals estimated: Lekagul and Round 1991, Collar *et al*. 1994, Robson 2000), **Laos** (widespread and locally abundant,

but heavily snared and therefore probably threatened nationally: Davidson *et al*. 1997, Duckworth *et al*. 1999), **Cambodia** (locally common and widespread: Robson 2000) and **Vietnam** (locally common and widespread in central and southern regions: J. C. Eames and A. W. Tordoff *in litt*. 2000). The global population has been estimated to fall below 10,000 individuals (McGowan and Garson 1995), but this is likely to be an underestimate given its tolerance of considerably degraded forest habitat (Duckworth *et al.* 1999). It occurs in evergreen, semi-evergreen and bamboo forest, second growth and scrub, often near roads and tracks through the forest, chiefly in the plains and foothills to 500 m, but occasionally up to 800 m (Eames *et al.* 1992, Robson *et al.* 1993b, Collar *et al*. 1994, Duckworth *et al*. 1999), and perhaps 1,150 m (Showler *et al*. 1998a). It remains under pressure, however, from continuing extensive lowland forest destruction within its range and, perhaps more severely, from hunting and snaring (Lekagul and Round 1991, Ling *et al*. 1999). **Criteria nearly met:** A1b,c; A2b,c; C1.

TIBETAN EARED-PHEASANT *Crossoptilon harmani* has been recorded in south-east Tibet, mainland **China**, and at least one locality in extreme northern Arunachal Pradesh, **India**, where it occurs in tall dense scrub in dry river valleys, the borders of mixed broadleaf and coniferous forest, coniferous forest, and grassy hill slopes, from 3,000 to 5,000 m (rarely down to 2,400 m); it is locally common, and adaptable to disturbed habitats (Ali and Ripley 1968–1998, C. R. Robson *in litt.* 1991, Grimmett *et al.* 1998). Recent surveys have indicated that its population must be greater than 10,000 individuals (Lu Xin verbally 1998). Deforestation and hunting may, however, be significant threats in Tibet, and it is probably declining (Zhang Zhengwang in McGowan and Garson 1995). **Criteria nearly met:** C1; C2a.

WHITE EARED-PHEASANT *Crossoptilon crossoptilon* is found in mainland **China**, where it is known from Qinghai, Sichuan, Yunnan and Tibet, and occurs in coniferous and mixed forests near the treeline, plus subalpine birch and rhododendron scrub, at 3,000–4,300 m (Meyer de Schauensee 1984, Cheng Tso-hsin 1987). It was estimated to number only 10,000–50,000 and to be declining (McGowan and Garson 1995), but given the extent of its range it is unlikely that its population is less than 10,000 individuals (Zheng Guangmei verbally 1996). It is threatened by hunting for food and deforestation, but the high-altitude forests that it inhabits are not being lost at a very rapid rate (Zheng Guangmei verbally 1996), and there are recent records from several protected areas (Zhang Zhengwang *in litt.* 1993). **Criteria nearly met:** C1; C2a.

MIKADO PHEASANT *Syrmaticus mikado* is found in the mountains of central **Taiwan** (China), where it inhabits forest with dense undergrowth and bamboo on steep mountain slopes between 1,800 and 3,300 m and possibly higher (Severinghaus and Blackshaw 1976, del Hoyo *et al.* 1994, Stattersfield *et al.* 1998). Yushan National Park was recently estimated to hold c.10,000 individuals, and the species is also known from several other protected areas (del Hoyo *et al.* 1994). Heavy hunting pressure was a problem for it in the past, but is not a serious threat nowadays (L. L. Severinghaus *in litt.* 1996). There may be some decline in its numbers outside protected areas because of habitat loss, but its high-altitude habitats are relatively secure and it is tolerant of secondary growth (del Hoyo *et al.* 1994). **Criteria nearly met:** B1+2c; C1; C2a.

COPPER PHEASANT *Syrmaticus soemmerringii* is found on the islands of Honshu, Shikoku and Kyushu in **Japan**, in coniferous, broadleaf and mixed forest from sea-level to 1,800 m (Brazil 1991, del Hoyo *et al.* 1994). It was once very common, but it appears to have declined substantially because of large-scale hunting and is now considered to be uncommon and hard to find (Brazil 1991, McGowan and Garson 1995). During the 1970s, the number of birds shot by hunters declined rapidly from c.800,000 to c.300,000 per year, but appeared to stabilise at c.100,000 per year during the 1990s (N. Kawaji *in litt.* 1997). The shooting of

females has been illegal since 1976 (Brazil 1991). In addition to the effects of hunting and habitat loss, feral cats and dogs may be causing a reduction in breeding success, and hybridisation (of the five subspecies) between wild and captive-reared stock may also be a problem (H. Higuchi in McGowan and Garson 1995). There is a need for surveys to establish the status of all of the subspecies, and for regulation of sport hunting to ensure that it is made sustainable (McGowan and Garson 1995). **Criteria nearly met:** A1cd; A2cd.

GREAT ARGUS *Argusianus argus* is confined to the Sundaic lowlands, where it is recorded from southern Tenasserim, **Myanmar**, peninsular and south-west **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct: Lim Kim Seng *in litt*. 2001), **Brunei** (extirpated from many areas), Kalimantan and Sumatra, **Indonesia** (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). It is generally uncommon in tall, dry, lowland primary and logged forests, up to 1,300 m, but principally occurs below 900 m (Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in much of Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*). However, this species has probably not declined very rapidly because it ranges up to elevations where forest loss is less severe and occurs in selectively logged sites, even though it is much sparser in deciduous forest and rare to absent from lowland peatswamp and white-sand heath forests (Wells 1999). Excessive trapping occurs in many areas of Borneo (MacKinnon and Phillipps 1993). **Criteria nearly met:** A1c,d; A2c,d.

WHITE-STRIPED FOREST-RAIL *Rallina leucospila* is endemic to north-west Papua (formerly Irian Jaya), **Indonesia**, where it is known from the Tamrau, Arfak and Wandammen mountain ranges (Beehler *et al*. 1986, Taylor 1998). There are very few recent records but one was seen and others heard in 1994–1995 and it was reported by local guides to be rare and to have a very patchy distribution (D. Gibbs *in litt*. 1996). It is a terrestrial inhabitant of interior montane forest from 1,450 to 1,600 m (Beehler *et al*. 1986). The extensive rainforests within its range remain largely undisturbed, owing to their geographical isolation and the low-density and traditional lifestyle of the human population (Sujatnika *et al.* 1995). There are two protected areas within its range, the Wondiwoi/Wandammen Nature Reserve (730 km^2) and the Pegunungan Arfak Nature Reserve, (currently 683 km^2), and the proposed Pegunungan Tamrau Nature Reserve (c.4,415 km^2) in the Tamrau mountains (Sujatnika *et al.* 1995). The only likely threats to this species are localised forest loss and incidental kills from dogs on hunting trips (N. Bostock *in litt*. 1993), but it may be threatened by introduced predatory mammals even in these remote mountains (D. Gibbs *in litt*. 2000). **Criteria nearly met:** B1+2c,e; C2a.

BAND-BELLIED CRAKE *Porzana paykullii* is known to breed in the middle and lower Amur Valley and Primorye in south-east **Russia** (Dement'ev and Gladkov 1951–1954) and Heilongjiang, Jilin, Liaoning, Hebei and northern Henan in north-east mainland **China** (Cheng Tso-hsin 1987), and it has been reported to breed in both **North Korea** and **South Korea** although this is unproven (Austin 1948, Gore and Won 1971). It occurs on passage and/or in winter in North Korea, South Korea, mainland China (Inner Mongolia, Shandong and southwards to southern China, including Hong Kong), central **Thailand**, **Vietnam**, **Malaysia** (Peninsular Malaysia, Sabah and Sarawak) and **Indonesia** (northern Sumatra, Java and northern Kalimantan, with one record from southern Sulawesi); it is a vagrant on Sakhalin Island in Russia and in Japan, and there is an erroneous record from the Philippines (del Hoyo *et al.* 1996, Taylor 1998). It breeds in lowland marshes and meadows with tussocks, thickets or small trees, and is often found near villages and along field edges; it

winters in wet grassland, swamps and paddyfields (del Hoyo *et al.* 1996). In Russia, it has a restricted distribution, and although it is still "abundant" in parts of its breeding range in most areas it appears only sporadically; it is threatened there by the intensification of agriculture and industrial development and other forms of habitat loss (Potapov and Flint 1987). Very little information is available from elsewhere, and its status and distribution urgently require investigation, especially in the wintering range (del Hoyo *et al.* 1996). **Criterion nearly met:** C1.

NEW GUINEA FLIGHTLESS RAIL *Megacrex inepta* is endemic to the lowlands of central New Guinea (Papua, formerly Irian Jaya, **Indonesia** and **Papua New Guinea**), where it inhabits lowland forest, especially wet thickets, swamp forest and mangroves, and although there are very few records, it is reported to be locally common, especially where sago is harvested (Gregory 1996, Taylor 1998). Although flightless, it appears to be able to fight off attacks from dogs and to climb into trees to escape predators (Ripley 1964, Coates 1985). It is hunted in some areas, sometimes targeted specifically with sago lures, and it may be affected by feral pigs, and lowland forests across its range are under pressure from logging (K. D. Bishop *in litt.* 1994, P. Gregory *in litt.* 1994, Gregory 1996, Taylor 1998, T. Leary *in litt.* 2000). Although its population size is unknown, it appears to be small, but there is no evidence of a significant decline. **Criterion nearly met:** C1.

LITTLE BUSTARD *Tetrax tetrax* has two widely separated breeding populations: in its eastern range it occurs in **Russia** (9,000 displaying males), **Kazakhstan** (c.20,000 individuals), **Ukraine** (100–110 individuals) (Y. Andryuschenko *in litt*. 1999), north-west mainland **China**, northern **Iran** and **Turkey** (20–100 pairs) (Eken and Magnin in press), while its western range covers **Spain** (100,000–200,000 displaying males) and **Portugal** (20,000 individuals), with smaller populations in **Italy** (1,500–2,200 individuals), **France** (1,087–1,256 displaying males) (Jolivet 1997) and possibly **Morocco** (de Juana and Martínez in press). Most birds disperse in winter but stay within the countries above, but at least **Azerbaijan** plays host to wintering populations, including one of 10,000–30,000 birds (Heath and Evans 2000). It is a rare and erratic visitor to **Pakistan** and **India** (Grimmett *et al.* 1998) and a winter visitor to **Afghanistan** (Paludan 1959). The global population (excluding Kazakhstan) is estimated at a minimum of 240,000 individuals (C. Martínez *in litt*. 1999). Whilst it remains widespread and numerous, in some parts of its range it has declined dramatically since the nineteenth century (Tucker and Heath 1994). It inhabits dry grassland and, in Europe, it also occurs in areas of low-intensity arable cultivation and pastoral land, selecting areas with a high diversity of ground cover such as mosaics of pasture, long-rotation fallow land and legume crops (Tucker and Heath 1994). The primary cause of its decline has been conversion of dry grassland and low-intensity cultivation to intensive arable agriculture, especially where this has included the planting of monocultures, irrigation or afforestation (Tucker and Heath 1994). This continues to be the primary threat and cause of continuing declines, but it also suffers from illegal hunting (Y. Andryuschenko *in litt*. 1999). **Criteria nearly met:** A1c,d; A2c,d.

HOUBARA BUSTARD *Chlamydotis undulata* occurs over a huge range from the Canary Islands, **Spain**, to mainland **China** (del Hoyo *et al.* 1996). The population has been estimated at 49,000–62,000 individuals, but it is likely to exceed 100,000 birds (F. Launay *in litt.* 2000). *C. u. undulata* (9,800 birds) is resident in North Africa where it has declined in **Libya**, **Egypt** and **Tunisia**, and probably also in **Algeria**, **Mauritania**, **Morocco** and **Sudan**; *C. u. fuertaventurae* (700–750 birds) occurs on the Canary Islands, **Spain** (Goriup 1997). *C. u. macqueenii* is thought to occupy six sub-regions: resident and migratory birds occur in the Middle East (**Turkey**, **Jordan**, **Israel**, **Iraq**, **Kuwait**, **Bahrain**, **Oman**, **Qatar**, **Saudi Arabia**, **United Arab Emirates**, **Syria**, **Yemen**), and in **Russia** (including in the Asian region: see

Dement'ev and Gladkov 1951–1954), **Iran**, **Pakistan**, **India**, **Afghanistan**, **Uzbekistan**, **Tajikistan**, from western **Kazakhstan** to **Turkmenistan**, and on the Mongolian plateau and in the Gobi desert of **Mongolia** and western China (Goriup 1997). The population of this subspecies is estimated at 39,000–52,000 individuals, mostly breeding in Kazakhstan (30,000–40,000) (F. Launay *in litt.* 2000), although numbers in China are likely to be far higher than the current estimate of 500 birds (F. Launay *in litt.* 2000). Declines are reported from Bahrain, Jordan, Iran, Iraq and India (Goriup 1997). Populations from some subregions are believed to mix on the wintering grounds (F. Launay *in litt.* 2000). The main threats are habitat loss and degradation as desert areas are developed for agriculture and infrastructure projects; these are compounded by high hunting pressure from falconers, with new areas in Central Asia, close to breeding grounds, increasingly being exploited (Uzbekistan Zoological Society *per* P. D. Goriup *in litt.*1999, F. Launay *in litt.* 2000). There are no reliable data for rates of decline, but given the substantial threats declines are likely to be significant and possibly widespread; moreover, they may accelerate if hunting pressure in Central Asia increases. ***Taxonomy*** There has been a suggestion that the two main races represent separate species (Gaucher *et al.* 1996), but the differences are not great and the distributions unclear; this matter requires urgent attention. **Criteria nearly met:** A1c,d; A2c,d.

MALAYSIAN PLOVER *Charadrius peronii* is a breeding resident in **Vietnam** (scarce in the south), **Cambodia** (rare), Peninsular **Thailand** (local and uncommon), Peninsular **Malaysia** (scarce to locally common), East Malaysia, **Singapore** (rare), **Brunei** (apparently declining), **Philippines** (widespread but uncommon and suffering from substantial loss of range with human disturbance of its nesting beaches) and **Indonesia** (local around coasts and offshore islands of Sumatra, uncommon on and around Borneo and Bali, very rare on mainland Java; uncommon and sparsely distributed in the Lesser Sundas and the Sulawesi subregion) (Mann 1987, 1988, Lekagul and Round 1991, Dickinson *et al.* 1991, MacKinnon and Phillipps 1993, del Hoyo *et al.* 1996, Coates and Bishop 1997, Smythies and Davison 1999, Robson 2000). The global population has been estimated at fewer than 10,000 individuals (Rose and Scott 1997), but this may be an underestimate given the number of islands that it inhabits. However, there is growing concern that numbers are limited by its highly linear distribution and under intense pressure from the development and disturbance of coastal habitat (del Hoyo *et al.* 1996, Collar *et al.* 1999). It frequents quiet sandy bays, coral sand beaches, open dunes and artificial sand-fills, where it lives in pairs, generally not mixing with other waders (del Hoyo *et al.* 1996). **Criterion nearly met:** C1.

JAVAN PLOVER *Charadrius javanicus* is restricted to the Javan Coastal Zone Endemic Bird Area and the Kangean Secondary Area, occurring on sandy beaches and mudflats on Java and the Kangean Islands, **Indonesia** (Stattersfield *et al.* 1998). The taxonomic status of this species is extremely unclear and records attributable to it are therefore very sparse, meaning that its conservation status is also unclear; nevertheless, it was recently found common on southern Madura (del Hoyo *et al.* 1996). While it may prove widespread, it is likely to be declining with human disturbance of beaches within its limited range. **Criterion nearly met:** C1.

SULAWESI WOODCOCK *Scolopax celebensis* is restricted to the Sulawesi Endemic Bird Area (excluding the South-East Peninsula but with a recent record from the East Peninsula: White and Bruce 1986), **Indonesia**, where it inhabits montane forest and bamboo thickets at 1,700–2,300 m (Stattersfield *et al.* 1998; also Riley 1924), or 1,100–2,300 m, with records from very few localities (Coates and Bishop 1997). Although it is possibly fairly common and simply overlooked owing to its cryptic plumage, silent ground-haunting behaviour and nocturnal habits (see Riley 1924, Stresemann and Heinrich 1939–1941), there have been few

if any recent records and there must be concern that it is local, possibly affected by habitat disturbance and loss, and even by the spread of feral cats into montane Sulawesi (see entry for Geomalia *Geomalia heinrichi* in this section). **Criterion nearly met:** C1.

GREAT SNIPE *Gallinago media* breeds primarily in **Russia**, east to 95°E (150,000–250,000 pairs), with large numbers in **Belarus** (12,000–20,000 pairs) and **Norway** (5,000–15,000 pairs). It also breeds in **Poland**, **Finland**, **Sweden**, **Estonia** (500–700 males), **Latvia**, **Lithuania**, and **Ukraine** (Tucker and Heath 1994). From early August, it migrates through **Turkey**, **Cyprus** and **Egypt**, with birds gathering in grasslands in **Ethiopia** (Tucker and Heath 1994, J. S. Ash *in litt.* 1999). When these dry out in October, birds follow the rains south and west to **Sudan**, **Chad**, **Burkina Faso**, **Mali**, **Mauritania**, **Senegal**, **Sierra Leone**, **Liberia**, **Ivory Coast**, **Ghana**, **Togo**, **Benin**, **Nigeria**, **Cameroon**, **Gabon**, **Congo**, **Democratic Republic of Congo**, **Kenya**, **Uganda**, **Rwanda**, **Burundi**, **Tanzania**, **Malawi**, **Zambia**, **Zimbabwe**, **Mozambique**, **South Africa**, **Angola** and **Namibia** (Tucker and Heath 1994, del Hoyo *et al.* 1996). It occurs rarely in **India**, **Sri Lanka** (Grimmett et al. 1998), and **Myanmar** (Robson 2000). Its range has contracted and numbers have declined since the late nineteenth century: although the Scandinavian population has stabilised, there are continuing rapid declines in the southern forest and forest-steppe zones of Russia and Ukraine, largely as a result of the loss (to farmland and reservoirs) of nesting habitats—which include floodplain and tussock meadows, natural fens with scattered bushes, and peatlands up to 1,200 m—compounded by hunting in eastern Europe and in its wintering range. (Tucker and Heath 1994, del Hoyo *et al.* 1996, Kålås *et al.* 1997). CMS Appendix II. **Criteria nearly met:** A1c,d; A2c,d.

FAR EASTERN CURLEW *Numenius madagascariensis* breeds in eastern **Russia**, from the upper reaches of the Nizhnyaya Tunguska river east though the Verkhoyarsk mountains to Kamchatka, and south to Primorye and north-eastern **Mongolia** (del Hoyo *et al.* 1996). It has been recorded as a non-breeding visitor to **Japan**, **North Korea**, **South Korea**, mainland **China**, **Hong Kong**, **Taiwan** (China), **Bangladesh**, **Thailand**, **Vietnam**, **Philippines**, **Malaysia**, **Singapore**, **Brunei**, **Indonesia**, **Papua New Guinea**, **Australia** and **New Zealand** (Collar *et al.* 1994, also Lim and Gardner 1997, Grimmett *et al.* 1998, Eames and Tordoff in prep.). It breeds on open mossy or transitional bogs, moss-lichen bogs and wet meadows, and on the swampy shores of small lakes; in the non-breeding season it is essentially coastal, occurring at estuaries, mangrove swamps, saltmarshes and intertidal flats, particularly those with extensive seagrass (Zosteraceae) meadows (del Hoyo *et al.* 1996). Its total population was recently estimated at 21,000 at most, of which c.19,000 winter in Australia with smaller numbers in the Philippines, Sumatra and New Zealand, and it is believed to be declining (del Hoyo *et al.* 1996, Rose and Scott 1997). The main threats are habitat loss, hunting and a decrease in the availability of food because of pollution; the reclamation of coastal wetlands in the Chinese section of its flyway (for which see Threats under Saunders's Gull *Larus saundersi*) may represent a major constraint. A potential threat may be that females probably tend to migrate further south, to the more threatened south Australian wetlands (del Hoyo *et al.* 1996). **Criteria nearly met:** A1c,d,e; A2c,d,e; C1.

ASIAN DOWITCHER *Limnodromus semipalmatus* has a disjunct breeding range in the steppe regions that extend from West to East Siberia in **Russia**, and south into **Mongolia** and Heilongjiang in north-east mainland **China** (del Hoyo *et al.* 1996). It has been recorded as a non-breeding visitor to **Japan**, **North Korea**, **South Korea**, mainland China, **Hong Kong**, **Taiwan** (China), **Kazakhstan**, **Uzbekistan**, **India**, **Bangladesh**, **Sri Lanka**, **Myanmar**, **Thailand**, **Cambodia**, **Vietnam**, **Philippines**, **Malaysia**, **Singapore**, **Brunei**, **Indonesia**, **Papua New Guinea**, **Australia** and **New Zealand** (Collar *et al.* 1994, also King *et al.* 1975, Mundkur *et al*. 1995a, Lim and Gardner 1997, Grimmett *et al.* 1998, Eames and Tordoff in prep.). It breeds in

extensive freshwater wetlands in the steppe and forest-steppe zones, and during the non-breeding season it occurs in sheltered coastal environments, primarily estuarine and intertidal mudflats, lagoons, creeks and saltworks (del Hoyo *et al.* 1996). Recent population estimates include 15,000–20,000 birds (Rose and Scott 1997) and at least 20,000 (del Hoyo *et al.* 1996). It is dependent on rather a small number of wetlands, notably the wintering sites at the Banyuasin delta on Sumatra, where up to 13,000 were estimated in 1988, and Ujung Pangkah in East Java (del Hoyo *et al.* 1996). It may therefore be particularly vulnerable to habitat loss, hunting, pollution and other pressures on both the breeding and wintering grounds (see Threats under Spoon-billed Sandpiper *Eurynorhynchus pygmeus* and Spotted Greenshank *Tringa guttifer*). **Criteria nearly met:** A1c,d,e; A2c,d,e; C1.

BEACH THICK-KNEE *Esacus magnirostris* is widespread around coasts from the Andaman Islands, **India**, Mergui archipelago, **Myanmar**, islands off peninsular **Thailand**, and Peninsular **Malaysia** and **Singapore** through **Indonesia**, the **Philippines**, **Papua New Guinea**, the **Solomon Islands** and **Australia** (del Hoyo *et al*. 1996, Lim Kim Seng *in litt*. 2001). The range is, however, essentially linear so that, despite ranging from the farthest point west in Australia north round to south of the farthest point east, its total population there may be as few as 1,000 birds (Marchant and Higgins 1993); it is very rare on and around Sumatra (Holmes 1996), and there appears to be extensive (but wholly unquantified) human disturbance of beach habitats in many areas (del Hoyo *et al.* 1996). **Criterion nearly met:** C1.

BUSH STONE-CURLEW *Burhinus grallarius* is a resident of all but the most arid parts of mainland **Australia**, and many offshore islands, with a tiny breeding population in southern New Guinea (Papua, formerly Irian Jaya, **Indonesia**, and **Papua New Guinea**), where it affects open forest and woodland (Marchant and Higgins 1993). In Australia, it is now largely absent south and east of the Great Dividing Range, and is scarce elsewhere in southern Australia; the northern Victorian population is apparently declining, with counts of 328 birds in 1985 and 141 in 1991 (Marchant and Higgins 1993). Most recent records from South Australia are from islands, and it has been declining in south-western Australia since the 1920s, but it remains common in northern Australia and on many continental islands, even within towns, although it has declined near Rockhampton; the total Australian population has been estimated at 15,000 individuals (Marchant and Higgins 1993). Its rarity and continuing decline in its southern range has been attributed to predation by introduced foxes, habitat clearance for agriculture, habitat degradation by pastoralism, and removal of leaf-litter from habitat remnants; hunting is now illegal but may continue, and other threats include predation by cats and domestic dogs, and poisoning by ingesting bait set for rabbits (Garnett and Crowley 2000). **Criterion nearly met:** C2b.

BLACK-BELLIED TERN *Sterna acuticauda* is known from southern mainland **China** (previously regular in Yunnan, now very rare), **Pakistan** (frequent in northern Sind and Punjab), **India** (widespread and locally fairly common; declining in Gujarat), **Nepal** (locally fairly common), **Bangladesh** (previously common; now a local breeder), **Myanmar** (previously abundant, now a scarce to uncommon resident, apparently fewer than 25 pairs surviving), **Thailand** (formerly resident in the north-west, now very rare and probably extinct as a breeding species), **Laos** (previously bred in large numbers along the Mekong channel, now very rarely recorded), **Cambodia** (in the early 1960s apparently fairly common along the Mekong, but very few recent records) and **Vietnam** (formerly occurred regularly in Cochinchina, and occasionally in Annam, but no recent records) (Cripps 1878, Deignan 1945, Thomas 1964, Ali and Ripley 1968–1998, Smythies 1986, Cheng Tso-hsin 1987, Lekagul and Round 1991, Roberts 1991–1992, del Hoyo *et al*. 1996, Thompson and Johnson 1996, Grimmett *et al*. 1998, Thewlis *et al*. 1998, Timmins and Soriyun 1998, Duckworth *et al*. 1999, MacKinnon

and Phillipps 2000, Robson 2000). It is found on large rivers (usually breeding on sandspits and islands) and marshes, occasionally on smaller pools and ditches, in lowlands (but not on the coast) up to 730 m (Inskipp and Inskipp 1991). There has been an extremely rapid decline in South-East Asia and it is now almost extinct in the region (Lekagul and Round 1991, J. W. Duckworth *in litt*. 1998, Duckworth *et al*. 1999). Nevertheless, the suggestion that the world population could be below 10,000 (Rose and Scott 1997) may be over-cautious given its status in South Asian countries. Threats include the destruction of breeding habitat (islands and sandspits in larger rivers are increasingly cultivated), the collection of eggs for food, and natural or dam-determined flooding of nests (Scott 1989, Roberts 1991–1992, Collar *et al*. 1994). **Criteria nearly met:** A1a,c,d; A2a,c,d; C1.

ANDAMAN WOOD-PIGEON *Columba palumboides* is endemic to the Andaman and Nicobar (including Great Nicobar, Nancowry, Car Nicobar and Batti Malv) archipelagos, **India** (Stattersfield *et al*. 1998). It is uncommon in the Andamans (Davidar *et al*. 1996, Grimmett *et al*. 1998). Pairs or small parties wander from island to island in search of fruiting figs and other trees in dense broadleaved evergreen forest (del Hoyo *et al*. 1997, Grimmett *et al*. 1998). Its highly restricted range and predilection for the densest portions of forest imply that it is potentially threatened by habitat loss and fragmentation (del Hoyo *et al*. 1997). Indeed, although forest remains fairly extensive on the Andamans and Nicobars, the human population on larger islands is rising rapidly and habitat is consequently under pressure from agriculture, grazing and logging (Whitaker 1985, Curson, 1989, Pande *et al.* 1991, Sinha 1992, Stattersfield *et al*. 1998). Hunting is also apparently common on the islands and may affect this species, while planned development projects could seriously affect its habitat (Stattersfield *et al*. 1998). **Criterion nearly met:** C1.

JAPANESE WOOD-PIGEON *Columba janthina* is an uncommon and local resident in **Japan**, on small islands off southern Honshu, Shikoku and Kyushu, south through the Nansei Shoto islands to the Yaeyama Islands and south through the Izu Islands to the Ogasawara and Iwo Islands (Brazil 1991). It occurs locally on small islands off the south coast of **South Korea** (Gore and Won 1971), and it has been recorded (presumably as a vagrant) in eastern Russia (Knystautas 1993), Shandong, mainland **China** (Cheng Tso-hsin 1987) and on **Taiwan** (China) (Wang Chia-hsiong *et al.* 1991). It inhabits dense subtropical forest and warm temperate evergreen broadleaf forests, and is heavily dependent on mature forest (Brazil 1991). Although it is still relatively common on the Izu Islands, it has apparently declined there since the 1950s, it was though to have declined on Okinawa during the 1980s because of forestry activities, and the subspecies *C. j. nitens* (which used to occur on the Ogasawara and Iwo Islands) is almost certainly extinct (Brazil 1991). **Criterion nearly met:** C1.

ANDAMAN CUCKOO-DOVE *Macropygia rufipennis* is endemic to the Andaman and Nicobar (Nancowry subgroup and Great Nicobar) archipelagos, **India**, where it is locally frequent on the former and scarce on the latter. It frequents dense broadleaved primary and secondary evergreen forest, tolerating some degree of habitat degradation. While forest remains fairly extensive on the Andamans and Nicobars, the human population on larger islands is rising rapidly and habitat is consequently under pressure from agriculture, grazing and logging. Hunting is also apparently common on the islands, possibly affecting this species, and planned development projects on the Nicobars could seriously affect its habitat. **Criteria nearly met:** B1+2a,b,c,d,e; C1.

NICOBAR PIGEON *Caloenas nicobarica* occurs on the Andaman and Nicobar Islands (**India**), Mergui Archipelago (Myeik Kyunzu) (**Myanmar**), islands off south-west peninsular **Thailand**, islands around Peninsular **Malaysia**, islands off southern **Vietnam**, islands around

Sumatra, **Indonesia**, islands in Wallacea and West Papua, possibly also East Timor, many islands in the **Philippines**, islands in eastern **Papua New Guinea** and throughout the **Solomon Islands**, plus **Palau** (with endemic race *pelewensis*) in the Caroline Islands (**USA**) (Ali and Ripley 1968–1998, del Hoyo *et al.* 1997, Wells 1999). Cambodia records are in error (C. M. Poole *in litt.* 2001). It breeds, often in dense colonies, on normally extremely small wooded offshore islands, and forages *in situ* or (at least at times) on adjacent mainland (or larger island) areas (del Hoyo *et al.* 1997, Wells 1999). Relentless trapping for food, the pet trade and perhaps still their (certainly once-prized) gizzard-stones seriously suppresses populations, as does clearance of small islands for plantations and, almost certainly, the colonisation of such islands by rats, cats and other alien predators (Collar and Andrew 1988, Wells 1999, NJC). The race *pelewensis* (King 1978–1979) may still number up to 1,000 birds (Engbring and Pratt 1985). **Criteria nearly met:** A1c,d; A2c,d.

LUZON BLEEDING-HEART *Gallicolumba luzonica* is near-endemic to Luzon, **Philippines**, where it is recorded from a wide range of localities in lowland (below 1,400 m) forest, including selectively logged and otherwise poor-quality secondary growth and even plantations, in the Sierra Madre south to Quezon National Park and Mt Makiling (Baud 1978, Dickinson *et al.* 1991, Altamirano 1993, Hornbuckle 1994, Gonzalez 1995), with a recent report (including a captive bird) from Mt Bulusan in the south in October 1995 (R. Altamirano and A. C. Diesmos verbally 1995). Habitat loss within its range has been extensive, and it is vulnerable to snaring, as with the more range-restricted Whiskered Pitta *Pitta kochi*, but it remains widespread and locally numerous (Poulsen 1995). Evidently very small populations occur in Polillo (Gonzalez and Dans 1996) and Catanduanes, where the subspecies *rubiventris* is known from a single specimen (Gonzales 1983). **Criteria nearly met:** A1c,d; A2c,d; C2a.

CINNAMON-HEADED GREEN-PIGEON *Treron fulvicollis* is confined to the Sundaic lowlands, where it is known from southern Tenasserim, **Myanmar**, peninsular **Thailand** (no recent records), Sabah, Sarawak and Peninsular **Malaysia**, **Singapore** (non-breeding visitor), **Brunei** (very uncommon) and Kalimantan and Sumatra (including the Riau and Lingga archipelagos, Bangka, Belitung, Siberut and Nias islands), **Indonesia** (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). It occurs in freshwater swamp forest, peatswamp forest, mangroves, coastal forest, open scrub and second growth, including wooded gardens, to 1,200 m (Smythies 1981, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). Non-breeders occasionally visit higher elevation dry land forests (Wells 1999). It is threatened throughout its range by the clearance, fragmentation and degradation of lowland forest (Collins *et al.* 1991), including swamp forest (see Threats under Hook-billed Bulbul *Setornis criniger*). Although the species has almost certainly experienced a significant population decline in the past decade owing to habitat loss, it occurs in relatively less threatened forests on slopes and appears to be able to survive in secondary forest. **Criteria nearly met:** A1c,d; A2c,d.

SUMBA GREEN-PIGEON *Treron teysmannii* is restricted to the Sumba Endemic Bird Area, **Indonesia**, where it occurs at all altitudes in open forest, particularly disturbed areas with tall trees (Stattersfield *et al.* 1998; also del Hoyo *et al.* 1997). The population in the early 1990s was estimated to be somewhat larger than 14,000 birds and presumably declining with forest clearance (M. J. Jones *et al.* 1995). **Criterion nearly met:** C2b.

SUMATRAN GREEN-PIGEON *Treron oxyura* is restricted to the Sumatra and Peninsular Malaysia Endemic Bird Area and the Java and Bali Forests Endemic Bird Area, occurring only in Sumatra (350–1,800 m) and West Java (600–3,000 m), **Indonesia**, in hill and montane forest (Stattersfield *et al.* 1998; also Gibbs in press). Forest destruction has been widespread

in the lower parts of its elevation range (del Hoyo *et al*. 1997; see, e.g., Threats under Crestless Fireback *Lophura erythrophthalma* and Sumatran Ground-cuckoo *Carpococcyx viridis*), and it is now apparently very hard to find on both islands (D. Liley *in litt*. 1998, P. Morris *in litt*. 1998). **Criterion nearly met:** C1.

WHISTLING GREEN-PIGEON *Treron formosae*, considered a small-island specialist by Rand (1970), nests in the Nansei Shoto islands of **Japan**, **Taiwan** (China) and the Batanes islands north of Luzon, **Philippines**; it is uncommon and local on the Batanes, rare in Taiwan and common only in the Nansei Shoto (Dickinson *et al.* 1991, del Hoyo *et al.* 1997). It inhabits subtropical broadleaf evergreen forest, cultivated fields where there are trees nearby, and even towns where there are gardens (Brazil 1991), mainly on small islands but at up to 2,000 m on Taiwan (Stattersfield *et al.* 1998). In the Philippines, threats presumably comprise deforestation and hunting, as with other columbids, while commercial development through tourism was perceived as an imminent threat to Mt Irada (R. J. Timmins *in litt.* 1997). **Criterion nearly met:** C1.

CREAM-BELLIED FRUIT-DOVE *Ptilinopus merrilli* is endemic to Luzon—where it is found chiefly in primary and selectively logged forest up to 1,100 m in the Sierra Madre (race *faustinoi* north of and nominate *merrilli* south of Quirino)—and to two of its satellites, Polillo and Catanduanes, **Philippines**, where, despite being considered rare, nomadic and under pressure from hunting (Dickinson *et al*. 1991, Poulsen 1995, BirdLife database), the total number of localities for the species is at least 24 (from the map in Poulsen 1995) and it appears to be fairly common, even in ultrabasic forest (Altamirano 1993, Hornskov 1995a, Poulsen 1995), and in 1996 it was found to be abundant on Patnanungan island in the Polillo group (Gonzalez and Dans 1996). Nevertheless, habitat destruction, both actual and potential, represents a significant threat to this species (see, e.g., Threats *Habitat loss: Luzon* under Philippine Eagle *Pithecophaga jefferyi*). **Criteria nearly met:** A1c,d; A2c,d; C2a.

JAMBU FRUIT-DOVE *Ptilinopus jambu* is confined to the Sundaic lowlands, from peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, **Singapore** (scarce and irregular non-breeding visitor), Kalimantan, Sumatra and West Java, **Indonesia** and **Brunei** to 1,500 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). It inhabits mangrove and evergreen forests in lowlands and submontane regions (Wells 1999). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*). However, the species's ability to persist in second growth and at higher elevations, where forest destruction has been less severe, means that its decline has been less rapid. **Criteria nearly met:** A1c,d; A2c,d.

MAROON-CHINNED FRUIT-DOVE *Ptilinopus subgularis* is restricted to the Sulawesi Endemic Bird Area (not the South or South-East Peninsulas) and Banggai and Sula Islands Endemic Bird Area, **Indonesia**, where it is a moderately common inhabitant mainly of primary but also some secondary forest from the lowlands only up to 800 m (race *epia*) on Sulawesi, a little higher (900 m) in the Banggai Islands (race *subgularis*) and Sula Islands (race *mangoliensis*) (Stattersfield *et al*. 1998; also Stresemann and Heinrich 1939–1941, Coates and Bishop 1997, Gibbs in press). Forest destruction within its elevation range has been extensive in recent decades (see Threats under Lombopatang Flycatcher *Ficedula bonthaina*), and its populations must have suffered a commensurate decline. **Criteria nearly met:** A1c; A2c.

BLUE-CAPPED FRUIT-DOVE *Ptilinopus monacha* is restricted to the Northern Maluku Endemic Bird Area, **Indonesia**, where it inhabits lowland forest, mangroves and farmland at

0–750 m (Stattersfield *et al*. 1998). It is known from at least 13 islands in North Maluku and is moderately common, but is mainly a small-island and coastal species (Coates and Bishop 1997) and is therefore likely to be suffering from substantial habitat loss within its area of occupancy (e.g. for Halmahera see Threats under Sombre Kingfisher *Todiramphus funebris*). **Criterion nearly met:** C1.

PINK-BELLIED IMPERIAL-PIGEON *Ducula poliocephala* is endemic to the **Philippines** on at least 15 islands (Basilan, Biliran, Catanduanes, Cebu, Dinagat, Leyte, Luzon, Masbate, Mindanao, Mindoro, Negros, Panay, Samar, Sibuyan and Tawitawi) but is uncommon (Dickinson *et al*. 1991), local, perhaps semi-nomadic and certainly declining owing to the combination of heavy hunting and the extensive clearance of its preferred (though not exclusive, as the bird ranges up to 1,500 m) lowland forest habitat (R. J. Timmins *in litt*. 1994, G. C. L. Dutson *in litt*. 1997). It is extinct on Cebu but was recently found on Bohol (Brooks *et al.* 1995a). **Criteria nearly met:** A1c,d; A2c,d; C2a.

PINK-HEADED IMPERIAL-PIGEON *Ducula rosacea* is restricted to four Endemic Bird Areas (Northern Nusa Tenggara; Timor and Wetar; Banda Sea Islands; Northern Maluku) and five Secondary Areas (Seribu Islands; Masalembu; Kangean; Salayar and Bonerate Islands; Tukangbesi Islands), where it occurs in forest, scrub and farmland at 0–600 m (Stattersfield *et al*. 1998). Despite this wide range, the species appears to have become very rare at least in some areas (B. F. King verbally 1998), although not necessarily owing to hunting, at least on Timor (R. Noske *in litt*. 2000). **Criterion nearly met:** C1.

BLUE-STREAKED LORY *Eos reticulata* is restricted to the Banda Sea Islands Endemic Bird Area, **Indonesia**, where it inhabits forest, coconut plantations and mangroves on Tanimbar (Yamdena and Larat), with populations reported on Babar, Damar and Kai in the late nineteenth century but not since (Stattersfield *et al*. 1998; also Inskipp *et al*. 1988, Coates and Bishop 1997). An average 3,198 birds were traded annually in the period 1983–1989, but fieldwork in 1993 on Yamdena revealed an estimated 220,000 ± 52,000 birds, and past catch rates were judged unlikely to have caused any decline (del Hoyo *et al*. 1997, Bishop and Brickle 1998). However, habitat loss continues apace in the south of Yamdena (Bishop and Brickle 1998) and this, combined with continuing trapping, may cause it to decline in the future. **Criteria nearly met:** A2c,d; B1+2c,e.

MINDANAO LORIKEET *Trichoglossus johnstoniae* is endemic to Mindanao, **Philippines**, where it occurs at Mt Kitanglad (Ripley and Rabor 1961, Mallari and Gonzalez 1993, many observers *in litt*., plus 61 specimens in AMNH, ANSP, FMNH, PNM, UMMZ, USNM, ZMC); near Lake Lanao (Hachisuka 1930); Mt Piapayungan (13 specimens in DMNH); Mt Apo (Ogilvie-Grant 1906b, many observers *in litt*.); Mt Matutum (eight specimens in FMNH, UMMZ, USNM); Lake Parker (male in PNM); Luhan, New Dumangas, T'boli, South Cotabato (Krupa *et al.* 1985); Lake Sebu (C. R. Robson *in litt*. 1994, P. A. J. Morris *in litt.* 1996); and Mt Malindang (Rand and Rabor 1960). It was described as uncommon and local (Dickinson *et al.* 1991; see also Goodfellow 1906) and treated as threatened (Collar *et al*. 1994), but now appears to be moderately secure and numerous in montane forest above around 1,000 m (various observers), this being habitat which is unlikely to be greatly affected by human activities in the medium term. **Criterion nearly met:** B1+2c.

IRIS LORIKEET *Psitteuteles iris* is restricted to the Timor and Wetar Endemic Bird Area, **Indonesia**, where it occurs on both Timor and Wetar at 0–1,500 m in monsoon forest (Stattersfield *et al*. 1998). Although there appear to be no records from East Timor, and the species is sometimes trapped, it is still in reasonable numbers in East Timor (Noske 1995) and

it is or at least once was locally not uncommon on Wetar (Coates and Bishop 1997). It was, however, the least common of the three lorikeets on Timor in 1995, although many were not identified to species (Verbelen 1996). It is included as threatened in the *Parrot Action Plan* (Snyder *et al.* 2000) and its status certainly requires monitoring. **Criterion nearly met:** C1.

STRIATED LORIKEET *Charmosyna multistriata* is a poorly known New Guinea species, ranging from the Snow mountains of Papua (formerly Irian Jaya), **Indonesia**, to Crater mountain in **Papua New Guinea**, where records are from very few localities but assessments of its range and populations are hindered by its presumed nomadic habits (Coates 1985, Beehler *et al.*1986, Gregory 1995a). There are no recent records from Papua (K. D. Bishop *in litt.* 1997). In Papua New Guinea, there are recent records from Crater mountain (Mack and Wright 1996), and frequent records of up to 125 birds around Ok Tedi (Gregory 1995a). It inhabits lower montane forest at 80–1,770 m (Coates 1985), which is locally under threat from logging and clearance for agriculture, but although it may have a small overall population, it is not believed to be declining rapidly. **Criterion nearly met:** C1.

TANIMBAR COCKATOO *Cacatua goffini* is restricted to the Banda Sea Islands Endemic Bird Area, **Indonesia**, where it inhabits forest and agricultural land on Tanimbar (Yamdena, Larat) with an introduced population on Kai (Stattersfield *et al.* 1998). In the 1980s the numbers of birds entering trade annually exceeded 10,000 and there were fears for its security, but this exploitation proved to be in retaliation for crop depredations from a global population estimated at some 255,000 birds (Cahyadin 1993, 1996, del Hoyo *et al.* 1997). Nevertheless, habitat loss continues apace in the south of Yamdena (Bishop and Brickle 1998) and this, combined with continuing trapping, must be producing a decline. **Criteria nearly met:** A1c,d; A2c,d; B1+2c.

GEELVINK PYGMY-PARROT *Micropsitta geelvinkiana* is endemic to Numfor and the twin islands of Biak-Supiori off Papua, **Indonesia**, where it occurs in forest, secondary forest and forest gardens, up to at least 400 m (Mayr and Meyer de Schauensee 1939, Bishop 1982, Beehler *et al.* 1986, K. D. Bishop *in litt.* 2000). Recent visitors to Biak-Supiori have found it to be widespread but only in small numbers (Gibbs 1993, Poulsen and Frolander 1994, Eastwood 1996, B. M. Beehler and SvB *in litt.* 2000, M. van Beirs *in litt.* 2000), although it is also described as well distributed in good numbers (in suitable habitat) on Biak-Supiori (del Hoyo *et al.* 1997). On Biak and Numfor, forest is under heavy threat from logging and subsistence farming, but there appear to be large areas of forest remaining in interior Supiori (Bishop 1982, K. D. Bishop *in litt.* 1996, D. A. Holmes *in litt.* 2000). This species tolerates highly degraded habitats and occurs in the 110 km^2 Biak-Utara protected area (B. M. Beehler and SvB *in litt.* 2000), suggesting that it is not threatened. However, there is very little recent information on bird and especially forest status on Biak-Supiori, and further research may show that this species is declining rapidly. **Criteria nearly met:** C1; C2a.

BLUE-RUMPED PARROT *Psittinus cyanurus* is confined to the Sundaic lowlands, where it is known from southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, **Singapore**, Kalimantan, Sumatra (including the Riau and Lingga archipelagos, Bangka, Simeulue and the Mentawai islands), **Indonesia** and **Brunei** (uncommon), to 1,300 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). It inhabits primary, dryland evergreen and semi-evergreen lowland forest, both mature and selectively logged, and also visits edge vegetation, cultivated areas and gap-phase growth of forest clearings and occasionally mangroves (Wells 1999). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*).

However, the species's ability to persist in second growth and at higher elevations, where forest destruction has been less severe, means that its decline has been less rapid. **Criteria nearly met:** A1c; A2c.

LUZON RACQUET-TAIL *Prioniturus montanus* is endemic to Luzon, **Philippines**, where it is almost entirely confined to montane forest regions (above c.700 m) in the Cordillera Central and the Sierra Madre, with records including Mt Sicapo-o at Mt Simminublak (three specimens in FMNH); Liwan, Kenema, Mountain Province (male in DMNH); Sablan, Mountain Province (two females in FMNH); Massisiat, Abra (two males in FMNH); Mt Data (Ogilvie-Grant 1895c, with later specimens in AMNH, DMNH); Mt Pulog (McGregor 1910b); Mt Polis (Hornbuckle 1994); Mt Puguis (Morioka and Sakane 1979); Paoay ("Haights in the Oaks") (12 specimens in USNM); Irisan, Benguet (four specimens in FMNH, USNM; also McGregor 1904); Imugan, Nueva Vizcaya (female in DMNH); and three sites—Dipalayag, Los Dos Cuernos and Mt Cetaceo—in the Sierra Madre (Danielsen *et al.* 1994, Poulsen 1995); plus Pangil, Laguna (male in DMNH). The fact that it escaped detection in the Sierra Madre until 1991 is evidence that it lives in relatively inaccessible areas and its numbers may therefore not be in rapid decline; there is, however, a need for vigilance concerning habitat destruction, hunting and trapping for the cagebird trade (Danielsen *et al.* 1994, Poulsen 1995). **Criterion nearly met:** B1+2c.

MINDANAO RACQUET-TAIL *Prioniturus waterstradti* is endemic to Mindanao, **Philippines**, where it is known from nine montane localities, as follows: Mt Hilong-hilong (three specimens in DMNH, USNM); Mt Mayo (six specimens in CM, FMNH, USNM); Anakan and Civolig near Gingoog City (Meyer de Schauensee and duPont 1962); Mt Kitanglad (Ripley and Rabor 1961, Gonzalez and Mallari 1993, Heaney *et al.* 1993, many observers *in litt.*); Mt Apo (Rothschild 1904, Hartert 1906, Hachisuka 1931–1935, plus many modern specimens and observers); Mt Matutum (11 specimens in FMNH, USNM); Lake Sebu (P. A. J. Morris *in litt.* 1996); and Mt Malindang (Rand and Rabor 1960). The notion that it is "certainly local and uncommon, apparently occurring at lower density than some of its congeners" (Collar *et al.* 1994) appears to be over-cautious, with evidence to the contrary coming from several sources old and new (Hachisuka 1931–1935, Rand and Rabor 1960, Gonzalez and Mallari 1993, Heaney *et al.* 1993); moreover, the montane forest where the species lives (extreme elevational records are 820–2,700 m: BirdLife database) is relatively secure compared to lower-altitude growth. **Criterion nearly met:** B1+2c.

YELLOWISH-BREASTED RACQUET-TAIL *Prioniturus flavicans* is known with certainty only from the eastern two-thirds of the northern peninsula in the Sulawesi Endemic Bird Area, **Indonesia**, with populations on adjacent islands including the Togian Islands (Stattersfield *et al.* 1998, Walker and Cahill 2000). On the peninsula the extent of its lowland forest habitat is c.11,300 km^2, and the world population is in the order of 45,000 birds, but its numbers are declining in the face of "high rates" of habitat destruction, degradation and fragmentation (Walker and Cahill 2000). On this basis the species closely approaches the range-size threshold of under 20,000 km^2 and its rate of decline may be close to 20% in the past 10 years and future 10 years. **Criteria nearly met:** A1c,d; A2c,d; B1+2a,b,c,d,e.

BLUE-NAPED PARROT *Tanygnathus lucionensis* is confined to the **Philippines**, where there are records from at least 45 islands, plus the Talaud Islands, **Indonesia**, and islands off north-east Borneo belonging to **Malaysia** (Dickinson *et al.* 1991). As a bird of closed and open forest formations, including second growth, coconut plantations, banana patches and mangrove, chiefly in lowland and coastal areas (Rabor 1976, BRT), it was common on many islands in the Philippines a century or less ago (e.g. Tweeddale 1877a, Whitehead 1890, 1899c,

Zimmer 1918b), but has suffered from trapping (for both domestic and international trade) and habitat loss on such a scale that it is now rare or extinct on many islands (see, e.g., Brooks *et al*. 1992, Dutson *et al*. 1992, Evans *et al*. 1993a). However, while not a "small-island specialist" it does have a strong capacity to survive in small pockets of habitat on the smaller islands, so that its status overall is difficult to assess; moreover, it is still fairly numerous in some areas of Palawan and on Tawitawi (Lambert 1993c, D. Allen verbally 1997), and high numbers persist in a large tract of forest on Talaud (Riley 1997a,b). **Criterion nearly met:** C2a.

OLIVE-SHOULDERED PARROT *Aprosmictus jonquillaceus* is restricted to the Timor and Wetar Endemic Bird Area, **Indonesia** and **East Timor**, where it occurs on Timor, Wetar and Roti at 0–2,600 m in monsoon forest and acacia savanna (Stattersfield *et al*. 1998; also White and Bruce 1986, Forshaw 1989, Coates and Bishop 1997). The total population has been put at around 10,000 and judged probably stable (Juniper and Parr 1997), but habitat destruction (possibly compounded by trade, with 1,343 birds officially recorded as being exported from Indonesia, 1981–1985: Inskipp *et al*. 1988) appears to be responsible for a steady decline in numbers in the past 15 years (B. F. King verbally 1998). **Criterion nearly met:** C1.

RED-BILLED HANGING-PARROT *Loriculus exilis* is restricted to the Sulawesi Endemic Bird Area (but not the East or South Peninsulas), **Indonesia**, where it inhabits primary lowland and hill forest and mangroves up to 1,000 m (Stattersfield *et al*. 1998; also White and Bruce 1986, Coates and Bishop 1997). Although it is clearly inconspicuous and easily overlooked, records are so few that it is characterised as generally very uncommon and local, and it is apparently nomadic (Coates and Bishop 1997), so that the extensive forest destruction within its elevation range in recent decades (see Threats under Lombopatang Flycatcher *Ficedula bonthaina*) may have caused a significant decline in its population. **Criteria nearly met:** A1c; A2c.

YELLOW-THROATED HANGING-PARROT *Loriculus pusillus* occurs in Java and Bali, **Indonesia**, where it inhabits lowland, swamp and montane forests to 2,000 m (MacKinnon and Phillipps 1993). It has declined markedly in West Java, but appears to tolerate heavily degraded forest (as long as some tall trees are left) on Bali (SvB). Forest destruction in the lowlands of Java and Bali has been extremely extensive (see Threats under Javan Hawk-eagle *Spizaetus bartelsi*). However, the species's ability to tolerate degraded habitats and to persist at higher elevations, where forest destruction has been less severe, means that its decline has been less rapid. **Criteria nearly met:** A1c; A2c.

NICOBAR PARAKEET *Psittacula caniceps* is endemic to the Nicobar archipelago, **India**, where it inhabits tall forest on Great Nicobar, Little Nicobar, Menchal and Kondul islands, feeding in small groups in the canopy on the fruit of *Pandanus* palms (Grimmett *et al.* 1998). It is apparently common, but fairly large numbers are trapped for the cagebird trade (del Hoyo *et al.* 1997). Furthermore, increased settlement of the islands has led to increased pressure on natural resources, and planned development projects could severely affect the habitat of this species (Stattersfield *et al.* 1998). **Criteria nearly met:** B1+2a,b,c,d,e; C1.

LONG-TAILED PARAKEET *Psittacula longicauda* occurs in the Andaman, Cocos and Nicobar islands (to **India**), peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, **Singapore**, Kalimantan (including the Natuna islands), Sumatra (including the Riau archipelago), **Indonesia** and **Brunei**, in coastal and lowland areas to at least 300 m (Smythies 1981, Mann 1987, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000), preferring extreme lowland swamp (including peatswamp) forest in the Thai-Malay Peninsula (Wells 1999),

although it avoids primary forest in Borneo (Smythies 1981). Swamp-forest destruction in the Sundaic lowlands has been extensive (see Threats under Hook-billed Bulbul *Setornis criniger*), but the species remains numerous in a number of areas owing to its capacity to forage away from forested areas and nest communally (Wells 1999). **Criteria nearly met:** A1c,d; A2c,d.

MOUSTACHED HAWK-CUCKOO *Cuculus vagans* is confined to the Sundaic lowlands, in southern Tenasserim, **Myanmar**, peninsular **Thailand**, southern **Laos**, Sabah, Sarawak and Peninsular **Malaysia**, **Brunei**, Kalimantan, Sumatra and, occasionally, West Java, **Indonesia**, where it is generally uncommon in evergreen forest to 915 m (Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Duckworth *et al*. 1999, Wells 1999, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*). The species's preference for forest edge and secondary forest (King *et al.* 1975, MacKinnon and Phillipps 1993) and use of mid-altitude forest implies that it is not immediately threatened; however, it is commonest in plains-level forest, and thus may become threatened in future (Wells 1999). **Criteria nearly met:** A1c; A2c.

MOLUCCAN CUCKOO *Cacomantis heinrichi* is restricted to the Northern Maluku Endemic Bird Area, **Indonesia**, where it inhabits montane forest at 1,000–1,500 m on Halmahera, lower (800–1,200 m) on Bacan (Stattersfield *et al*. 1998; also Stresemann 1931, Heinrich 1956), with recent reports from as low as 100–450 m (D. Gibbs and R. Bright *per* MKP, specimen from Kobe in MZB). **Criterion nearly met:** C2b.

BLACK-BELLIED MALKOHA *Phaenicophaeus diardi* occurs in southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct: Lim Kim Seng *in litt*. 2001), **Brunei**, and Kalimantan and Sumatra, **Indonesia**, where it inhabits primary or secondary evergreen forest or mangroves to 1,220 m (Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*). However, the species's ability to persist in second growth and at higher elevations, where forest destruction has been less severe, means that its decline has been less rapid. **Criteria nearly met:** A1c; A2c.

CHESTNUT-BELLIED MALKOHA *Phaenicophaeus sumatranus* occurs from southern Tenasserim, **Myanmar**, through peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, **Singapore** and **Brunei** (uncommon), to Kalimantan (including the North Natuna islands) and Sumatra, **Indonesia**, in primary and secondary forest, including mangroves and peatswamp forest, to 1,000 m (Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*). However, the species's ability to persist in second growth and at higher elevations, where forest destruction has been less severe, means that its decline has been less rapid. **Criteria nearly met:** A1c; A2c.

BORNEAN GROUND-CUCKOO *Carpococcyx radiatus*, now treated separately from the threatened Sumatran Ground-cuckoo *C. viridis*, is endemic to the island of Borneo (**Brunei**, Sabah and Sarawak, **Malaysia**, and Kalimantan, **Indonesia**) (Collar and Long 1995), where it inhabits the floor of mostly lowland forest (Thompson 1966, Davison 1979b). While this habitat

is under severe pressure in Borneo, and although this cuckoo is widely regarded as scarce to very rare (Hose 1893, Finsch 1905, Gore 1968, Fogden 1976), it has proved to be widespread and common in many areas (including secondary habitats and lower hill country) surveyed for pheasants and, as they are generally regarded as uneatable, natives who catch them in snares for other species let them go (R. Sözer verbally 1999). **Criteria nearly met:** A1c; A2c.

BIAK COUCAL *Centropus chalybeus* is endemic to the twin islands of Biak-Supiori, off Papua (formerly Irian Jaya), **Indonesia** (Mayr and Meyer de Schauensee 1939, Beehler *et al.* 1986). Recent observations indicate that it is probably not uncommon in forest and secondary forest, but elusive and more often heard than seen, and probably commoner on Supiori (Bishop 1982, Gibbs 1993, K. D. Bishop *in litt.* 1996, 2000, Eastwood 1996, B. M. Beehler and SvB *in litt.* 2000, M. van Beirs *in litt.* 2000). On Biak, forest is under heavy threat from logging and subsistence farming, but there appear to be large areas of forest remaining in interior Supiori (Bishop 1982, K. D. Bishop *in litt.* 1996). The species's tolerance of moderately degraded forest, and its occurrence in the 110 km^2 Biak-Utara protected area (B. M. Beehler and SvB *in litt.* 2000), suggest that it is not threatened. However, there is very little recent information on bird or forest status on Biak-Supiori, and further research may show that this species is declining rapidly. **Criterion nearly met:** C1.

RUFOUS COUCAL *Centropus unirufus* is endemic to Luzon and its satellites Polillo and Catanduanes, **Philippines**, where it is uncommon and local (Dickinson *et al.* 1991). It appears to be a bird of lowland forest (R. J. Timmins *in litt.* 1994) and therefore merits attention owing to the actual and potential destruction of this habitat (see, e.g., Threats *Habitat loss: Luzon* under Philippine Eagle *Pithecophaga jefferyi*), but it persists well in mature secondary growth, selectively logged and degraded forest, and even prefers bamboo areas (N. Bostock *in litt.* 1994, Gonzalez 1995, P. Davidson *in litt.* 1996). **Criteria nearly met:** A1c; A2c.

ENGGANO SCOPS-OWL *Otus enganensis* is endemic to Enggano island (Enggano Endemic Bird Area), off south-west Sumatra, **Indonesia**, where it occupies forest edge and wooded areas (MacKinnon and Phillipps 1993, Stattersfield *et al.* 1998). Reports on its status differ, from very rare to quite uncommon (del Hoyo *et al.* 1999). Habitat destruction on Enggano has, thus far, been relative slight ("quite extensive forest": Holmes 1994), but recent tentative proposals for agricultural development suggest the possibility for future radical change (del Hoyo *et al.* 1999), and this may exert pressure on the small population of this species. **Criteria nearly met:** B1+2a,b,c,d,e; C1; C2b.

REDDISH SCOPS-OWL *Otus rufescens* occurs in the Sundaic lowlands, in south peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan, Sumatra (including Bangka Island) and Java, **Indonesia** and **Brunei**, in primary and tall secondary forest, including logged forest, to 1,000 m (King *et al.* 1975, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*). However, the species's ability to persist in second growth and at higher elevations, where forest destruction has been less severe, means that its decline has been less rapid, but it is likely to be close to extinction in Thailand (Wells 1999). **Criteria nearly met:** A1c; A2c.

ANDAMAN SCOPS-OWL *Otus balli* is an endemic resident in the Andaman islands, **India**, where it was common, at least early in the twentieth century, in trees in semi-open or cultivated areas and around human settlements (Grimmett *et al.* 1998, del Hoyo *et al.* 1999). Its current

status is unclear, although it appears to be easily found and therefore probably common (del Hoyo *et al.* 1999). There seems little reason to expect its population to be under immediate threat given its tolerance of disturbed areas; however, forest loss is accelerating on the Andamans and further research is required to clarify its ecological requirements, population size and trends (Stattersfield *et al.* 1998, del Hoyo *et al.* 1999). **Criteria nearly met:** B1+2a,b,c,d,e.

SIMEULUE SCOPS-OWL *Otus umbra* is endemic to Simeulue Island, off north-west Sumatra, **Indonesia**, where it favours forest edge and remnants, especially on coasts, as well as clove plantations (MacKinnon and Phillipps 1993). Reports as to its status differ, from rare, or possibly rare, to not uncommon (del Hoyo *et al.* 1999). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010 (see Threats under Crestless Fireback *Lophura erythrophthalma*), and although in 1989 there was still 60% forest cover, "rapid clearance for the cultivation of cloves has been reported" (Holmes 1994). Against this, the tolerance of secondary and edge habitats shown by this species renders it unlikely to be under immediate threat. **Criteria nearly met:** B1+2a,b,c,d,e; C1; C2b.

LUZON SCOPS-OWL *Otus longicornis* is endemic to Luzon, **Philippines**, where it is recorded from rain- and pine forest, extending to at least 2,200 m, in the provinces of Ilocos Norte, Benguet, Cagayan, Isabela, Nueva Ecija, Bulacan, Quezon, Camarines Norte and Camarines Sur (Dickinson *et al.* 1991, Collar *et al.* 1994, Hornbuckle 1994, Poulsen 1995, Mallari and Diesmos in press). Despite its fairly wide distribution and elevation tolerance, the species is generally uncommon and appears to be sensitive to habitat alteration in the lowlands (Dickinson *et al.* 1991, N. Bostock verbally 1993, Poulsen 1995). **Criteria nearly met:** A1c; A2c.

MINDORO SCOPS-OWL *Otus mindorensis* is endemic to Mindoro, **Philippines**, where it is now judged likely to occur throughout the mountains in the centre of the island, and it appears to be common (*contra* Dickinson *et al.* 1991) above c.1,000 m in closed-canopy montane forest (Collar and Andrew 1988, Dutson *et al.* 1992, Evans *et al.* 1993a), which is being lost or degraded at a relatively low rate (BRT). Very recently, the species was commonly observed in Mt Iglit-Baco National Park at 700–900 m, in patches of highly fragmented lowland second-growth forest that are confined to this elevation band (A. T. L. Dans and J. C. T. Gonzalez verbally 1995). **Criterion nearly met:** C2a

MINDANAO SCOPS-OWL *Otus mirus* is endemic to Mindanao, **Philippines**, where it has been recorded from Mt Hilong-hilong (Ripley and Rabor 1968); Mt Apo (Dickinson *et al.* 1991); Mt Kitanglad (Gonzalez and Mallari 1993, Heaney *et al.* 1993, Collar *et al.* 1994); and Lake Sebu (*per* G. C. L. Dutson *in litt.* 1994). Although described as rare (Dickinson *et al.* 1991), this is a species of higher-elevation forest, which is being lost at a relatively low rate, and it is probably therefore not seriously threatened (C. R. Robson *in litt.* 1994, BRT). **Criterion nearly met:** B1+2c.

ELEGANT SCOPS-OWL *Otus elegans* is found on the Nansei Shoto islands, southern **Japan**, on Lanyu island, off south-east **Taiwan** (China), and on the Batanes and Babuyan islands off northern Luzon, **Philippines**, where it occurs in subtropical evergreen forest, and locally in or near to villages, from sea-level to 550 m or higher (del Hoyo *et al.* 1999). It is common wherever suitable habitat remains on the Nansei Shoto, and is presumed to have quite a large population there (Brazil 1991); it has a population estimated at c.1,000 birds on Lanyu island (del Hoyo *et al.* 1999), and it has been described as fairly common on the Batanes and Babuyan islands (Dickinson *et al.* 1991). However, its range must have been much reduced and fragmented in the Philippines by deforestation (L. L. Severinghaus verbally 1999), although its population is

thought to be stable on Lanyu and its prospects for survival there are good so long as suitable habitat is protected (del Hoyo *et al.* 1999) it is presumably also relatively secure on the Nansei Shoto (see Threats under Ryukyu Woodcock *Scolopax mira*). **Criterion nearly met:** B1+2c.

MANTANANI SCOPS-OWL *Otus mantananensis* is endemic to islets off Sabah (e.g. Mantanani), **Malaysia**, and off Palawan (e.g. Rasa, Ursula), **Philippines**, as well as to small islands in the Sulu archipelago and central Philippines, where it is apparently common in coconut groves and in other areas with trees (Dickinson *et al*. 1991). Despite this assessment, the total extent of its range is small, and it is confined to increasingly disturbed and developed islands such that its population may be incurring substantial declines (del Hoyo *et al.* 1999, W. L. R. Oliver verbally 1999). **Criteria nearly met:** C1; C2a.

MENTAWAI SCOPS-OWL *Otus mentawi* is endemic to the larger islands of the Mentawai group, off west Sumatra, **Indonesia**, where it occurs in lowland forest and second growth, including around villages (MacKinnon and Phillipps 1993). Its status is poorly known, but it may be locally common (del Hoyo *et al*. 1999). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010 (see Threats under Crestless Fireback *Lophura erythrophthalma*); although there is much damage on Siberut (Holmes 1994) and further deforestation looms (see Threats under Storm's Stork *Ciconia stormi*), extensive forest remains on the islands as a whole (Holmes 1994) and the species's ability to persist in human-modified habitats suggests that its decline may not be rapid. **Criteria nearly met:** B1+2a,b,c,d,e; C1.

PALAWAN SCOPS-OWL *Otus fuliginosus* is endemic to Palawan (Dickinson *et al.* 1991) and the adjacent islands of Alabagin (specimen in PNM) and Balabac (male from Bulan-bulan in DMNH), with an as-yet unconfirmed report from Calauit (J. C. T. Gonzalez verbally 1995), **Philippines**. Records on Palawan include Linapacan at Kinalaykayan and Dicabaitot (four specimens in PNM); St Paul's Subterranean River National Park (Collar *et al.* 1994); Cleopatra's Needle (male in PNM); Buenavista (female in DMNH); Iwahig Penal Colony (Collar *et al.* 1994); vicinity of Puerto Princesa (Sharpe 1888); Quezon at Tabon (two specimens in DMNH, PNM); Singnapan at Kabasakan, Pinikpikan and Tining-luan (specimens in PNM; see also Sison 1983); Taguso (Dickinson *et al.* 1991); Mt Mantalingahan at Pinigisan (male in DMNH); and Tigwayan, Batarasa (three specimens in DMNH). Although described as rare (Dickinson *et al.* 1991) and evidently a bird of lowland forest, which is being rapidly cleared, the emerging evidence from knowledge of its voice is that it is much commoner and widerspread than supposed, and can adapt to mixed cultivation and plantations (Sison 1983). **Criteria nearly met:** B1+2c; C2b.

CHESTNUT-BACKED OWLET *Glaucidium castanonotum* is endemic to **Sri Lanka**, where it is a rare resident in dense wet forests of the lowlands and hills (Grimmett *et al.* 1998), also occurring in logged forests, scrub and cultivation (Kotagama and Fernando 1994, Jones *et al.* 1998). It is shy and retiring, generally keeping to the canopy of large trees, and is therefore probably overlooked (Grimmett *et al.* 1998). However, while it may thus be commoner than records suggest, its range has diminished dramatically since the nineteenth century when it was found throughout lowland Sri Lanka to the outskirts of Colombo (Legge 1880, Jones *et al.* 1998). Forest on the island has suffered rapid degradation and fragmentation in the past decades through excessive gathering of fuelwood, clearance for permanent agriculture, shifting cultivation, fire, urbanisation and logging (Stattersfield *et al.* 1998). Closed-canopy forest is estimated to have declined from 29,000 km^2 (44% of the island's area) in 1956 to 12,260 km^2 in 1983 (Collins *et al.* 1991). It is feared that this loss will continue. **Criteria nearly met:** B1+2a,b,c,d,e; C1.

SUMBA BOOBOOK *Ninox rudolfi* is restricted to the Sumba Endemic Bird Area, **Indonesia**, where it occurs at up to 1,000 m in primary, disturbed primary and secondary forest and forest edge, in both deciduous and evergreen formations (Stattersfield *et al*. 1998; also M. J. Jones *et al*. 1995). Although characterised as "uncommon or rare", occurring singly, in pairs or small dispersed groups of up to four birds (Coates and Bishop 1997), recent observations on Sumba have revealed that the species is widespread and moderately common, although still at some risk from steady habitat loss (B. F. King verbally 1998, S. J. Marsden verbally 1998). **Criteria nearly met:** C1; C2b.

ANDAMAN HAWK-OWL *Ninox affinis* is endemic to the Andaman and Nicobar archipelagos, India, where it occurs in mangrove forest, lightly wooded areas and forest clearings, apparently hawking insects at dusk (Grimmett *et al.* 1998). Although its tolerance of degraded habitats gives cause for optimism, the human population on larger islands in the Andaman group is rising rapidly and habitat is consequently under severe pressure from agriculture, grazing and logging (Whitaker 1985, Curson, 1989, Pande *et al.* 1991, Sinha 1992, Stattersfield *et al.* 1998). **Criteria nearly met:** B1+2a,b,c,d,e.

OCHRE-BELLIED HAWK-OWL *Ninox ochracea* is restricted to the Sulawesi Endemic Bird Area (but not the South Peninsula), **Indonesia**, where it is generally uncommon in primary and tall secondary lowland forest, riverine and lower montane forest up to 1,000 m (Stattersfield *et al*. 1998, Rasmussen 1999 *contra* 1,780 m in Coates and Bishop 1997), with an assertion that it is chiefly a species of drier formations (Stresemann and Heinrich 1939–1941). Forest destruction within its elevation range has been extensive in recent decades (see Threats under Lombopatang Flycatcher *Ficedula bonthaina*), and its populations must have suffered a commensurate decline. **Criteria nearly met:** C1; C2a.

LARGE FROGMOUTH *Batrachostomus auritus* occurs very sparsely from south peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan (including the Natuna and Labuan islands) and Sumatra, **Indonesia** and **Brunei** in evergreen and secondary forest, to at least 250 m, perhaps 1,000 m (Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Cleere 1998, Wells 1999, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but this bird's use of and even preference for regenerating and second-growth areas (Wells 1999) places it at less risk. **Criteria nearly met:** A1c; A2c.

DULIT FROGMOUTH *Batrachostomus harterti* is restricted to the Bornean Mountains Endemic Bird Area in hill and lower montane forest at 300–1,500 m (Stattersfield *et al*. 1998; also Smythies 1981). Discounting its erroneous listing for Tanjung Puting National Park (MacKinnon *et al*. 1996), it is known by only eight specimens from four areas in Sarawak, **Malaysia**, and one in West Kalimantan, **Indonesia** (Cleere 1998), and there is steady habitat loss on Mt Dulit itself (B. F. King verbally 1998). **Criteria nearly met:** A1c; A2c.

GOULD'S FROGMOUTH *Batrachostomus stellatus* occurs from central peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct: Lim Kim Seng *in litt*. 2001), Kalimantan (including the North Natuna islands) and Sumatra (including the Riau and Lingga archipelagos and Bangka island), **Indonesia** and **Brunei**, where it is fairly widespread and common in evergreen and secondary forest to at least 500 m (once to 920 m) (Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary

formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia, but its ability to breed in second-growth places it at less risk (del Hoyo *et al.* 1999). **Criteria nearly met:** A1c; A2c.

SHORT-TAILED FROGMOUTH *Batrachostomus poliolophus* is endemic to Sumatra, **Indonesia**, where it is known from a small number of records from the Barisan range, in forest between 660 and 1,400 m, but is presumably greatly under-recorded because it is nocturnal (Cleere 1998, del Hoyo *et al.* 1999). Deforestation in the lower part of its altitudinal range could be a threat (see Threats under Crestless Fireback *Lophura erythrophthalma* and Sumatran Ground-cuckoo *Carpococcyx viridis*). **Criteria nearly met:** A1c; A2c.

BORNEAN FROGMOUTH *Batrachostomus mixtus* (when split from *B. poliolophus*) is restricted to the Bornean Mountains Endemic Bird Area, where it occurs in Sabah and Sarawak, **Malaysia**, and Kalimantan, **Indonesia**, in lower montane forest at 600–1,675 m (Stattersfield *et al.* 1998; also Cleere 1998). It is thought "probably not common" (Cleere 1998). **Criteria nearly met:** A1c; A2c.

SALVADORI'S NIGHTJAR *Caprimulgus pulchellus* is restricted to the Sumatra and Peninsular Malaysia Endemic Bird Area and the Java and Bali Forests Endemic Bird Area, occurring only in Sumatra (South Barisan mountains) and Java, **Indonesia**, in montane forest at 1,350–2,200 m or lower, 800–1,800 m (Stattersfield *et al.* 1998; also Collar and Andrew 1988, van Marle and Voous 1988, Holmes 1996). Although previously reckoned to be Data Deficient, reassessment of evidence suggests that it is likely to be moderately secure in small numbers at higher elevations but losing habitat and populations at lower ones where encroachment is occurring (see, e.g., Threats under Crestless Fireback *Lophura erythrophthalma* and Sumatran Ground-cuckoo *Carpococcyx viridis*). **Criteria nearly met:** C1; C2a.

WATERFALL SWIFT *Hydrochous gigas* is restricted to the Sumatra and Peninsular Malaysia Endemic Bird Area and the Java and Bali Forests Endemic Bird Area, occurring in the Main Range, West **Malaysia**, and Sumatra and West Java, **Indonesia**, in hill and montane forest and over open country at 300–2,400 m, but strongly associated with waterfalls as nest sites (Stattersfield *et al.* 1998, Chantler 2000). Records from Borneo, in Sarawak and Sabah, Malaysia, and Brunei, involve no specimen evidence (see Smythies 1981, Chantler 2000) and are thus treated provisionally here. The individual populations appear to be small, although one, at Gede-Pangrango, was once fairly large (Andrew 1985) but is seemingly now vanished (Chantler 2000). **Criterion nearly met:** C1.

VOLCANO SWIFTLET *Collocalia vulcanorum* is restricted to the Java and Bali Forests Endemic Bird Area, nesting in crevices at 2,200–3,000 m in West Java, **Indonesia**, and foraging in airspace around peaks and ridges of volcanoes (Stattersfield *et al.* 1998). Certain records are from four sites in West Java, with likely but unconfirmed records from five sites in West, Central and East Java (SvB); taking an average colony size of 25 pairs and assuming only one colony per site, this would make a total of under 400 birds for all localities, but this may be highly precautionary and an additional 4–5 craters may also contain the species (SvB). As the species nests in crater crevices and all known localities are active volcanoes, colonies are believed to be susceptible to periodic extinction (MacKinnon and Phillipps 1993). **Criteria nearly met:** A2c; D1.

PHILIPPINE NEEDLETAIL *Mearnsia picina* is endemic to the **Philippines** where it has been described as fairly common on Mindanao, Samar, Leyte, Biliran, Cebu and Negros (Dickinson *et al.* 1991), with a recent record from Tawitawi (Lambert 1993c), but it now

appears to be scarce and local even in the best sites, presumably as a consequence of the widespread reduction in its lowland forest habitat (G. C. L. Dutson *in litt*. 1996, Chantler 2000). **Criteria nearly met:** A1c; A2c; C1.

RED-NAPED TROGON *Harpactes kasumba* is confined to the Sundaic lowlands, where it is known from peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct), Kalimantan, **Indonesia** and **Brunei** in many forest types, but most commonly in primary or lightly logged formations, to 1,200 m (Smythies 1981, Mann 1987, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000, del Hoyo *et al*. 2001). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysian Borneo (see Threats under Crestless Fireback *Lophura erythrophthalma*), but it appears to be able to survive, to some extent, in logged forest (del Hoyo *et al*. 2001), a factor that increases its long-term security. **Criteria nearly met:** A1c; A2c.

DIARD'S TROGON *Harpactes diardii* occurs in the Sundaic lowlands, from peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct), Kalimantan and Sumatra, **Indonesia** and **Brunei** in evergreen forest below 1,300 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000, del Hoyo *et al*. 2001). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysian Borneo (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species appears to be able to persist in hill-slope and logged forest, and second growth, provided the canopy is not heavily disturbed (Pearson 1975). **Criteria nearly met:** A1c; A2c.

WHITEHEAD'S TROGON *Harpactes whiteheadi* is restricted to the Bornean Mountains Endemic Bird Area, in hill and lower montane forest at 900–1,500 m, with populations on Mt Kinabalu, Sabah, and Mt Dulit, Sarawak, **Malaysia**, and Gunung Lunjut, Kalimantan, **Indonesia**; it is apparently uncommon everywhere and reputedly declining on Kinabalu, for reasons unknown (Stattersfield *et al*. 1998; also van Balen 1997, del Hoyo *et al*. 2001), and habitat loss in its lower elevation range may be a factor (see Threats under Mountain Serpent-eagle *Spilornis kinabaluensis*). **Criterion nearly met:** C1.

CINNAMON-RUMPED TROGON *Harpactes orrhophaeus* is known from peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan and Sumatra, **Indonesia** and **Brunei**, where it occurs in the lower storey of humid evergreen forest to 1,500 m (Smythies 1981, Mann 1987, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000, del Hoyo *et al*. 2001). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*). In southern Thailand and Peninsular Malaysia its dependence on plains-level, closed-canopy forest places it at more serious risk than its congeners (Wells 1999), but in Borneo it occurs on relatively less threatened forests on slopes (e.g. Banks 1937a). **Criteria nearly met:** A1c; A2c.

SCARLET-RUMPED TROGON *Harpactes duvaucelii* occurs in the Sundaic lowlands, from southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan and Sumatra, **Indonesia** and **Brunei**, where it is generally abundant in lowland primary and logged forests (including swamp forest, although there are no records from peatswamp) to 1,065 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000, del Hoyo *et al*. 2001). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary

formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia, such that the species is likely to be declining sufficiently rapidly to approach the threshold for Vulnerable. Although intolerant of heavy disturbance to canopy cover, the species's ability to use regenerated forest and its wide occurrence on slopes suggests that it is not immediately threatened (Wells 1999). **Criteria nearly met:** A1c; A2c.

WARD'S TROGON *Harpactes wardi* is known from the eastern Himalayas in **Bhutan** (uncommon and local, although recorded regularly in recent years: Inskipp *et al.* 1999a), **India** (small numbers seen recently in Arunachal Pradesh, where it is apparently local and rare: Singh 1994, Grimmett *et al.* 1998), **Myanmar** (formerly locally common in the north, but no recent records and thought to be generally uncommon: Smythies 1986, Collar *et al.* 1994, Robson 2000), mainland **China** (three collected in north-west Yunnan, 1973: Pang Yanzhang *et al.* 1980) and **Vietnam** (previously common on Fan Si Pan, north-west Tonkin, but no recent records despite intensive searching: Delacour 1930, A. W. Tordoff verbally 2000). It is found in the lower storey, undergrowth and bamboo of tall broadleaf evergreen forest between 1,500 and 3,200 m (Stanford and Mayr 1940–1941, Grimmett *et al.* 1998), perhaps moving downslope during the cold season to c.1,220 m in some areas (Robson 2000). It is threatened by forest clearance and degradation in much of its range, particularly through logging and shifting cultivation (Collins *et al.* 1991, Collar *et al.* 1994, Stattersfield *et al.* 1998). It is also possibly susceptible to hunting (J. C. Eames *in litt*. 1997). **Criteria nearly met:** C1; C2a.

BLYTH'S KINGFISHER *Alcedo hercules* ranges from eastern **Nepal** (vagant: Inskipp and Inskipp 1991), **Bhutan** (rare: Clements 1992, Inskipp *et al.* 1999a), north-east **India** (rare: Grimmett *et al.* 1998), **Bangladesh** (vagrant: Thompson *et al.* 1993, Grimmett *et al.* 1998), **Myanmar** (scarce to fairly common in the north, west and south: Smythies 1986, Robson 2000), mainland **China** (southern Yunnan, e.g. Mengyang Nature Reserve, and Hainan island: Collar *et al.* 1994), north-west **Thailand** (very rare visitor: Lekagul and Round 1991), **Laos** (uncommon to locally common in the north and Annamite mountains, scarce further south: Tobias 1997, Thewlis *et al.* 1998, Duckworth *et al.* 1999) and **Vietnam** (locally fairly common, at least historically, on Mount Fan Si Pan, west Tonkin and Annam: Collar *et al.* 1994, Robson 2000). It is found along streams in evergreen forest from 200–1,200 m (Ali and Ripley 1968–1998, Duckworth *et al.* 1999), mainly at 400–1,000 m (Fry *et al*. 1992, Grimmett *et al.* 1998). It is thus still widespread at low densities within its historical range, although deforestation is reducing and fragmenting its habitat (see Collins *et al.* 1991) and human disturbance and river pollution are possibly also threats (Eames *et al.* 1992). Given its linear distribution along rivers, and thus restricted extent of occurrence, the total population size is potentially modest. **Criteria nearly met:** C1; C2a.

SULAWESI KINGFISHER *Ceyx fallax* is restricted to the Sulawesi Endemic Bird Area (race *fallax*) and Sangihe and Talaud Endemic Bird Area (race *sangirensis*), **Indonesia**, where it inhabits drier primary lowland forest (unassociated with water) up to 1,000 m, chiefly 600 m (Stattersfield *et al*. 1998; also Stresemann and Heinrich 1939–1941). It is generally uncommon and possibly extinct on Sangihe owing to habitat loss (Coates and Bishop 1997). Forest destruction within its elevation range has been extensive in recent decades (see Threats under Lombopatang Flycatcher *Ficedula bonthaina*), and its populations must have suffered a commensurate decline. **Criteria nearly met:** C1; C2a.

LILAC-CHEEKED KINGFISHER *Cittura cyanotis* is restricted to the Sulawesi Endemic Bird Area (race *cyanotis*) and Sangihe and Talaud Endemic Bird Area (race *sanghirensis*), **Indonesia**, where it inhabits primary and tall secondary lowland forest up to 1,000 m (Stattersfield *et al*. 1998). Its distribution is apparently much fragmented, with no records

from South Sulawesi, and it is generally uncommon (White and Bruce 1986, Coates and Bishop 1997). Forest destruction within its elevation range has been extensive in recent decades (see Threats under Lombopatang Flycatcher *Ficedula bonthaina*), and its populations must have suffered a commensurate decline. **Criteria nearly met:** A1c; A2c; C1.

BROWN-WINGED KINGFISHER *Pelargopsis amauropterus* occurs in **India** (locally common in West Bengal, rare in Orissa and Assam; largely resident but apparently also non-breeding visitor: Grimmett *et al.* 1998), **Bangladesh** (locally common: Grimmett *et al.* 1998), **Myanmar** (fairly common to locally common resident in the south-west and Tenasserim: Robson 2000), Peninsular **Thailand** (uncommon to locally fairly common on the west coast: Lekagul and Round 1991, Robson 2000) and Peninsular **Malaysia** (fairly common but restricted to islands of the north-west: Wells 1999, Robson 2000). It is usually restricted to coasts, favouring mangroves (particularly old growth), creeks and tidal rivers, although it has been recorded occasionally far inland (Grimmett *et al.* 1998, Robson 2000). Despite being locally common, its total population may not be very large within its linear distribution and it is presumably negatively affected by the ongoing clearance and degradation of mangroves in South-East Asia. **Criteria nearly met:** A1a,c,d; A2a,c,d; C1.

LAZULI KINGFISHER *Todiramphus lazuli* is restricted to the Seram Endemic Bird Area, **Indonesia**, where it inhabits lowland forest and secondary growth in farmland (Stattersfield *et al*. 1998). Despite fears that it might be suffering from habitat loss (Collar *et al*. 1994), and although its range remains very restricted, it appears to thrive in heavily degraded coastal habitats (MKP, JAT), with an overall density in logged and unlogged forest of four birds per km^2 (Marsden 1998). **Criterion nearly met:** C2b.

TALAUD KINGFISHER *Todiramphus enigma* is restricted to the Sangihe and Talaud Endemic Bird Area, **Indonesia**, where it is found only on the Talaud islands of Karakelang, Salibabu and Kabaruan, occupying (*contra* "all habitats" in Stattersfield *et al.* 1998) the mid-canopy of chiefly undisturbed but also secondary forest, often along streams (being replaced in cultivated and coastal areas by Collared Kingfisher *T. chloris*), but is common and currently little affected by habitat loss, a key area being the Karakelang Hunting Reserve (Riley 1997b, Wardill *et al*. 1997, Riley *et al*. 1998a; also White and Bruce 1986). This reserve and some other areas of forest on the island are, however, under some logging pressure, and the total area in question is not great (see Threats under Red-and-Blue Lory *Eos histrio*). **Criterion nearly met:** C2b.

CINNAMON-BANDED KINGFISHER *Todiramphus australasia* is restricted to four Endemic Bird Areas (Northern Nusa Tenggara; Sumba; Timor and Wetar; and the Banda Sea Islands, the first three with nominate *australasia*, the last one with races *dammeriana* and *odites*), **Indonesia** (possibly also East Timor), where it is found in monsoon forest at 0–700 m (Stattersfield *et al*. 1998; also White and Bruce 1986). Its distribution within this fairly wide area is, however, very patchy, and it is generally uncommon (Coates and Bishop 1997), perhaps because it is a closed-canopy specialist (P. Andrew, R. Noske *in litt*. 2000), so that habitat loss and degradation (e.g. for Timor see Threats under Wetar Ground-dove *Gallicolumba hoedtii*) seem likely to be considerable negative factors. **Criteria nearly met:** C1; C2a.

RUFOUS-COLLARED KINGFISHER *Actenoides concretus* is confined to the Sundaic lowlands, from southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct), Kalimantan, Sumatra (including offshore islands) and Java, **Indonesia** and **Brunei**, in lowland and hill forest, to 1,500 m (Smythies 1981, Mann 1987, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). Forest destruction in the

Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but its presence in hill forest suggests no immediate pressure from outright loss of habitat, although it is unlikely to survive logging events that heavily disrupt the canopy (Wells 1999). **Criteria nearly met:** A1c; A2c.

GREEN-BACKED KINGFISHER *Actenoides monachus* is restricted to the Sulawesi Endemic Bird Area (race *monachus* in North and Central, race *capucinus* in East, South-East and South Sulawesi), **Indonesia**, where it is generally uncommon (but highly inconspicuous) in primary and tall secondary lowland forest up to 900 m (Stattersfield *et al.* 1998; also Stresemann and Heinrich 1939–1941, White and Bruce 1986, Coates and Bishop 1997). Forest destruction within its elevation range has been extensive in recent decades (see Threats under Lombopatang Flycatcher *Ficedula bonthaina*), and its populations must have suffered a commensurate decline. **Criteria nearly met:** A1c; A2c.

BIAK PARADISE-KINGFISHER *Tanysiptera riedelii* is endemic to the twin islands of Biak-Supiori off Papua (formerly Irian Jaya), **Indonesia** (Beehler *et al.* 1986). It is sometimes considered to be a subspecies of Common Paradise-kingfisher *T. galeata* but there are significant morphological differences (Forshaw 1985, Sibley and Monroe 1990, 1993). It is reported to be fairly common in primary, secondary and logged forest, up to at least 300 m, and possibly 600 m (Gibbs 1993, Poulsen and Frolander 1994, Eastwood 1996, B. M. Beehler and SvB *in litt.* 2000, M. van Beirs *in litt.* 2000). However, it has also been reported as rare and restricted to tall forest (K. D. Bishop *in litt.* 2000). On Biak, forest is under heavy threat from logging and subsistence farming, but there appear to be large areas of forest remaining in interior (Supiori Bishop 1982, K. D. Bishop *in litt.* 1996, D. A. Holmes *in litt.* 2000). Its reported abundance, tolerance of degraded forest, and presence in the 110 km^2 Biak-Utara protected area (B. M. Beehler and SvB *in litt.* 2000) suggest that it is not threatened. However, there are contradictory reports of its status, and further research may show it to be threatened. **Criteria nearly met:** C1; C2a.

NUMFOR PARADISE-KINGFISHER *Tanysiptera carolinae* is endemic to the island of Numfor (330 km^2) in Geelvink Bay, Papua (formerly Irian Jaya), **Indonesia**, where it is common and widespread in all lowland habitats, including beach vegetation and highly degraded forest (Beehler *et al.* 1986, Diamond 1986). Much of the forest on Numfor has already been destroyed or degraded by logging and subsistence farming (Bishop 1982, K. D. Bishop *in litt.* 1996). No protected area currently exists on Numfor, but a reserve has been proposed (Diamond 1986). Although this species appears to be secure in highly degraded forest, the ecological requirements of this species and the current status of forest on Numfor are poorly known and leave cause for concern. **Criterion nearly met:** C2b.

MALABAR PIED-HORNBILL *Anthracoceros coronatus* is restricted to central and southern **India** (common in a few areas, but declining and confined to land under 300 m: Grimmett *et al.* 1998) and **Sri Lanka** (local and moderately plentiful, but now restricted to more secluded forest of the dry lowlands: Grimmett *et al.* 1998). It is subject to seasonal movements in open moist broadleaved deciduous and evergreen forests, visiting fruit trees in cultivated areas (Grimmett *et al.* 1998). Forest on Sri Lanka has suffered rapid degradation and fragmentation in the past decades through excessive gathering of fuelwood, clearance for permanent agriculture, shifting cultivation, fire, urbanisation and logging (Stattersfield *et al.* 1998). Closed-canopy forest is estimated to have declined from 29,000 km^2 (44% of the island's area) in 1956 to 12,260 km^2 in 1983 (Collins *et al.* 1991). Similar losses are occurring in

mainland India and the status of this species therefore requires monitoring. **Criterion nearly met:** C1.

BLACK HORNBILL *Anthracoceros malayanus* is confined to the Sundaic lowlands of peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct: Lim Kim Seng *in litt*. 2001), Kalimantan and Sumatra (including the Lingga archipelago plus Bangka and Belitung islands), **Indonesia** and **Brunei** (common), in lowland primary and logged forests and swamp forests, usually below 500 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*). It is more tolerant of secondary and regenerating areas than other hornbills, but its dependence on plains-level habitats has removed it from much of its former range, especially in Thailand where it is on the verge of extinction (Wells 1999). **Criteria nearly met:** A1c; A2c.

RHINOCEROS HORNBILL *Buceros rhinoceros* is confined to the Sundaic lowlands of extreme southern peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct), Kalimantan, Sumatra and Java, **Indonesia** and **Brunei** (uncommon), in lowland and hill forests and swamp forests, to 1,400 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's presence in hill forest suggests no immediate pressure from outright loss of habitat, although in Borneo it is shot for food and hat feathers by local tribes (Smythies 1981). It returns to customary nest holes, even after surrounding forest has been disturbed; but studies demonstrate that logging has a long-term negative impact on overall numbers (Johns 1987), perhaps by reducing nest sites. **Criteria nearly met:** A1c; A2c.

GREAT HORNBILL *Buceros bicornis* occurs in mainland **China** (rare resident in west and south-west Yunnan and south-east Tibet), **India** (locally fairly common, but declining), **Nepal** (local and uncommon, largely in protected areas), **Bhutan** (fairly common), **Bangladesh** (vagrant), **Myanmar** (scarce to locally common resident throughout), **Thailand** (widespread, generally scarce but locally common), **Laos** (formerly common; currently widespread but scarce, a major decline having clearly occurred), **Vietnam** (rare and declining resident), **Cambodia** (rare), Peninsular **Malaysia** (uncommon to more or less common) and **Indonesia** (uncommon on Sumatra) (Engelbach 1932, David-Beaulieu 1944, Lekagul and Round 1991, Grimmett *et al.* 1998, Thewlis *et al.* 1998, Duckworth *et al.* 1999, Wells 1999, MacKinnon and Phillipps 2000, Robson 2000). It frequents evergreen and mixed deciduous forests, ranging out into open deciduous areas to visit fruit trees, and ascending slopes to at least 1,560 m (Robson 2000). The abundance of this species tends to be correlated with the density of large trees (Datta 1998), and it is therefore commonest in unlogged forest and threatened by logging. It is particularly susceptible to hunting pressure as it is large, visits predictable feeding sites (such as fruiting trees) and casques are kept or sold as trophies (Duckworth *et al.* 1999). Hunting pressure has caused its extirpation from many areas (Round 1984, 1985, MacKinnon and Phillipps 2000). **Criteria nearly met:** A1a,c,d; A2a,c,d; C1.

RUFOUS HORNBILL *Buceros hydrocorax* is endemic to the **Philippines**, where it occurs in primary, mature secondary and disturbed forests on 11 islands, Luzon and Marinduque (race *hydrocorax*), Samar, Leyte, Bohol, Panaon, Biliran, Calico-an and Buad (race

semigaleatus), Dinagat, Siargao, Mindanao (plus Balut, Bucas and Talicud) and Basilan (race *mindanensis*), where it is still patchily common—notably in the Sierra Madre of Luzon—but continues to suffer from substantial hunting pressure and widespread loss of habitat (Dickinson *et al.* 1991, Gonzalez 1995, Kemp 1995; also R. J. Timmins *in litt.* 1994). **Criteria nearly met:** A1c; A2c.

HELMETED HORNBILL *Rhinoplax vigil* is confined to the Sundaic lowlands, where it is known from southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct: Lim Kim Seng *in litt.* 2001), Kalimantan and Sumatra, **Indonesia** and **Brunei**, in semi-evergreen and evergreen lowland forest to 1,500 m (Smythies 1981, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*). However, the species's ability to persist at higher elevations, where forest destruction has been less severe, means that its decline has been less rapid. It is also threatened by hunting, particularly for its casque (del Hoyo *et al.* 2001), and its population should therefore be monitored carefully. **Criteria nearly met:** A1c,d; A2c,d.

BROWN HORNBILL *Anorrhinus tickelli* occurs in mainland **China** (rare; recorded in southern Xishuangbanna, south Yunnan and south-east Tibet), **India** (a small population is resident in hills bordering the Brahmaputra valley), **Myanmar** (uncommon to locally common in the west and Tenasserim), **Thailand** (generally uncommon in the west, north-west and north-east), **Laos** (historically numerous, currently widespread and locally common but declining), **Vietnam** (rare to uncommon in Tonkin and Annam) and **Cambodia** (scarce), where it inhabits evergreen broadleaved forest in foothills up to 1,000 m in India and 1,500 m in South-East Asia and 1,800 m in China. (Lekagul and Round 1991, Grimmett *et al.* 1998, Thewlis *et al.* 1998, Duckworth *et al.* 1999, MacKinnon and Phillipps 2000, Robson 2000). It is threatened by forest loss through intensive shifting agriculture and widespread logging activities, and because of high levels of hunting in many portions of its range (Round 1984, 1985, Lambert *et al.* 1994, Thewlis *et al.* 1998). However, its population probably exceeds 10,000 individuals, and it therefore fails to meet the thresholds for Vulnerable. **Criterion nearly met:** C1.

WHITE-CROWNED HORNBILL *Aceros comatus* is confined to the Sundaic lowlands of southern Tenasserim, **Myanmar**, peninsular and south-west **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan and Sumatra, **Indonesia** and **Brunei** (apparently rare), where it is generally uncommon in semi-evergreen and evergreen forests to 1,675 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*). However, the species's ability to persist at higher elevations, where forest destruction has been less severe, means that its decline has been less rapid. **Criteria nearly met:** A1c; A2c.

SUNDA WRINKLED HORNBILL *Aceros corrugatus* is confined to the Sundaic lowlands of peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan and Sumatra (including the Batu islands), **Indonesia** and **Brunei** (widespread), in lowland, evergreen and swamp forests, to 1,000 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by

2010, and the situation is little different in Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*). However, the species's ability to persist at higher elevations, where forest destruction has been less severe, means that its decline has been less rapid, although it appears to be on the verge of extinction in Thailand (Lekagul and Round 1991). **Criteria nearly met:** A1c; A2c.

MINDANAO WRINKLED HORNBILL *Aceros leucocephalus* is endemic to the **Philippines** on three islands, Mindanao (Dickinson *et al*. 1991) and its two small satellites Camiguin Sur (BRT; specimens in DMNH) and Dinagat (duPont and Rabor 1973b). Most distributional data are from between 300 and 1,000 m (BirdLife database) in original native lowland forest (Rand and Rabor 1960, duPont and Rabor 1973b), but despite clearly losing ground to habitat clearance, hunting and trapping for trade, in recent years the species has been judged to remain fairly common (Gonzales and Rees 1988, Dickinson *et al.* 1991, Brooks *et al*. 1992, Evans *et al*. 1993a, Kemp 1995). **Criteria nearly met:** A1c; A2c; C1.

RED-CROWNED BARBET *Megalaima rafflesii* is confined to the Sundaic lowlands of southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, **Singapore**, Kalimantan and Sumatra, **Indonesia** and **Brunei**, in lowland evergreen forest to 800 m (Smythies 1981, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's ability to persist in secondary forest and rubber plantations (Smythies 1981) provides some hope, although it is gone from virtually all its Thai range and populations are highly fragmented everywhere (Wells 1999). **Criteria nearly met:** A1c; A2c.

RED-THROATED BARBET *Megalaima mystacophanos* is confined to the Sundaic lowlands of southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sabah, Sarawak and Peninsular (including the Tana islands) **Malaysia**, Singapore (extinct: Lim Kim Seng *in litt*. 2001), Kalimantan and Sumatra, **Indonesia** and **Brunei**, where it is generally abundant in lowland primary and logged forest, and somewhat rarer in flooded and peatswamp forest, to 1,000 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's numerical strength in hill-slope forest implies no immediate threat from habitat loss (Wells 1999). **Criteria nearly met:** A1c; A2c.

BLACK-BANDED BARBET *Megalaima javensis* is endemic to Java and Bali, **Indonesia**, where it inhabits open evergreen and hill forest, mostly in the lowlands, but locally to 1,500 m (MacKinnon and Phillipps 1993). Forest destruction in the lowlands of Java and Bali has been extremely extensive (see Threats under Javan Hawk-eagle *Spizaetus bartelsi*), but this bird's ability to persist at higher elevations suggests no immediate substantial threat, but it is also sometimes trapped and traded as a cagebird (JAT). **Criteria nearly met:** A1c,d; A2c,d.

YELLOW-CROWNED BARBET *Megalaima henricii* is confined to the Sundaic lowlands, from peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct: Lim Kim Seng *in litt*. 2001), Kalimantan and Sumatra, **Indonesia** and **Brunei** (common) in lowland primary (occasionally secondary) broadleaved evergreen (including swamp) forests to 975 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia

has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's occurrence in hill-slope forest implies no immediate threat from habitat loss, although loss of fig trees may result in reduced numbers (Wells 1999). **Criteria nearly met:** A1c; A2c.

MALAYSIAN HONEYGUIDE *Indicator archipelagicus* is very local and sparse to uncommon (although easily overlooked) in peninsular and west **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan and Sumatra, **Indonesia** and **Brunei**, in broadleaved, lowland evergreen forest to 1,000 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species appears to occur also in open country and secondary forest, in Borneo at least (Smythies 1981), and in more secure hill-slope forest in the Thai-Malay Peninsula (Wells 1999). **Criteria nearly met:** A1c; A2c.

YELLOW-RUMPED HONEYGUIDE *Indicator xanthonotus* occurs in **Pakistan** (possibly a seasonal or nomadic visitor, but there are no recent records), **Nepal** (local and uncommon), **India** (very rare and local), **Bhutan** (local and uncommon), mainland **China** (rare in south-east Tibet) and **Myanmar** (rare resident in the north), and occurs throughout in broadleaved or coniferous forest, usually between 1,450 and 3,500 m, males defending giant rock bee *Apis dorsata* nests that are usually attached to vertical cliffs (Cronin and Sherman 1976, Roberts 1991–1992, Grimmett *et al.* 1998, MacKinnon and Phillipps 2000, Robson 2000). It is generally assumed to be a rare species, but it is easily overlooked due to its inconspicuous behaviour, and it may be commoner than the paucity of records suggests; nevertheless, it has been suggested that over-exploitation of bees' nests for honey by human populations in the Himalayas might have a deleterious impact on its populations (Underwood 1992). **Criterion nearly met:** C1.

ANDAMAN WOODPECKER *Dryocopus hodgei* is endemic to the Andaman islands, **India**, where it is a common resident in large trees of evergreen forest (Davidar *et al.* 1996, Grimmett *et al.* 1998). Although forest remains fairly extensive on the Andamans, the human population on larger islands is rising rapidly and habitat is consequently under severe pressure from agriculture, grazing and logging (Whitaker 1985, Curson, 1989, Pande *et al.* 1991, Sinha 1992, Stattersfield *et al.* 1998). **Criteria nearly met:** B1+2a,b,c,d,e.

RED-COLLARED WOODPECKER *Picus rabieri* is known from mainland **China** (one historical record from extreme south Yunnan), **Laos** (widespread and locally fairly common) and **Vietnam** (uncommon in west and east Tonkin and north and central Annam), being found in evergreen and semi-evergreen forest and locally in tall mixed deciduous forest, usually below 700 m, ascending locally to c.1,000 m (Cheng Tso-hsin 1987, Collar *et al.* 1994, Tobias 1997, Thewlis *et al.* 1998, Duckworth *et al.* 1999, Robson 2000). The species has recently been recorded in **Cambodia** (C. M. Poole *in litt*. 2001). It tolerates, and in some regions even appears to prefer, logged and disturbed forest, provided that some large trees remain (Collar *et al.* 1994, Tobias 1997, Duckworth *et al.* 1999). It is nevertheless susceptible to the ongoing destruction and degradation of lowland forest that is taking place within its range (Collar *et al.* 1994). **Criteria nearly met:** A1b,c; A2b,c.

OLIVE-BACKED WOODPECKER *Dinopium rafflesii* is found from southern Tenasserim, **Myanmar**, peninsular and west **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct), Kalimantan and Sumatra, **Indonesia** and **Brunei** (uncommon), in mangroves, lowland

and hill forests to 1,200 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's abundance in hill-slope forest is unlikely to be at serious risk yet. **Criteria nearly met:** A1c; A2c.

BUFF-NECKED WOODPECKER *Meiglyptes tukki* is confined to the Sundaic lowlands, from southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct), Kalimantan (including the North Natuna islands) and Sumatra (including most offshore islands), **Indonesia** and **Brunei**, where it is generally fairly common in evergreen and semi-evergreen lowland and peatswamp forests (including secondary formations) to 1,250 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Wells 1999, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia, but the species's abundance in hill-slope forest means that it is unlikely to be at serious risk yet (Wells 1999). **Criteria nearly met:** A1c; A2c.

BLACK-AND-YELLOW BROADBILL *Eurylaimus ochromalus* is confined to the Sundaic lowlands, where it is recorded from southern Tenasserim, **Myanmar**, peninsular and west **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct: Lim Kim Seng *in litt*. 2001), Kalimantan (including the Natuna islands) and Sumatra (including offshore islands), Java and Bali, **Indonesia** and **Brunei**, where it is locally common in primary and secondary forest, and adjacent plantations, to 1,200 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Duckworth *et al*. 1999, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's use of secondary forest and submontane slopes implies that it is not immediately threatened (Medway and Wells 1976, MacKinnon and Phillipps 1993). **Criteria nearly met:** A1c; A2c.

GREEN BROADBILL *Calyptomena viridis* occurs in the Sundaic lowlands, from southern Tenasserim, **Myanmar**, peninsular and west **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct: Lim Kim Seng *in litt*. 2001), Kalimantan (including the Natuna islands) and Sumatra (including offshore islands), **Indonesia** and **Brunei**, in lowland and hill forests, principally below 800 m, but occasionally to 1,700 m (Smythies 1981, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Lambert 1996, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's use of secondary forest and submontane slopes implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

HOSE'S BROADBILL *Calyptomena hosii* is patchily distributed in Sabah and Sarawak, **Malaysia** and Kalimantan, **Indonesia**, where it occurs to 1,000 m (Smythies 1981, MacKinnon and Phillipps 1993, Lambert 1996). It is apparently quite common in the Baram drainage, very common on Gunung Liang Kubung and recorded in the upper Telen, but absent or rare on intervening mountains (MacKinnon and Phillipps 1993). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's use of submontane slopes, where the

destruction has been considerably less severe, implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

GIANT PITTA *Pitta caerulea* occurs in the Sundaic lowlands, from southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan and Sumatra (not recorded this century), **Indonesia** and **Brunei**, in forests and bamboo to 1,200 m (Smythies 1981, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Lambert 1996, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's use of submontane slopes, where the destruction has been considerably less severe, and its tolerance of second growth implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c; C1.

SULA PITTA *Pitta dohertyi* is endemic to the Banggai and Sula Islands Endemic Bird Area (on Peleng, Banggai, Taliabu and Mangole), **Indonesia**, where it is local and generally uncommon in lowland evergreen forest (seriously degraded selectively logged forest), apparently not ascending above 200 m (Davidson *et al.* 1993, 1995, Lambert 1996). Clearance, disturbance and degradation of lowland forests is increasing in the small range of this species (Stattersfield *et al.* 1998), and its population is therefore likely to be declining, although at a slow rate given its tolerance of secondary habitats. ***Taxonomy*** This form was treated as a subspecies by Stattersfield *et al.* (1998), but it is sufficiently distinct to warrant full species status. **Criterion nearly met:** C1.

GARNET PITTA *Pitta granatina* is confined to the Sundaic lowlands, where it is known from southern Tenasserim, **Myanmar** (no records this century), peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct), Kalimantan and north Sumatra, **Indonesia** and **Brunei**, where it is locally common in lowland evergreen forest to 600 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Lambert 1996, Erritzoe and Erritzoe 1998, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia, but the species's use of heavily logged forest and plantations implies that it is not immediately threatened (Erritzoe and Erritzoe 1998). **Criteria nearly met:** A1c; A2c.

MANGROVE PITTA *Pitta megarhyncha* occurs in **Bangladesh** (probably a very local resident in the Sundarbans, common at Khulna: Grimmett *et al.* 1998), **Myanmar** (scarce to locally common: Robson 2000), Peninsular **Thailand** (uncommon to locally common in west: Lekagul and Round 1991), Peninsular **Malaysia** (locally common: Wells 1999, Robson 2000; also one record from East Malaysia: MacKinnon and Phillipps 1993), **Singapore** and **Indonesia** (south from Riau archipelago, through eastern lowlands of Sumatra, to Bangka: MacKinnon and Phillipps 1993). It is restricted to coastal mangroves (Robson 2000), a habitat that is suffering severe pressure through clearance for fuel-wood, charcoal production and construction materials (J. W. K. Parr *in litt.* 1999). **Criteria nearly met:** A1b,c; A2b,c; C1.

NILGIRI PIPIT *Anthus nilghiriensis* is endemic to the Western Ghats of Kerala and Tamil Nadu, southern **India**, where it is locally fairly common on grassy upland slopes interspersed with bushes and trees, mainly above 1,500 m, but sometimes descending to c.1,000 m (Grimmett *et al.* 1998). Its range is small and its grassland habitat is gradually being converted to plantations of tea, eucalyptus and wattle *Acacia dealbata* (Collins *et al.* 1991, Stattersfield *et al.* 1998). **Criteria nearly met:** B1+2a,b,c,d,e.

BURU CUCKOO-SHRIKE *Coracina fortis* is restricted to the Buru Endemic Bird Area, **Indonesia**, where it inhabits lowland, montane and monsoon forest at 0–1,500 m (Stattersfield *et al.* 1998). In 1921–1922 it was unobtrusive but not particularly rare (Siebers 1930), in late 1989 it was sufficiently rarely encountered for no density estimates to be formulated (Marsden *et al.* 1997), and in 1995–1996 a mere 14 birds were encountered at three localities (MKP), all reinforcing the notion in White and Bruce (1986) that "it is actually very local". Nevertheless it occurs in disturbed habitat (Jepson 1993, MKP) and its preference appears to be higher-altitude primary forest with an open understorey (Marsden *et al.* 1997), so that it is probably little affected by current habitat changes (for which see Threats under Blue-fronted Lorikeet *Charmosyna toxopei*). **Criterion nearly met:** C1.

PIED CUCKOO-SHRIKE *Coracina bicolor* is restricted to the Sulawesi Endemic Bird Area (including the Togian islands, Muna and Buton) and Sangihe and Talaud Endemic Bird Area, **Indonesia**, where it inhabits lowland forest with clearings, scrub and mangroves up to 900 m, although in many places much lower (Stattersfield *et al.* 1998; also Stresemann and Heinrich 1939–1941, Coates and Bishop 1997). It is uncommon on Buton (Caterall undated), and generally uncommon to rare on Sulawesi (apparently absent from large parts of Central and East), with only two records from Sangihe (Coates and Bishop 1997). Forest destruction within its elevation range has been extensive in recent decades (see Threats under Lombopatang Flycatcher *Ficedula bonthaina*), and its populations must have suffered a commensurate decline. **Criteria nearly met:** A1c; A2c.

KAI CICADABIRD *Coracina dispar* is restricted to the Banda Sea Islands Endemic Bird Area, **Indonesia**, where it inhabits forest, edge and secondary woodland on the islands of Romang, Damar, Tanimbar, Kai, Seram Laut/Watubela and Banda (Stattersfield *et al.* 1998; also Coates and Bishop 1997). It is apparently generally uncommon, but may be locally numerous and is easily overlooked (Coates and Bishop 1997). The condition of many of the islands from which it has been reported is wholly unknown, and on this basis it is extremely difficult to assess its status. **Criteria nearly met:** A1c; A2c; B1+2c.

BLACK-BIBBED CICADABIRD *Coracina mindanensis* is endemic to the **Philippines** in five subspecies: *lecroyae* on Luzon (considered very rare: T. H. Fisher verbally 1997), *elusa* on Mindoro ("not common" in 1954: Ripley and Rabor 1958), *ripleyi* on Samar (rare: Rand and Rabor 1960), Biliran, Leyte and Bohol (rarely recorded in Rajah Sikatuna National Park: Brooks *et al.* 1995c), nominate *mindanensis* on Mindanao (fairly common at the PICOP concession, Bislig: T. H. Fisher verbally 1997) and Basilan, and *everetti* on Jolo, Lapac, Tawitawi and Bongao (Dickinson *et al.* 1991). It is probably highly elusive, sitting singly and silently in the forest canopy (G. C. L. Dutson *in litt.* 1995), but on Mt Malindang in 1956 it "seemed to prefer forests of lower elevation, down to the lowland" (Rand and Rabor 1960), and indeed throughout its range the great majority of records are from well below 1,000 m (BirdLife database); this suggests that it must have suffered population declines as a result of the loss of lowland forest in the Philippines. **Criteria nearly met:** A1c; A2c.

MINDANAO CUCKOO-SHRIKE *Coracina mcgregori* inhabits montane-mossy forest at 1,000 to 1,900 m on Mindanao, **Philippines** (Gonzalez and Mallari 1993, Heaney *et al.* 1993), where it is present in at least eight general localities (not four, as reported in Collar *et al.* 1994), namely Mt Mayo (four specimens in USNM); Daggayan (Meyer de Schauensee and duPont 1962); "Claveria" in Misamis Oriental (male in DMNH); Mt Kitanglad (Salomonsen 1953, Ripley and Rabor 1961, Gonzalez and Mallari 1993, Heaney *et al.* 1993, Lambert 1993c); Lake Sebu at Sitio Siete (Evans *et al.* 1993a, B. Gee *in litt.* 1997); Mt Busa (12 specimens in CMNH); Mt Malindang (Mearns 1907, McGregor 1921b, Rand and Rabor 1960); and

Mt Lamut (R. J. Timmins *in litt.* 1994). It is described as common within the confines of its range ("common locally": Dickinson *et al.* 1991), and this is confirmed by the sheer number of specimens obtained in short periods of time at various other localities, e.g. 27 in 34 days on Mt Malindang, 1956 (material in FMNH), 11 in 19 days on Mt Busa, 1993 (material in CMNH). Much forest above 1,300 m at Mt Kitanglad is now seriously threatened by clearance (Lambert 1993c, hence Collar *et al.* 1994), but in general forest at this altitude is relatively secure. **Criterion nearly met:** B1+2c.

FIERY MINIVET *Pericrocotus igneus* is confined to the Sundaic lowlands, from southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct), Kalimantan and Sumatra, **Indonesia**, **Brunei** and Palawan, **Philippines**, in coastal mangroves and broadleaf forest to 1,220 m, but principally below 600 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's use of submontane slopes, where forest destruction has been considerably less severe, and second growth implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

SPOT-NECKED BULBUL *Pycnonotus tympanistrigus* is restricted to the Sumatra and Peninsular Malaysia Endemic Bird Area, occurring only in Sumatra, **Indonesia**, in hill and lower montane forest at 600–1,400 m (Stattersfield *et al*. 1998). In the early 1990s it was known by few records and appeared to be very local and at risk from deforestation in its lower elevation range (most populations are below 900 m) (Collar *et al*. 1994), but it is now known to be relatively common and widespread, albeit still declining with habitat loss (P. A. J. Morris *in litt*. 1998, D. Yong verbally 1999). **Criteria nearly met:** A1c; A2c.

BLACK-AND-WHITE BULBUL *Pycnonotus melanoleucus* is confined to the Sundaic lowlands, where is it known from peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct), Kalimantan and Sumatra (including Mentawai island), **Indonesia** and **Brunei** (widespread) in forest to at least 1,830 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's use of submontane slopes, where forest destruction has been considerably less severe, and scrubby forest (King *et al.* 1975, Smythies 1981) implies that it is not immediately threatened; however, its apparently nomadic behaviour (D. Yong verbally 1999) suggests dependence on a temporally patchy resource, as yet unidentified. **Criteria nearly met:** A1c; A2c.

SCALY-BREASTED BULBUL *Pycnonotus squamatus* is confined to the Sundaic lowlands, from southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan, Sumatra and Java, **Indonesia** and **Brunei** (scarce) in lowland and hill forest to 1,000 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's use of submontane slopes, where forest destruction has been considerably less severe, implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

GREY-BELLIED BULBUL *Pycnonotus cyaniventris* is confined to the Sundaic lowlands, from southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan and Sumatra (including Mentawai island), **Indonesia** and **Brunei**, in lowland and hill forest to 1,200 m (Medway and Wells 1976, Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's use of submontane slopes, where forest destruction has been considerably less severe, and use of open wooded country (King *et al.* 1975) and forest edge (Medway and Wells 1976) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

PUFF-BACKED BULBUL *Pycnonotus eutilotus* occurs in the Sundaic lowlands, from southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct), Kalimantan and Sumatra (including Bangka island), **Indonesia** and **Brunei**, in lowland evergreen forest to 400 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's use of second growth (MacKinnon and Phillipps 1993), forest edge (Robson 2000) and even scrub (King *et al.* 1975) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

YELLOW-EARED BULBUL *Pycnonotus penicillatus* is an endemic resident in the wet-zone highlands of **Sri Lanka**, where it is common in forest and nearby gardens, at middle and higher elevations (Grimmett *et al.* 1998). Forest on the island has suffered rapid degradation and fragmentation in the past decades through excessive gathering of fuelwood, clearance for permanent agriculture, shifting cultivation, fire, urbanisation and logging (Stattersfield *et al.* 1998). It is feared that habitat loss will continue in the hills and the status of this species therefore requires monitoring. **Criteria nearly met:** B1+2a,b,c,d,e.

FINSCH'S BULBUL *Alophoixus finschii* occurs in the Sundaic lowlands, from extreme southern peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan and Sumatra (one record), **Indonesia** and **Brunei**, in lowland and hill forest to 760 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's ability to persist in secondary forest (Lekagul and Round 1991) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

BUFF-VENTED BULBUL *Iole olivacea* is restricted to the Sundaic lowlands of southern Tenasserim, **Myanmar**, peninsular and west **Thailand**, **Singapore** (scarce), Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan (including the Natuna and Anamba islands) and Sumatra (including offshore islands), **Indonesia** and **Brunei**, where it is generally a common bird of lowland and hill forest to 825 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but its ability to survive in second growth and orchards (King *et al*. 1975) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

ZAMBOANGA BULBUL *Ixos rufigularis* is endemic to Mindanao, **Philippines**, where it is known from around Lake Lanao (Meyer de Schauensee and duPont 1962) and the Zamboanga Peninsula, plus Basilan including Malamaui (Dickinson *et al.* 1991). Although characterised as "common in forest and forest edge" (Dickinson *et al.* 1991), recent evidence—an almost fruitless search in the Zamboango City watershed—suggests that it is fairly uncommon and possibly suffering from habitat loss within its somewhat restricted range (G. C. L. Dutson *in litt.* 1994). **Criteria nearly met:** B1+2c; C1; C2a.

STREAKED BULBUL *Ixos malaccensis* occurs in the Sundaic lowlands, from southern Tenasserim, **Myanmar**, peninsular **Thailand**, **Singapore** (vagrant), Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan and Sumatra, **Indonesia** and **Brunei**, in lowland and hill forest to 915 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's ability to survive in secondary forest (Smythies 1981) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

LESSER GREEN LEAFBIRD *Chloropsis cyanopogon* is restricted to the Sundaic lowlands, from southern Tenasserim, **Myanmar**, peninsular **Thailand**, **Singapore** (scarce), Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan (including Banggi Island) and Sumatra, **Indonesia** and **Brunei**, where it is generally fairly common in forest to 700 m (Smythies 1981, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's ability to survive in second growth and forest edge (Robson 2000) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

BLUE-MASKED LEAFBIRD *Chloropsis venusta* is restricted to the Sumatra and Peninsular Malaysia Endemic Bird Area, occurring only in Sumatra, **Indonesia**, in hill and lower montane forest at 600–1,500 m (Stattersfield *et al*. 1998). It is judged to be probably not uncommon (van Marle and Voous 1988) but upslope forest loss is reducing the lower elevation ranges of several rare Sumatran endemics (see, e.g., Threats under Crestless Fireback *Lophura erythrophthalma* and Sumatran Ground-cuckoo *Carpococcyx viridis*), and is likely to be causing a decline in this species also. **Criteria nearly met:** A1c; A2c.

GREEN IORA *Aegithina viridissima* is restricted to the Sundaic lowlands, where it occurs from southern Tenasserim, **Myanmar**, peninsular **Thailand**, Singapore (extinct: Lim Kim Seng *in litt*. 2001), Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan (including the Natuna islands) and Sumatra (including offshore islands), **Indonesia** and **Brunei**, in mangroves and forests to 825 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's ability to survive in second growth (Robson 2000) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

MOUNTAIN SHRIKE *Lanius validirostris* is endemic to the **Philippines**, where it occurs in clearings within forest in the highlands (above 1,000 m) of Luzon (nominate *validirostris*, in

the Cordillera Central and Sierra Madre), Mindoro (race *tertius*) and Mindanao (race *hachisuka*, including on Mt Kitanglad, Civolig, Mt Malindang and Mt Apo); although its area of occupancy must be very small, its habitat is apparently relatively secure and it is present throughout this habitat (Dickinson *et al.* 1991, Danielsen *et al.* 1994, Poulsen 1995, Lefranc 1997, NADM), so that with fuller assembly of information this species may well be judged not to be in any danger. **Criteria nearly met:** B1+2c; C2a.

BORNEAN BRISTLEHEAD *Pityriasis gymnocephala* is restricted to Borneo—Sabah and Sarawak, **Malaysia**, Kalimantan, **Indonesia** and **Brunei** (scarce)—principally in peatswamp forest up to 1,000 m (Smythies 1981, Mann 1987, MacKinnon and Phillipps 1993). Forest destruction in the Sundaic lowlands of Indonesian Borneo has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysian Borneo (see Threats under Crestless Fireback *Lophura erythrophthalma*). In addition, peatswamp forest has been severely affected by recent ENSO- and arson-related conflagrations (see Threats under Hook-billed Bulbul *Setornis criniger*). Although the species has therefore almost certainly experienced a significant population decline in the past decade owing to habitat loss, it appears to find some security in relatively less threatened forests on slopes. **Criteria nearly met:** A1c; A2c.

JAPANESE WAXWING *Bombycilla japonica* breeds only in the far east of **Russia**, where it has been found nesting in eastern Yakutia, Khabarovsk and Amur, in coniferous forests among larch and cedar trees; it is locally common on the breeding grounds, but given its limited range its total population must be small (Dement'ev and Gladkov 1951–1954). It is a non-breeding visitor to **Japan**, where it is uncommon and sporadic (Brazil 1991), **South Korea**, where it is irregular and uncommon (Austin 1948, Gore and Won 1971), and mainland **China**, where it is uncommon in the north and rare in the south (Cheng Tso-hsin 1987) and **Taiwan** (China) (Wang Chia-hsiong *et al.* 1991). It has presumably been affected by the logging and development of its forest habitat (see Threats under Blakiston's Fish-owl *Ketupa blakistoni*). **Criterion nearly met:** C1.

GEOMALIA *Geomalia heinrichi* is restricted to the Sulawesi Endemic Bird Area, **Indonesia**, where it inhabits montane forest (particularly moss forest) at 1,700–3,400 m (Stattersfield *et al*. 1998; also Stresemann and Heinrich 1939–1941). It is unrecorded in the East Peninsula *fide* Stattersfield *et al*. (1998) and from South Sulawesi *fide* Coates and Bishop (1997). The comment that its legs and feet are "notably weak and small for a large terrestrial bird" (White and Bruce 1986) is mistaken (NJC), and it is in fact an agile and wary species of the forest floor (Stresemann and Heinrich 1939–1941); nevertheless, the number of feral domestic cats observed in Lore Lindu at night in 1999 was such as to raise serious concern over the security of the mid-sized and smaller ground-haunting birds endemic to Sulawesi (R. F. A. Grimmett verbally 1999), and it may be that declines have resulted from this additional predation pressure. **Criteria nearly met:** A1c,e; A2c,e.

SLATY-BACKED THRUSH *Zoothera schistacea* is restricted to the Banda Sea Islands Endemic Bird Area, **Indonesia**, where it inhabits forest on Tanimbar only (Stattersfield *et al*. 1998). It is known from Larat and Yamdena, and is common in lowland forest (Collar and Andrew 1988), but there is significant logging in the south of Yamdena (Bishop and Brickle 1998; see Threats under Lesser Masked-owl *Tyto sororcula*), so the species is likely to be in steady decline. **Criteria nearly met:** C1; B1+2c.

MOLUCCAN THRUSH *Zoothera dumasi* is restricted to the Buru Endemic Bird Area (race *dumasi*) and the Seram Endemic Bird Area (race *joiceyi*), **Indonesia**, where it inhabits montane

forest at 800–1,500 m (900–1,280 m on Seram) (Stattersfield *et al.* 1998). The Seram race appears to be very rare (Bowler and Taylor 1989), while on Buru it was "not very rare" in 1922 (Hartert 1924a) and less rare than it at first appears (Stresemann 1914b). Hill forest on both islands is somewhat more secure than low-lying formations (see Threats under Purple-naped Lory *Lorius domicellus* and Blue-fronted Lorikeet *Charmosyna toxopei*). **Taxonomy**: The two races *dumasi* and *joiceyi* are sufficiently morphologically distinct to be treated as two separate species (P. C. Rasmussen verbally 1998), which might qualify one or both for threatened status. **Criteria nearly met:** A1c,d; A2c,d.

RED-BACKED THRUSH *Zoothera erythronota* is restricted to the Sulawesi Endemic Bird Area (but not the East Peninsula) (race *erythronota*) and Banggai (race *mendeni*) and Sula Island (Peleng and Taliabu only, race undescribed) Endemic Bird Area, **Indonesia**, where it inhabits lowland forest (and bamboo in the Banggai/Sulas) up to 1,000 m (Stattersfield *et al.* 1998; also Stresemann and Heinrich 1939–1941, Coates and Bishop 1997). It is uncommon but easily overlooked on Buton (Catterall undated; also Viney 1995), and generally uncommon on Sulawesi (Coates and Bishop 1997). Forest destruction within its elevation range has been extensive in recent decades (see Threats under Lombopatang Flycatcher *Ficedula bonthaina*), and its populations must have suffered a commensurate decline; there is concern that trade may also be a factor (see under Chestnut-backed Thrush *Zoothera dohertyi*). **Criteria nearly met:** A1c,d,e; A2c,d,e.

CHESTNUT-BACKED THRUSH *Zoothera dohertyi* is restricted to three Endemic Bird Areas (Northern Nusa Tenggara; Sumba; Timor and Wetar), **Indonesia**, where it occurs at higher altitudes in semi-evergreen, lower montane and montane forest (Stattersfield *et al.* 1998). Although considered generally uncommon to rare, and only locally common at higher elevations (Coates and Bishop 1997), it has been traded in very high volume within Indonesia in recent years because of its voice, and is probably already extinct on Lombok and close to extinction on Sumbawa (R. F. A. Grimmett verbally 1999). There is a single village on Sumbawa whose inhabitants have in the past two years devoted themselves to the capture, to virtual extinction, of the Lesser Sundas population of the Chestnut-capped Thrush *Zoothera interpres*, with some birds reaching Singapore (*BirdLife International–Indonesia Programme Annual Report* 1999), and it is feared that the Chestnut-backed and other *Zoothera* thrushes may already have become the new victims of this relentless pressure (R. F. A. Grimmett verbally 1999). **Criteria nearly met:** A1c,d,e; A2c,d,e; C1.

ORANGE-BANDED THRUSH *Zoothera peronii* is restricted to the Timor and Wetar Endemic Bird Area and the Banda Sea Islands Endemic Bird Area, **Indonesia** and **East Timor**, where it occurs and is generally common on Roti and West Timor (race *peronii*), East Timor, Wetar and Romang, Babar and Damar (race *audacis*) at 0–1,200 m in forest including monsoon forest (Stattersfield *et al.* 1998; also Coates and Bishop 1997). Although it has been found in degraded patches (R. Noske *in litt.* 2000), it appears to favour areas with closed-canopy forest (P. Andrew *in litt.* 2000), which are constantly diminishing (for Timor see Threats under Wetar Ground-dove *Gallicolumba hoedtii*), and there is great concern that so beautiful a songster (see Coates and Bishop 1997) will be targeted by traders already skilled in trapping *Zoothera* thrushes out of Lesser Sunda forests (see above under Chestnut-backed Thrush *Z. dohertyi*). **Criteria nearly met:** A1c,d; A2c,d.

EVERETT'S THRUSH *Zoothera everetti* is restricted to the Bornean Mountains Endemic Bird Area, in montane forest at 1,200–2,300 m on Mts Kinabalu, Murud, Mulu, Dulit, Trus Madi and the Kelabit uplands in Sabah and Sarawak, **Malaysia** (Stattersfield *et al.* 1998; also Clement in press). Despite its highly secretive nature, it may be genuinely rare and

declining in some areas with loss of forest at lower altitudes (see Threats under Mountain Serpent-eagle *Spilornis kinabaluensis*). **Criteria nearly met:** C1; C2b.

SPOT-WINGED THRUSH *Zoothera spiloptera* is an endemic resident in southern **Sri Lanka** where it is fairly common to common in lowlands and hills of the wet zone from 300–1,220 m (Grimmett *et al.* 1998, Jones *et al.* 1998). It inhabits the leaf-litter of damp, dense, wooded areas and occasionally gardens near forest, also occurring sporadically (but not recently) in the dry zone (Henry 1955, Grimmett *et al.* 1998). It is apparently most abundant in primary habitat, but is also recorded in selectively logged forest, forest edges or near tea cultivation and scrub, although it may be dependent on relatively intact forest remaining nearby (Jones *et al.* 1998). Forest on the island has suffered rapid degradation and fragmentation in past decades through excessive gathering of fuelwood, clearance for permanent agriculture, shifting cultivation, fire, urbanisation and logging (Stattersfield *et al.* 1998). Closed-canopy forest is estimated to have declined from 29,000 km^2 (44% of the island's area) in 1956 to 12,260 km^2 in 1983 (Collins *et al.* 1991). It is feared that this loss will continue and the status of this species therefore requires monitoring. **Criteria nearly met:** B1+2a,b,c,d,e.

FAWN-BREASTED THRUSH *Zoothera machiki* is restricted to the Banda Sea Islands Endemic Bird Area, **Indonesia**, where it inhabits forest and scrub on Tanimbar (Yamdena, Larat) only (Stattersfield *et al.* 1998). It was long known from just three specimens (White and Bruce 1986) but in the past 15 years has been found locally common in primary forest, occasionally secondary scrub, and sometimes concentrating in recently burnt areas (Collar and Andrew 1988, Lewis 1993, Bishop and Brickle 1998). Logging in the south of Yamdena has continued apace since 1992 (Bishop and Brickle 1998; see Threats under Lesser Masked-owl *Tyto sororcula*), so the species is likely to be in steady decline. **Criteria nearly met:** A1c; A2c; B1+2c.

FIRETHROAT *Luscinia pectardens* breeds in Sichuan, Yunnan and south-east Tibet, mainland **China**, in dense scrub in valley bottoms, and in mountain forest, at 2,800–3,700 m (Ali and Ripley 1968–1998, Cheng Tso-hsin 1987), and it is a non-breeding visitor to the mountains of Arunachal Pradesh and Meghalaya, **India** and **Bangladesh** (Grimmett *et al.* 1998). It appears to be very rare, with the only recent records on the breeding grounds from Wolong Biosphere Reserve in Sichuan (P. Alström *in litt.* 1993), although it must be greatly under-recorded as much of its range is remote and seldom visited by ornithologists. It has presumably been affected by deforestation on both the breeding and wintering grounds (see Smil 1984, 1993, Collins *et al.* 1991, MacKinnon *et al.* 1996). **Criterion nearly met:** C1.

RUFOUS-TAILED SHAMA *Trichixos pyrropyga* occurs in the Sundaic lowlands of peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan and Sumatra, **Indonesia** and **Brunei**, where it inhabits lowland (including peatswamp) and hill-slope forests to 1,200 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*). Although the species has almost certainly experienced a significant population decline in the past decade owing to habitat loss, it appears to find some security in relatively less threatened forests on slopes. **Criteria nearly met:** A1c; A2c.

ALA SHAN REDSTART *Phoenicurus alaschanicus* is known from Qinghai, Gansu and Ningxia, mainland **China**, with presumed non-breeding records from Shaanxi, Shanxi, Hebei

and Beijing (Cheng Tso-hsin 1987). Its breeding range and habitat requirements remain poorly understood because of the sparse ornithological coverage of much of its range, but there have been recent breeding records on scrub-covered hillsides from c.3,300 m to the treeline (He Fenqi verbally 1993, P. Alström *in litt.* 1996). It appears to be rare (P. Alström *in litt.* 2000), and has presumably been affected by habitat loss in its breeding range (see Smil 1984, 1993, MacKinnon *et al.* 1996). It has been recorded in the Helan Mountain Nature Reserve, and in or near to Yanchiwan Nature Reserve (Stattersfield *et al.* 1998). **Criterion nearly met:** C1.

CHESTNUT-NAPED FORKTAIL *Enicurus ruficapillus* is patchily distributed through the Sundaic lowlands, from southern Tenasserim, **Myanmar**, peninsular and west **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan and Sumatra, **Indonesia** and **Brunei**, where it occurs along rivers and streams in lowland and hill forests to 1,300 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but this species's ability to survive in submontane forest implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

WHITE-BELLIED BUSHCHAT *Saxicola gutturalis* is restricted to the Timor and Wetar Endemic Bird Area, **Indonesia**, where it occurs on Timor (possibly also East Timor) and its satellite islands of Roti and Semau (nominate *gutturalis* on Timor and Roti, race *luctuosa* on Semau) at 0–1,200 m in monsoon forest and scrubby savanna (Stattersfield *et al.* 1998; also White and Bruce 1986). In West Timor it is present in even very small remnant pockets of woodland but is largely excluded from savanna and open scrub by the Pied Bushchat *S. caprata* (Collar and Andrew 1988, R. Noske *in litt.* 2000); forest loss on Timor has been extensive (see Threats under Wetar Ground-dove *Gallicolumba hoedtii*) and the total number of this species may not be great. **Criterion nearly met:** C1.

RAIL-BABBLER *Eupetes macrocerus* is restricted to the Sundaic lowlands, from peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan (including the Natuna islands) and Sumatra, **Indonesia** and **Brunei**, in lowland broadleaf forest up to 1,000 m (Smythies 1981, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's preference for submontane forests (Smythies 1981) and ability to persist in logged habitat (MacKinnon and Phillipps 1993) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

RUFOUS-FRONTED LAUGHINGTHRUSH *Garrulax rufifrons* is restricted to the Java and Bali Forests Endemic Bird Area, where it occurs in montane forest at 1,000–2,400 m in West and Central Java, **Indonesia** (Stattersfield *et al.* 1998). It is fairly heavily exploited as a cagebird, which has rendered it uncommon in otherwise moderately secure habitat (D. A. Holmes verbally 1998). **Criteria nearly met:** A1c,d; A2c,d.

BLACK-HOODED LAUGHINGTHRUSH *Garrulax milleti* is confined to the Da Lat and Di Linh plateaus, **Vietnam** (locally common: Robson 2000), and the Kon Tum plateau of Vietnam (fairly common around Ngoc Linh: Tordoff *et al.* 2000) and **Laos** (locally fairly common in Xe Kong and Attapu provinces: Davidson *et al.* 1997, Thewlis *et al.* 1998). It is

usually found in flocks in dense undergrowth of broadleaved evergreen forest between 800 and 1,650 m (Delacour and Jabouille 1931, Brunel 1978, Eames *et al.* 1992, Eames 1995a, Robson 2000), although it has been found in remnant patches of habitat in Laos (Davidson *et al.* 1997). As it occurs at relatively low altitudes, it is vulnerable to habitat destruction through agricultural encroachment, charcoal burning and fuelwood collection, particularly as the human population of the area is increasing because of government resettlement programmes (Collar *et al.* 1994, Stattersfield *et al.* 1998). However, the known range of the species is sufficiently large, and the loss of habitat within it sufficiently gradual (particularly in Laos: Davidson *et al.* 1997), and it occurs in a few protected areas, including Dong Ampham NBCA (Laos), Ngoc Linh, Kon Ka Kinh and Kon Cha Rang nature reserves (Vietnam). **Criteria nearly met:** B1+2a,b,c,d,e.

CHESTNUT-BACKED LAUGHINGTHRUSH *Garrulax nuchalis* is resident in the hills of north-east **India** (apparently rare in Assam, Manipur, Nagaland and Arunachal Pradesh: Grimmett *et al.* 1998) and **Myanmar** (uncommon in the north: Robson 2000). It inhabits tall grass or dense bushes in stony scrub-covered ravines and hills, from the lowlands to c.900 m (Grimmett *et al.* 1998, Robson 2000). There have been very few recent records and its current status is poorly known, although this is at least in part because much of its range is infrequently visited by ornithologists. Surveys are required to establish its current distribution and population size, the extent of its habitat and the threats it faces. **Criteria nearly met:** C1; C2a.

GREY-BREASTED LAUGHINGTHRUSH *Garrulax jerdoni* is endemic to the Western Ghats of southern Kerala, western Tamil Nadu and southern Karnataka, **India**, where it is locally common from 1,100 m to the summits (Gaston and Zacharias 1996, Grimmett *et al.* 1998). It frequents undergrowth and low bushes, raspberry and bracken thickets growing along banks of streams and edges of sholas, and tea and cardamom plantations; its distribution is apparently tied to that of the wild raspberry (Grimmett *et al.* 1998). While its range is small, its tolerance for disturbed habitats suggests that it is not immediately threatened by habitat modification. However, an increasing human population has led to increased illegal encroachment into Western Ghat forests, livestock-grazing and the harvesting of fuelwood and huge quantities of forest products such as bamboo and canes; furthermore, hydropower development and road-building are causing reductions in forest cover in some areas (Stattersfield *et al.* 1998). **Criteria nearly met:** B1+2a,b,c,d,e.

WHITE-CHESTED BABBLER *Trichastoma rostratum* occurs in the Sundaic lowlands, from southern Tenasserim, **Myanmar**, peninsular **Thailand**, **Singapore** (rare), Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan (including the northern islands) and Sumatra (including offshore islands), **Indonesia** and **Brunei** (widespread and fairly common), near water in mangroves and lowland evergreen forest to 500 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's ability to survive in second growth (Robson 2000) and overgrown plantations (Smythies 1981) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

SHORT-TAILED BABBLER *Malacocincla malaccensis* occurs in the Sundaic lowlands, from southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, **Singapore**, Kalimantan (including the Natunas and Anamba islands) and Sumatra (including offshore islands), **Indonesia** and **Brunei**, where it is generally very common in lowland evergreen forest, including swamp forest, to 1,000 m (Smythies 1981, Mann 1987,

Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's ability to survive in hill-slope forest and second growth (King *et al*. 1975, Robson 2000) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

SOOTY-CAPPED BABBLER *Malacopteron affine* occurs in the Sundaic lowlands, from peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan and Sumatra (including Banyak and Bangka), **Indonesia** and **Brunei**, where it is generally common in lowland forest (including freshwater and peatswamp forest) to 700 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's ability to survive in forest edge and secondary formations (King *et al*. 1975, MacKinnon and Phillipps 1993) implies that it is not unduly threatened. **Criteria nearly met:** A1c; A2c.

RUFOUS-CROWNED BABBLER *Malacopteron magnum* occurs in the Sundaic lowlands, from southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan (including Natuna Islands) and Sumatra (including Bangka), **Indonesia**, **Brunei** and Palawan, **Philippines**, where it is generally common in lowland evergreen forest to 800 m (Dickinson *et al*. 1991, Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's ability to survive in second growth (King *et al.* 1975) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

MELODIOUS BABBLER *Malacopteron palawanense* is endemic to Palawan and Balabac in the **Philippines** (Dickinson *et al.* 1991), with records on Palawan including from Tanabag (specimen in USNM); Iwahig Penal Colony (four males in DMNH, many observers); Puerto Princesa (specimens in BMNH, CM, DMNH, MNHN, USNM; also Baud 1978); Inagauan (Hartley and McGowan 1991); Quezon, including Tabon and Bungalon (four specimens in DMNH, FMNH); Taguso (three specimens in MNHN and RMNH); Pulot, Brooke's Point (male in DMNH); and Singnapan and adjacent areas (Sison 1983). It is described as uncommon (Dickinson *et al.* 1991) and may be an extreme lowland forest specialist (C. R. Robson verbally 1994), but in the pockets of remaining habitat to which it has become confined it remains at good densities (J. Hornbuckle *in litt.* 1994). Oddly it has not been recorded from St Paul's Subterranean River National Park, which suggests some ecological feature of significance which needs to be identified; currently the best protection in the face of continuing forest loss appears to be afforded by the Iwahig Penal Colony (many observers). **Criteria nearly met:** A1c; A2c; B1+2c; C1; C2a.

GREY-BREASTED BABBLER *Malacopteron albogulare* is restricted to Sabah, Sarawak, and Peninsular **Malaysia**, Singapore (extinct: Lim Kim Seng *in litt*. 2001), Kalimantan and Sumatra (including the Batu islands and Lingga archipelago), **Indonesia**, in lowland peatswamp and heath forest (Smythies 1981, Davies and Payne 1982, MacKinnon and Phillipps 1993). Swamp-forest destruction in Indonesian Borneo has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysian

Borneo (see Threats under Hook-billed Bulbul *Setornis criniger*), but the species's ability to survive in second growth and to some extent in other habitats (Dutson *et al*. 1991, Duckworth *et al*. 1996) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

SHORT-TAILED SCIMITAR-BABBLER *Jabouilleia danjoui* is known from east Tonkin, north, central and south Annam, **Vietnam**, where small numbers have been recorded at many sites, and central **Laos**, where a fairly large population survives near the Vietnamese border (Evans and Timmins 1994, Tobias 1997, Robichaud 1999). In south Annam, the subspecies *danjoui* is found in montane evergreen forest between 1,500 and 2,100 m, but the northern subspecies *parvirostris* is mainly found in lowland forest between 50 and 900 m (Robson *et al.* 1993b, Robson 2000), ascending locally to 1,650 m (Kalyakin and Korzun 1997, Robichaud 1999). The species frequents the lower storey, often foraging on the ground (Thewlis *et al.* 1998). While there is no direct evidence of a decline, it is threatened by deforestation throughout its range, particularly where it prefers forest at lower altitudes (Collar *et al.* 1994). It occurs in several protected areas throughout its range, including Nakai-Nam Theun NBCA in Laos (Robichaud 1999), Ke Go, Vu Quang and Pu Mat nature reserves in Vietnam (Robson *et al.* 1991, Kalyakin and Korzun 1997, Round 1999). **Criteria nearly met:** A1b,c; A2b,c; C1; C2a.

STRIPED WREN-BABBLER *Kenopia striata* is restricted to the Sundaic lowlands, from peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct), Kalimantan and Sumatra, **Indonesia** and **Brunei**, in forest (including swamp forest) to 1,000 m (King *et al.* 1975, Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's presence in hill-slope and logged forest implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

LARGE WREN-BABBLER *Napothera macrodactyla* occurs in the Sundaic lowlands, from peninsular **Thailand**, Peninsular **Malaysia**, Singapore (extinct), and Sumatra and Java (few records), **Indonesia** in evergreen forest to 1,200 m (Smythies 1981, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's use of submontane forest implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

BLACK-THROATED WREN-BABBLER *Napothera atrigularis* is restricted to Borneo in the Sundaic lowlands of Sabah and Sarawak, **Malaysia**, Kalimantan, **Indonesia** and **Brunei**, in broadleaf forest from lowlands to 1,500 m (Smythies 1981, Mann 1987, MacKinnon and Phillipps 1993). Forest destruction in the Sundaic lowlands of Indonesian Borneo has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysian Borneo (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's preference for submontane forests (MacKinnon and Phillipps 1993) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

LUZON WREN-BABBLER *Napothera rabori* is endemic to Luzon, **Philippines**, in three races, nominate *rabori* in the north at Tabbug (Ilocos Norte) (Rand 1960), Mt Cagua (male in FMNH), Baggao below Los Dos Cuernos (male in USNM) and four other sites—Maconacon, Minuma, Dimapnat and Palanan—in the Sierra Madre (Poulsen 1995),

mesoluzonica in the central portion at the Angat watershed (Lambert 1993c, Harrap and Mitchell 1994), Tibag (Gonzalez 1995), Saray (10 specimens in BMNH, CM, DMNH; also duPont 1971), Pangil (duPont 1971, Won 1986b; also three specimens in CM, DMNH), and (subspecies uncertain) Quezon National Park (de Roever 1990, Harrap and Mitchell 1994, Hornbuckle 1994), and *sorsogonensis* from the southern peninsula at Mt Isarog (Goodman and Gonzales 1990), Mt Bulusan (Rand and Rabor 1967) and Tugas (Dickinson *et al.* 1991). This ground-haunting bird, although described as rare and very local (Dickinson *et al.* 1991) and despite being a lowland specialist (highest elevation recorded is only 800 m: BirdLife database), appears in fact to be overlooked, and tape-recording has shown it to be commoner and widerspread than previously thought (P. A. J. Morris verbally 1996). **Criteria nearly met:** A1c; A2c.

RUFOUS-THROATED WREN-BABBLER *Spelaeornis caudatus* occurs in **Nepal** (very rare in the east: Grimmett *et al.* 1998), **Bhutan** (uncommon: Inskipp *et al.* 1999a) and **India** (where it is locally common in Sikkim, northern West Bengal and Arunachal Pradesh). It inhabits dense undergrowth of moist evergreen forest, often in steep gullies, especially where ferns, mossy rocks and fallen trees abound, from 1,500 m to 2,500 m, perhaps occasionally to 3,100 m (Ali and Ripley 1968–1998, Inskipp and Inskipp 1991, Grimmett *et al.* 1998). Within its small range, it is threatened by the destruction and fragmentation of forest, chiefly through logging and shifting cultivation (Inskipp 1989, Singh 1994, 1999, Stattersfield *et al.* 1998). However, as it is a highly skulking species and easily overlooked, further research will probably reveal it to be more widespread and abundant than current records imply. **Criteria nearly met:** B1+2c; C1.

WEDGE-BILLED WREN-BABBLER *Sphenocichla humei* occurs in mainland **China** (one record from Yunnan: Han Lianxian 1992), **India** (Sikkim, West Bengal, Arunachal Pradesh, Nagaland, Manipur, but few recent records), **Bhutan** (two recent records) and northern **Myanmar**, but it appears to be a rare and infrequently encountered species throughout, although this impression is at least partly exaggerated because of the infrequency with which its range is visited by ornithologists (Stanford and Ticehurst 1938–1939, Smythies 1949, Ali and Ripley 1968–1998, Ghosh 1987, Katti *et al.* 1990, Han Lianxian 1992, Singh 1994, 1999, Grimmett *et al.* 1998, Inskipp *et al.* 1999a, MacKinnon and Phillipps 2000, D. Farrow *in litt*. 2000). It inhabits the understorey and leaf-litter of broadleaved evergreen forest with large trees and bamboo, usually between 1,500 m and 2,300 m, wintering down to 800 m (Grimmett *et al.* 1998). While the high altitudes favoured by this species are relatively free from habitat loss, shifting cultivation and logging are nevertheless reducing forest cover in Bhutan, north-east India and north Myanmar, posing the only major threat to this species (*Oriental Bird Club Bull*. 21:15–20, Katti *et al*. 1990, 1992, Singh 1994, 1999, Stattersfield *et al.* 1998, D. Farrow *in litt*. 2000). It has been recorded in Thrumshingla National Park, Bhutan, and Namdapha National Park, India. **Taxonomy**: The two races *humei* and *roberti* are sufficiently morphologically distinct to be treated as two separate species (P. C. Rasmussen verbally 1998), which might qualify one or both for threatened status. **Criteria nearly met:** C1; C2a.

PYGMY BABBLER *Stachyris plateni* is endemic to the **Philippines**, where it occurs in two highly distinctive forms, *pygmaea* from Samar and Leyte and nominate *plateni* from Mindanao, with records on the latter from Kalambogan, Malabang, Mt Malindang, Mt Piapayungan, Mt Agtuuganon, Mt Apo, Mt Matutum (all in Dickinson *et al.* 1991), plus Mt Hilong-hilong (material in ZMC: M. Heegaard *in litt.* 1989), Mt Kitanglad (Hornskov 1995a), Cabanglasan in Bukidnon (material in ZMC: M. Heegaard *in litt.* 1989), Tu-od and Na-awan in Misamis Oriental (material in ZMC: M. Heegaard *in litt.* 1989; also Dickinson *et al.* 1991), Mt Lamut (R. J. Timmins *in litt*. 1994), Bislig at PICOP (T. H. Fisher verbally 1997) and Lake Sebu

(R. J. Timmins *in litt.* 1994). It occupies primary forest, adjacent second growth and trees in cultivated areas (Rand 1970b) but appears to be uncommon to rare throughout its range (G. C. L. Dutson *in litt.* 1994, T. H. Fisher verbally 1997). **Criterion nearly met:** B1+2c.

GOLDEN-CROWNED BABBLER *Stachyris dennistouni* is endemic to Luzon in Ilocos Norte and the northern Sierra Madre mountains from Cape Engaño south (including Palanan, Los Dos Cuernos and Minuma) to Aurora province (Dingalan, Maria Aurora and the Talaytay watershed), **Philippines**, where it is fairly common in primary forest and adjacent tall grass areas, but only below 1,000 m, and it is potentially at risk as deforestation in the region continues (Dickinson *et al.* 1991, Jensen and Hornskov 1992, J. Hornbuckle *in litt.* 1994, P. Davidson *in litt.* 1995, Poulsen 1995, D. W. Billing *in litt.* 1997, B. Gee *in litt.* 1997). **Criteria nearly met:** A1c; A2c; B1+2c.

LUZON STRIPED-BABBLER *Stachyris striata* is endemic to Luzon, **Philippines**, where records are from Lamao, Bataan (Gilliard 1950), Mt Mariveles (Fisher mss, Dickinson *et al.* 1991), Cape Engaño, Cagayan (Whitehead 1899b), Mt Cagua (four specimens in FMNH), Penablanca, Cagayan (Dickinson *et al.* 1991), Minuma (Danielsen *et al.* 1994, Poulsen 1995), Siagot (Danielsen *et al.* 1994, Poulsen 1995), Los Dos Cuernos (Danielsen *et al.* 1994, Poulsen 1995), Mt Cetaceo (two specimens in FMNH), San Mariano ("Molino"), Isabela (Whitehead 1899b, Parkes 1971a; nine specimens in AMNH, DMNH, PNM), Dilalongan, Aurora (Dickinson *et al.* 1991), and Angat Dam (Poulsen 1995, T. H. Fisher verbally 1997). The species has been characterised as rare and local, and a bird of forest understorey below 500 m (Dickinson *et al.* 1991), but in the Sierra Madre it is common locally up to 850 m, sometimes occupies the middle and upper storeys, and persists in heavily degraded forest and overgrown clearings (Danielsen *et al.* 1994, Poulsen 1995, T. H. Fisher verbally). Nevertheless, the actual and potential habitat loss within its range is substantial (see, e.g., Threats *Habitat loss: Luzon* under Philippine Eagle *Pithecophaga jefferyi*). **Criteria nearly met:** A1c; A2c; B1+2c.

PANAY STRIPED-BABBLER *Stachyris latistriata* is endemic to Panay, **Philippines**, where it is known from montane mossy forest above 1,100 m on Mt Baloy (Gonzales and Kennedy 1990), Mt Madja-as (78 specimens in CMNH, FMNH, USNM, UPLB; Kennedy 1992, NADM), and Dagsalan, upper Aklan river (F. R. Lambert verbally 1997). It is common in montane mossy forest on Mt Baloy, where it was the most abundant species (Gonzales and Kennedy 1990) and on Mt Madja-as (NADM); it is likely to occupy this habitat throughout the estimated 700 km^2 of Panay's remaining closed-canopy high-elevation forest, much of which is on steep slopes above 45° and hence relatively safe from both commercial logging and *kaingin* (NADM). **Criterion nearly met:** B1+2c.

PALAWAN STRIPED-BABBLER *Stachyris hypogrammica* is endemic to Palawan, **Philippines**, where it is known from primary montane forest areas (above 1,000 m) in the south of the island, as follows: Mt Victoria (three specimens in CMNH), Mt Mataling (Salomonsen 1961), Mt Gorang Bato (Dickinson *et al.* 1991), Mt Mantalingahan at the peak and at Magtaguimbong (16 specimens in DMNH, FMNH, USNM). All the evidence, including that of the original and recent observers, points to this species being common and relatively secure within its highly restricted but fairly pristine range (Salomonsen 1961, Collar *et al.* 1994), as with Panay Striped-babbler *Stachyris latistriata*. **Criterion nearly met:** B1+2c.

WHITE-BREASTED BABBLER *Stachyris grammiceps* is restricted to the Java and Bali Forests Endemic Bird Area, occurring in lowland and hill forest at 0–1,000 (–1,400) m in West Java, **Indonesia** (Stattersfield *et al.* 1998). Despite concerns that it is now highly

fragmented within its very restricted range (Collar and Andrew 1988, Collar *et al.* 1994), it has been found to be extremely common at Gunung Halimun, with a density calculated at 332 birds per km^2 in primary and slightly disturbed forest between 750 and 1,100 m (D. Liley *in litt*. 1999). **Criteria nearly met:** A1c; A2c; B1+2a,b,c,d,e.

SOOTY BABBLER *Stachyris herberti* was historically known from only two lowland localities in central **Laos**, where it was first collected around 1920 (Baker 1920), but in 1994 was found to be not uncommon in primary forest at a single locality in central Annam, **Vietnam** (Eames *et al.* 1995a). Recent observations suggest that it is locally common in Laos over a fairly wide area (Duckworth *et al*. 1998b, P. Davidson *in litt*. 1998, Duckworth *et al.* 1999, Robson 2000). It frequents stunted evergreen forest in limestone gullies, steep cliffs and adjacent rock, at 230–610 m (Duckworth *et al.* 1999, Robson 2000). It is presumably threatened by lowland deforestation (Collar *et al.* 1994), although this is proceeding at an extremely slow rate in the inaccessible karstic regions it occupies (Duckworth *et al.* 1999). It occurs in a few protected areas: Hin Nam No, Khammouane Limestone and Nam Kading NBCAs, Laos, and Phong Nha Nature Reserve, Vietnam (Duckworth *et al.* 1999). **Criterion nearly met:** C1.

WHITE-NECKED BABBLER *Stachyris leucotis* occurs in the Sundaic lowlands, from peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan and Sumatra, **Indonesia** and **Brunei** in hill forest to 1,000 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's numerical strength and preference for hill-slope forests (Smythies 1981), which have been subject to considerably less destruction, implies that it is not unduly threatened. **Criteria nearly met:** A1c; A2c.

BLACK-THROATED BABBLER *Stachyris nigricollis* occurs in the Sundaic lowlands, from peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct), Kalimantan and Sumatra, **Indonesia** and **Brunei** in forest, including freshwater and peatswamp forest, to 1,000 m (Smythies 1981, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*). In addition, peatswamp forest has been severely affected by recent ENSO- and arson-related conflagrations (see Threats under Hook-billed Bulbul *Setornis criniger*). However, the species's use of submontane forests implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

CHESTNUT-RUMPED BABBLER *Stachyris maculata* occurs in the Sundaic lowlands, from peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct), Kalimantan and Sumatra, **Indonesia** and **Brunei**, in broadleaf forest to 800 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's ability to persist in hill-slope forests implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

FLUFFY-BACKED TIT-BABBLER *Macronous ptilosus* occurs in the Sundaic lowlands, from peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct),

Kalimantan (including Anamba Islands) and Sumatra (including offshore islands), **Indonesia** and **Brunei**, where it is locally very common in lowland evergreen forest (including freshwater swamp forest, second growth and bamboo) to 700 m (Smythies 1981, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia, but the species's ability to survive in second growth (Smythies 1981, MacKinnon and Phillipps 1993) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

ORANGE-BILLED BABBLER *Turdoides rufescens* is an endemic resident in **Sri Lanka**, where it is fairly common to common in the wet lowlands, usually rare on adjacent hills but occasionally common above 700 m (Jones *et al.* 1998, Grimmett *et al.* 1998). It appears to be almost restricted to primary forest in some areas (usually where Yellow-billed Babbler *T. affinis* is present), but is sometimes common in selectively logged or secondary forest, scrub or tea plantations, and is a characteristic component of mixed-species flocks, with up to 45 individuals counted in a group (Jones *et al.* 1998). Forest on the island has suffered degradation and fragmentation in recent decades owing to excessive gathering of fuelwood, clearance for permanent agriculture, shifting cultivation, fire, urbanisation and logging (Stattersfield *et al.* 1998). Within such a restricted range there is a danger that severe clearance of wet-zone vegetation would eradicate this species in the long term. **Criteria nearly met:** B1+2a,b,c,d,e.

GIANT BABAX *Babax waddelli* occurs in southern Tibet, mainland **China**, and extreme north-eastern Sikkim, **India**, where it inhabits dense deciduous scrub above the treeline, particularly of *Hippophae rhamnoides*, and the edge of coniferous forest, at c.2,700–4,400 m (Cheng Tso-hsin 1987, Grimmett *et al.* 1998, Stattersfield *et al.* 1998). Within its restricted range, it has been described as "locally common" (Ali and Ripley 1968–1998) but, more recently, "rather scarce" (P. Alström *in litt.* 1993). It is presumably declining because of deforestation, although extensive pine and mixed coniferous forests with prickly oak and rhododendron still remain to the east of Lhasa (Robson 1986). **Criteria nearly met:** C1; C2a.

TIBETAN BABAX *Babax koslowi* is restricted to the eastern part of the Qinghai–Tibetan Plateau, in eastern Tibet and southern Qinghai, mainland **China**, where several large valleys cut into the plateau, including those of the Tongtian river (the upper reaches of the Chang Jiang or Yangtze), the Lancang Jiang (the upper Mekong) and the Nu Jiang (the upper Salween) (Cheng Tso-hsin 1987, Stattersfield *et al.* 1998). It is known by just a few scattered records in this inaccessible and poorly known area, but it appears to be genuinely rather scarce and localised (Robson 1986, P. Alström *in litt.* 1993), in juniper forest and scrub, mixed fir and juniper forest and scrub bordering agricultural land at c.3,650–4,500 m (Cheng Tso-hsin 1987, Stattersfield *et al.* 1998). There is some logging of forest within its range, but it does not appear to be immediately threatened by habitat loss (P. Alström *in litt.* 1993). **Criteria nearly met:** C1; C2a.

BROWN FULVETTA *Alcippe brunneicauda* occurs in the Sundaic lowlands of peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan (including the Natuna islands) and Sumatra (including the Batu islands), **Indonesia** and **Brunei**, where it is fairly common in broadleaf forest to 1,000 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's preference for submontane forests, at least in Borneo (Smythies 1981), and use of second growth

(MacKinnon and Phillipps 1993) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

SPOTTED CROCIAS *Crocias albonotatus* is restricted to the Java and Bali Forests Endemic Bird Area, occurring in montane forest at 900–2,400 m in West Java, **Indonesia** (Stattersfield *et al*. 1998). It is moderately common in Gunung Gede-Pangrango National Park (R. F. A. Grimmett verbally 1999) but seemingly rare on Gunung Halimun (D. Liley *in litt*. 1999) and is vulnerable to forest loss and perhaps trapping at lower levels. **Criteria nearly met:** B1+2a,b,c,d,e.

REED PARROTBILL *Paradoxornis heudei* is known from Lake Khanka in Primorye, **Russia**, **Mongolia**, and Heilongjiang, Liaoning, Inner Mongolia, Shandong, Jiangsu, Shanghai, Zhejiang and Jiangxi, mainland **China**, where it is locally common in reedbeds bordering rivers, lakes and the coast (Cheng Tso-hsin 1987, Rank 1989, *Oriental Bird Club Bull.* 19 [1994]: 18–19, Melville and Li Xuxiang 1998, P. Alström, U. Olsson and D. Zetterström *in litt.* 2000). It is presumably declining, as several of the sites where it has been recorded recently in eastern China are under severe threat (Rank 1989), and reedbeds appear to be under pressure everywhere in China (P. Alström *in litt.* 2000). **Criteria nearly met:** A1c; A2c.

RUFOUS-VENTED PRINIA *Prinia burnesii* has two disjunct populations, one (nominate *burnesii*) in the plains of the Indus, **Pakistan** (in Punjab and northern Sind; scarce in southern Sind) and adjacent north-west **India** (Punjab), and another (race *cinerascens*) in the plains of the Brahmaputra river, Assam, and western Bihar in north-east India and northern **Bangladesh** (Ali and Ripley 1968–1998, Roberts 1991–1992, Grimmett *et al.* 1998). The western population is locally common in its restricted habitat, as was the eastern population, although there have only been a few recent published records, mostly from Assam (Ali and Ripley 1968–1998, Roberts 1991–1992, Barua 1995b, Grimmett *et al.* 1998, Showler and Davidson 1999). There are no recent records from Bangladesh; it could still occur in any remaining suitable habitat in the north although these are now almost totally destroyed (P. M. Thompson *in litt*. 1997, Grimmett *et al.* 1998). It inhabits long grasses, sometimes where these are mixed with acacias and tamarisks, including *Saccharum* wet grassland in monoculture or where mixed with *Typha*, *Phragmites* and bushes, mainly in the vicinity of large rivers (Ripley 1982, Grimmett *et al.* 1998, Showler and Davidson 1999). Although it is negatively affected by the destruction and modification of this habitat (see Majumdar and Brahmachari 1988, Rahmani 1988b, Roberts 1991–1992, and Threats under Jerdon's Babbler *Chrysomma altirostre*) it has also apparently expanded its range northwards in Pakistan owing to the spread of man-modified wetlands and associated swamp vegetation (T. J. Roberts verbally 1998). **Taxonomy**: The two races *burnesii* and *cinerascens* are sufficiently morphologically distinct to be treated as two separate species (P. C. Rasmussen verbally 1998), which might qualify one or both for threatened status. **Criteria nearly met:** A1c; A2c.

TANIMBAR BUSH-WARBLER *Cettia carolinae* is restricted to the Banda Sea Islands Endemic Bird Area, **Indonesia**, where it is found only on the Tanimbar islands (Stattersfield *et al*. 1998). It is common in primary and secondary forest on Yamdena (Coates and Bishop 1997, Bishop and Brickle 1998), but there is significant logging in the south of the island (Bishop and Brickle 1998; see Threats under Lesser Masked-owl *Tyto sororcula*), so the species is likely to be in steady decline. **Criterion nearly met:** B1+2c.

LONG-BILLED BUSH-WARBLER *Bradypterus major* occurs in the western Himalayas, in Xinjiang, western mainland **China** (very rare, occurring in the Kunlun mountains of Xinjiang and west Tibet) and the eastern Pamir mountains in **Tajikistan** (perhaps only a vagrant),

northern **Pakistan** (few recent records and distribution imperfectly known) and north-west **India** (fairly common, but extremely local), breeding from 2,400 to 3,600 m (probably moving downslope in winter to 1,200 m) on open slopes in terraced cultivation, low thorny scrub, rank grass and bracken, often near forest edges (Bates and Lowther 1952, Ali and Ripley 1968–1998, Cheng Tso-hsin 1987, Collar and Andrew 1988, Roberts 1991–1992, Inskipp and Collins 1993, MacKinnon and Phillipps 2000). Its range is apparently contracting in Kashmir, possibly owing to changes in agricultural practices (Collar and Andrew 1988). It is poorly known and infrequently recorded, but this is at least in part due to its highly secretive behaviour and the current inaccessibility of its range. **Criteria nearly met:** C1, C2a.

SRI LANKA BUSH-WARBLER *Bradypterus palliseri* is an endemic resident in **Sri Lanka**, where it inhabits dense forest undergrowth, particularly along streams above 1,500 m, but is also recorded down to 350 m, occupying a very small range, within which it is apparently rare, although suitable habitat remains relatively intact and is disappearing at too slow a rate to assume any immediate threat (Grimmett *et al.* 1998, Stattersfield *et al.* 1998). **Criteria nearly met:** B1+2a,b,c,d,e; C1.

TYTLER'S LEAF-WARBLER *Phylloscopus tytleri* breeds in the western Himalayas in **Pakistan** and Kashmir, **India** (generally between 2,000 and 3,600 m), passes through **Nepal** in small numbers during migration and winters mainly in the Western Ghats of peninsular India, but is scarce to locally common in Pakistan and Kashmir, rare in Nepal, and infrequently recorded in its wintering range, partly owing to identification difficulties (Rasmussen 1998g, Grimmett *et al.* 1998). During the breeding season it inhabits coniferous forest, also subalpine dwarf willows and birches, while in winter its habitat preferences are less clear, with some records from the middle storey of shola forest (Grimmett *et al.* 1998). In its breeding range, forests are under constant threat from timber extraction, excessive cutting for fuelwood and animal fodder, livestock-grazing and burning (Stattersfield *et al.* 1998). In its wintering range, increasing encroachment into forests, livestock-grazing, hydropower development, road-building and the harvesting of fuelwood and huge quantities of forest products, such as bamboo and canes, are causing reductions in forest cover in the Western Ghats (Stattersfield *et al.* 1998). **Criterion nearly met:** C1.

YELLOW-BREASTED TAILORBIRD *Orthotomus samarensis* is endemic to Samar, Leyte and Bohol, **Philippines**, inhabiting forest and forest edge in dense undergrowth (Dickinson *et al.* 1991). Although characterised as the rarest member of the genus in the Philippines (Hachisuka 1936), and rare or very rare on Bohol and Samar (Rand and Rabor 1960), it is very shy (Ogilvie-Grant 1897, Brooks *et al.* 1995c) and is moderately common in Rajah Sikatuna National Park (N. Bostock verbally 1993, J. Hornbuckle *in litt.* 1994); the extent to which it suffers from habitat loss is unclear. On voice it should be considered conspecific with Black-headed Tailorbird *O. nigriceps* (Hornskov 1995a). **Criteria nearly met:** A1c; A2c; B1+2c; C2a.

RUFOUS-RUMPED GRASSBIRD *Graminicola bengalensis* is found in the terai of northern **India** and southern **Nepal** (highly localised but common at a few sites from Uttar Pradesh eastwards to the Brahmaputra lowlands: Grimmett *et al.* 1998), **Bangladesh** (rare and highly local in the north: Grimmett *et al.* 1998), mainland **China** (rare resident in Guanxi, northern Guangdong, Hong Kong and Hainan: MacKinnon and Phillipps 2000), **Myanmar** (formerly resident in Tenasserim, current status unknown: Robson 2000), **Thailand** (former resident in central region, last record in 1923, probably extinct: Robson 2000) and **Vietnam** (former resident in east Tonkin: Robson 2000). It occurs in tall emergent vegetation in or bordering freshwater swamps or along banks of rivers in the lowlands (Grimmett *et al.* 1998, Robson 2000). Although still common at a few sites, including protected areas such as Chitwan

National Park, Nepal, it must be suffering substantial long-term losses as its grassland and wetland habitat is steadily converted to agriculture, drained and overgrazed (see Threats under Bengal Florican *Houbaropsis bengalensis*). **Criteria nearly met:** A1c; A2c; C1; C2a.

GOLDEN-BELLIED FLYROBIN *Microeca hemixantha* is restricted to the Banda Sea Islands Endemic Bird Area, **Indonesia**, where it inhabits forest and mangroves in the Tanimbar Islands (Yamdena, Loetoe and Larat) (Stattersfield *et al.* 1998). It is common within its small range (Coates and Bishop 1997), but there is significant logging in the south of Yamdena (Bishop and Brickle 1998; see Threats under Lesser Masked-owl *Tyto sororcula*), so the species is likely to be in steady decline. **Criterion nearly met:** B1+2c.

OLIVE-YELLOW ROBIN *Poecilodryas placens* is very patchily distributed in New Guinea (Papua, formerly Irian Jaya, **Indonesia**, and **Papua New Guinea**), being recorded from Batanta Island, Wandammen mountains, Fakfak mountains, Kumawa mountains, Weyland mountains, Keku near Madang, Lake Kutubu, Mt Bosavi, Karimui, and a number of sites in Central Province (Beehler *et al.* 1986, Coates 1990, Diamond 1985). There are new recent records from Crater mountain (A. Mack *in litt.* 1999), where it is patchily abundant, and in limestone hill forest from Moro to Gobe, Gulf Province, where it is locally common (K. D. Bishop *in litt.*1999). It may prove to be more widespread through the central mountains but is believed to be absent from many intervening areas (Diamond 1985). It is a hill forest species, occurring to 1,450 m, and the one well-known site, at Veimauri, near Port Moresby, is being logged (P. Gregory *in litt.* 1994). Although the small total population may be isolated into subpopulations, some of which may be threatened by logging, its extensive and often inaccessible range suggests that there are more, safer subpopulations to be discovered. **Criterion nearly met:** C2a.

MAROON-BREASTED PHILENTOMA *Philentoma velatum* occurs in the Sundaic lowlands, from southern Tenasserim, **Myanmar**, peninsular and west **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct: Lim Kim Seng *in litt*. 2001), Kalimantan, Sumatra and Java, **Indonesia** and **Brunei** (very scarce), in broadleaf evergreen forest, including swamp forest, to 1,400 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's use of submontane forests implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

STREAKY-BREASTED JUNGLE-FLYCATCHER *Rhinomyias addita* is restricted to the Buru Endemic Bird Area, **Indonesia**, where it inhabits lowland and montane forest at 500–1,500 m (Stattersfield *et al.* 1998). It has been described as a rare bird of the higher mountain regions (Stresemann 1914b), but it was found to be a moderately common bird in logged and primary forest inland from Fogi at elevations of 500–1,100 m, late 1995 (MKP). Some deforestation has occurred in Buru's coastal lowlands, large areas have been disturbed and selectively logged, and others are targeted for conversion (see Threats under Blue-fronted Lorikeet *Charmosyna toxopei* and Rufous-throated White-eye *Madanga ruficollis*), but a relatively widespread and common upland species which can use logged forest is probably not threatened by this circumstance. **Criterion nearly met:** B1+2c.

GREY-CHESTED JUNGLE-FLYCATCHER *Rhinomyias umbratilis* is restricted to the Sundaic lowlands, from peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**,

Kalimantan and Sumatra (including offshore islands), **Indonesia** and **Brunei**, in lowland broadleaf (including peatswamp) forest to 1,160 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Dryland and swamp forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma* and Hook-billed Bulbul *Setornis criniger*), but the species's use of submontane forests, second growth and plantations (MacKinnon and Phillipps 1993) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

HENNA-TAILED JUNGLE-FLYCATCHER *Rhinomyias colonus* is restricted (if one excludes dubious records from East Sulawesi attributed to the race "*subsolanus*") to the Banggai (Peleng: race *pelingensis*) and Sula Islands (Taliabu, Seho, Mangole and Sanana: race *colonus*) Endemic Bird Area, **Indonesia**, where it inhabits lowland forest up to 300 m, and is apparently uncommon on Peleng and generally so on the Sulas (Stattersfield *et al*. 1998; also Coates and Bishop 1997). Although it has been found in degraded forest, its preferred habitat appears to be undisturbed forest, and this, combined with its restriction to the lowlands, is a cause for some concern (Davidson *et al*. 1995). **Criteria nearly met:** A1c; A2c; B1+2a,b,c,d,e.

SLATY-BACKED JUNGLE-FLYCATCHER *Rhinomyias goodfellowi* is endemic to Mindanao, **Philippines**, where it is known from six upland localities in the centre of the island, as follows: Civolig near Gingoog City (Meyer de Schauensee and duPont 1962); Daggayan inland from Gingoog City (Meyer de Schauensee and duPont 1962); Mt Kitanglad (Salomonsen 1952, Ripley and Rabor 1961, Gonzalez and Mallari 1993, Heaney *et al.* 1993); Mt Apo (Ogilvie-Grant 1906b, McGregor 1920, Gibbs 1984, 24 specimens in CMNH, many observers); Lake Sebu (R. J. Timmins *in litt.* 1994, B. Gee *in litt.* 1997); and Mt Busa (seven specimens in CMNH). It was regularly encountered in 1992 and 1993 on Mt Kitanglad, and no fewer than 31 specimens were collected in 22 days' fieldwork, 3–16 February, 3–9 March and 16–18 April 1993 (CMNH register data). It appears, therefore, to be reasonably numerous within the confines of its range, and as a species of the dipterocarp/mossy forest ecotone it is unlikely to be at great risk (A. C. Diesmos verbally 1996, BRT). **Criterion nearly met:** B1+2c.

SUMBA BROWN FLYCATCHER *Muscicapa segregata* (when split from Asian Brown Flycatcher *M. dauurica*) is restricted to the Sumba Endemic Bird Area, **Indonesia** (Stattersfield *et al*. 1998). An early suspicion was that it is "rather scarce and local" (White and Bruce 1986), a more recent assessment considered it "uncommon and sparse, perhaps rare" but indicated it to be "quiet and unobtrusive" (Coates and Bishop 1997), and the most recent observations confirm that it is widespread in remaining patches of both primary and secondary forest, albeit declining with steady habitat clearance (M. J. Jones *et al*. 1995, many observers *in litt*. 1999). **Criteria nearly met:** B1+2a,b,c,d,e.

RUFOUS-CHESTED FLYCATCHER *Ficedula dumetoria* occurs in the Sundaic lowlands of peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan, north Sumatra and principally West Java, **Indonesia** and **Brunei**, in bamboo and forest, locally to 1,500 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's use of hill and submontane forests (MacKinnon and Phillipps 1993) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

RUFOUS-THROATED FLYCATCHER *Ficedula rufigula* is restricted to the Sulawesi Endemic Bird Area (but not the East Peninsula), **Indonesia**, where it is locally common and widespread in the undergrowth of primary lowland and hill forest up to 600 m, occasionally to 1,000 m (Stattersfield *et al*. 1998; also Coates and Bishop 1997). Forest destruction within its elevation range has been extensive in recent decades (see Threats under Lombopatang Flycatcher *Ficedula bonthaina*), and its populations must have suffered a commensurate decline. **Criteria nearly met:** A1c; A2c.

FURTIVE FLYCATCHER *Ficedula disposita* is endemic to Luzon, **Philippines**, where it was considered threatened but (like Mindanao's Cryptic Flycatcher *F. crypta*, now considered at Low Risk owing to recent fieldwork) is no longer regarded as particularly rare, with records from above Crow Valley, Zambales mountains (Ripley and Marshall 1967); Mt High Peak, Zambales (Kennedy and Ruedas 1992); Minuma (Poulsen 1995); Siagot (Poulsen 1995); Talisis valley, Dinalongan, Aurora (D. W. Billing *in litt.* 1997, T. H. Fisher verbally 1997); two sites on the coastal road near Debutunan Point, Aurora (T. H. Fisher verbally 1997); and Angat Dam (Dutson 1993a; also Evans *et al.* 1993a, P. A. J. Morris *in litt.* 1997, AJL). It has now been found (on voice) to be relatively common, e.g. in forest along the main coastal road between Baler and Dinalongan in Aurora (T. H. Fisher verbally 1997). The species inhabits lowland forest, which is under great pressure, but appears to survive well in selectively logged and degraded areas (Dutson 1993a, Poulsen 1995, T. H. Fisher verbally 1997). **Criteria nearly met:** A1c; A2c; B1+2c.

BLACK-AND-RUFOUS FLYCATCHER *Ficedula nigrorufa* is an endemic resident in the Western Ghats of southern **India**, where it is locally common from 700 m to the highest summits (Grimmett *et al.* 1998). It inhabits shola forests with dense undergrowth and plenty of leaf-litter, cardamom and coffee plantations, and moist thickets in ravines (Grimmett *et al.* 1998). While its range is small, its tolerance for modified habitats suggests that it is not immediately threatened. However, an increasing human population has led to increased illegal encroachment into Western Ghat forests, livestock grazing and the harvesting of fuelwood and huge quantities of forest products such as bamboo and canes (Stattersfield *et al.* 1998). Furthermore, hydropower development and road-building are causing reductions in forest cover in some areas (Stattersfield *et al.* 1998). These factors require monitoring as, left uncontrolled, they will threaten the area's endemic avifauna. **Criteria nearly met:** B1+2a,b,c,d,e.

BLACK-BANDED FLYCATCHER *Ficedula timorensis* is restricted to the Timor and Wetar Endemic Bird Area, **Indonesia**, where it occurs on Timor at 0–1,200 m in the dense undergrowth of monsoon forest, apparently preferring areas with limestone boulders and rocky scree slopes (Stattersfield *et al*. 1998; also Collar and Andrew 1988, Coates and Bishop 1997). Although it has been found in degraded patches (R. Noske *in litt*. 2000), it appears to favour areas with closed-canopy forest (P. Andrew *in litt*. 2000), which are constantly diminishing (see Threats under Wetar Ground-dove *Gallicolumba hoedtii*); moreover, the ground vegetation in all such patches of forest is grazed by cattle (Collar and Andrew 1988). **Criteria nearly met:** A1c; A2c.

DULL-BLUE FLYCATCHER *Eumyias sordida* is an endemic resident in the uplands of **Sri Lanka**, where it is generally common from 1,220–1,830 m, scarce at lower altitudes, occurring in well-wooded areas and at forest edges, sometimes in shady gardens (Grimmett *et al.* 1998). While it may be resilient to forest degradation because of this preference for edge habitats, total clearance of habitat is a potential threat in its very small range. Forest on the island has suffered rapid degradation and fragmentation in past decades through excessive gathering of fuelwood, clearance for permanent agriculture, shifting cultivation, fire, urbanisation and

logging (Stattersfield *et al.* 1998). Closed-canopy forest is estimated to have declined from 29,000 km² (44% of the island's area) in 1956 to 12,260 km² in 1983 (Collins *et al.* 1991). It is feared that this loss will continue (Stattersfield *et al.* 1998). **Criteria nearly met:** B1+2a,b,c,d,e.

NILGIRI FLYCATCHER *Eumyias albicaudata* is an endemic resident in the Western Ghats of southern **India**, where it is common at forest edges, clearings, thick vegetation near streams, cardamom plantations and sholas, from 600 m to the summits (Grimmett *et al.* 1998). While its range is small, its tolerance of disturbed habitats suggests that it is not immediately threatened by habitat modification. However, an increasing human population has led to increased illegal encroachment into Western Ghat forests, livestock-grazing and the harvesting of fuelwood and huge quantities of forest products such as bamboo and canes (Stattersfield *et al.* 1998). Furthermore, hydropower development and road-building are causing reductions in forest cover in some areas (Stattersfield *et al.* 1998). **Criterion nearly met:** B1+2c.

PALAWAN BLUE-FLYCATCHER *Cyornis lemprieri* is endemic to Palawan, Balabac, Culion and Calauit, **Philippines**, where it is uncommon to locally common in lowland forest and second growth (Dickinson *et al.* 1991), with records including St Paul's Subterranean River National Park, Iwahig Penal Colony and Mt Mantalingahan (Jensen and Hornskov 1992, Hornbuckle 1994, R. J. Timmins *in litt.* 1994, P. Davidson *in litt.* 1995), with a breeding record from Mt Mananangob (Vrettos 1993) and a good population at Singnapan (Sison 1983). The extent to which it suffers from habitat loss is unclear. **Criteria nearly met:** A1c; A2c; B1+2c.

MALAYSIAN BLUE-FLYCATCHER *Cyornis turcosus* occurs in the Sundaic lowlands, from peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan and Sumatra, **Indonesia** and **Brunei** in bamboo, swamp and evergreen forest, usually in association with waterways, locally to 800 m (Smythies 1981, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Dryland and swamp forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma* and Hook-billed Bulbul *Setornis criniger*), but the species's use of secondary forest and brushwood of abandoned plantations (Smythies 1981) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

SHORT-CRESTED MONARCH *Hypothymis helenae* is endemic to the **Philippines** where it is distributed widely but patchily, on Camiguin Norte (race *personata*), Luzon, Polillo, Catanduanes and Samar (nominate *helenae*), Dinagat, Siargao and Mindanao (race *agusanae*), occupying the understorey of forest below 1,000 m; it is everywhere rare (very few records on Luzon and Mindanao) except on Camiguin Norte and Dinagat (Dickinson *et al.* 1991; also F. R. Lambert verbally 1993, N. J. Redman verbally 1993), although even in these cases there is room for doubt, the records from Camiguin Norte dating from 1907 (McGregor 1907f) and that from Dinagat (duPont and Rabor 1973b) not being confirmed in 1994 (J. Hornbuckle *in litt.* 1994); on the other hand it was supposedly common on Samar and locally so on Luzon (Rand 1970a). The extent to which it suffers from habitat loss is unclear. **Criteria nearly met:** A1c; A2c.

JAPANESE PARADISE-FLYCATCHER *Terpsiphone atrocaudata* breeds in the humid forests of southern Honshu, Shikoku, Kyushu and the Nansei Shoto islands in **Japan**, **South Korea**, **Taiwan** (including Lanyu island) (China) and the extreme northern **Philippines**, and has been recorded as a non-breeding visitor to mainland **China**, **Hong Kong**, **Thailand**, **Laos**, **Vietnam**, Philippines, **Malaysia**, **Singapore**, and Sumatra, **Indonesia** (Austin 1948, Gore and

Won 1971, King *et al.* 1975, Brazil 1991, Dickinson *et al.* 1991, Lim Kim Seng *in litt*. 2001). A recent survey detected a steep decline in part of the Japanese breeding population (H. Higuchi verbally 1997; also Brazil 1991), which has presumably occurred because of forest loss and degradation in its winter range. **Criteria nearly met:** A1c; A2c.

BLUE PARADISE-FLYCATCHER *Terpsiphone cyanescens* is endemic to the Palawan group (Palawan, Busuanga, Culion and Bantac), and **Philippines**, appears to be fairly common (100 records, 1964–1970: McClure 1974) throughout the island (including St Paul's Subterranean River National Park), but is confined to primary and secondary forest, mostly in the lower-lying areas, and hence it remains exposed to population decline as a result of continuing habitat conversion (Zimmer 1918b, Sison 1983, Dickinson *et al.* 1991, Hornbuckle 1994, R. J. Timmins *in litt*. 1994), especially as it does not occur in isolated tracts of secondary growth (N. Bostock verbally 1994). **Criteria nearly met:** A1c; A2c; B1+2c.

WHITE-TAILED MONARCH *Monarcha leucurus* is restricted to the Banda Sea Islands Endemic Bird Area, **Indonesia**, where it inhabits forest on the Kai Islands (Stattersfield *et al.* 1998). It is moderately common (Coates and Bishop 1997), but it may be vulnerable to forest loss within its small range. **Criterion nearly met:** B1+2c.

BIAK FLYCATCHER *Myiagra atra* is endemic to Biak-Supiori, Numfor and Rani in Geelvink Bay, Papua (formerly Irian Jaya), **Indonesia** (Beehler *et al.* 1986, Mayr and Meyer de Schauensee 1939). It is apparently largely confined to the interior hills, inhabiting primary, secondary and logged forest up to c.400 m, but is also occasionally recorded in mangroves (Mayr and Meyer de Schauensee 1939, Beehler *et al.* 1986, Gibbs 1993). Observations from recent visits suggest that it may be fairly common on Biak (N. Bostock *in litt.* 1993, Gibbs 1993, Eastwood 1996, B. M. Beehler and SvB *in litt*. 2000, M. van Beirs *in litt.* 2000). On Biak and Numfor, forest is under heavy threat from logging and subsistence farming, but there appear to be large areas of forest remaining in interior Supiori (Bishop 1982, K. D. Bishop *in litt*. 1996, D. A. Holmes *in litt.* 2000). Its reported abundance, tolerance of degraded forest and occurrence in the 110 km^2 Biak-Utara protected area (B. M. Beehler and SvB *in litt*. 2000) suggest that it is not threatened. **Criterion nearly met:** C1.

CINNAMON-TAILED FANTAIL *Rhipidura fuscorufa* is restricted to the Banda Sea Islands Endemic Bird Area, **Indonesia**, where it inhabits forest and mangroves on Babar and the Tanimbar Islands (Stattersfield *et al.* 1998). It is common within its small range, but there is significant logging in the south of Yamdena (Bishop and Brickle 1998; see Threats under Lesser Masked-owl *Tyto sororcula*), so the species is likely to be in steady decline. **Criterion nearly met:** B1+2c.

LONG-TAILED FANTAIL *Rhipidura opistherythra* is restricted to the Banda Sea Islands Endemic Bird Area, **Indonesia**, where it is found only on the Tanimbar Islands (Yamdena, Larat and Maroe) (Stattersfield *et al.* 1998). It is moderately common in the interior of primary and secondary forest, but there is significant logging in the south of Yamdena (Bishop and Brickle 1998; see Threats under Lesser Masked-owl *Tyto sororcula*), so the species is likely to be in steady decline. **Criterion nearly met:** B1+2c.

PALAWAN TIT *Parus amabilis* is considered "generally uncommon" on Palawan, Balabac and Calauit, **Philippines**, where it is possibly dependent on primary forest, and its occurrence at higher elevations (as reported in Dickinson *et al.* 1991) remains to be confirmed, so that it must be a potential cause for concern owing to the clearance and degradation of lowland

forest that is taking place within its range (Harrap and Quinn 1996, NADM). **Criteria nearly met:** A1c; A2c; B1+2c.

YELLOW TIT *Parus holsti* is confined to the mountains of central **Taiwan** (China), where it occurs in primary broadleaf forest and occasionally secondary growth at 700–2,500 m (Severinghaus and Blackshaw 1976, Stattersfield *et al.* 1998). It is generally found at low densities (Chang and Severinghaus 1979, Severinghaus 1989), and it was one of the species reported to be captured during large-scale netting of wild birds for export (Lin Chih-cheng 1987). However, much of its habitat is secure, as c.11% of Taiwan is protected in national parks, nature reserves and wildlife sanctuaries (P. K. D. Perng *in litt.* 1996). **Criteria nearly met:** C1; C2a.

WHITE-FRONTED TIT *Parus semilarvatus* is endemic to Luzon (race *snowi* in the northern Sierra Madre, nominate *semilarvatus* in the central and southern Sierra Madre and a few other sites to the south and west) and Mindanao (race *nehrkorni*, known from some eight localities), **Philippines**, inhabiting forest, edge and second growth up to around 1,150 m; it is generally characterised as "rare and local", possibly as a result of some unidentified biological specialisation, although it is moderately common in the Sierra Madre (Dickinson *et al.* 1991, Poulsen 1995, Harrap and Quinn 1996). **Criteria nearly met:** A1c; A2c.

YUNNAN NUTHATCH *Sitta yunnanensis* is known from south-east Tibet, southern Sichuan and northern Yunnan, and the western extreme of Guizhou, mainland **China** (Cheng Tso-hsin 1987; also Wu Zhikang *et al.* 1986). It affects open mature pine forest (avoiding other types of coniferous forest) with little undergrowth or scrub, mainly from 2,400 to 3,400 m (Dolan 1938, C. R. Robson *in litt.* 1990). It is locally common, but it has disappeared from several of the localities where it was recorded in the early twentieth century, and is presumably continuing to decline because of logging and forest fires (Yang Lan *in litt.* 1997; see Smil 1984, 1993, MacKinnon *et al.* 1996), although it does appear able to adapt to degraded and secondary forest (M. A. S. Beaman *in litt.* 1993). **Criteria nearly met:** A1c; A2c; C1; C2a.

YELLOW-BILLED NUTHATCH *Sitta solangiae* is known from three or four widely disjunct areas: the Fan Si Pan mountains in north-east Tonkin and Da Lat plateau in southern Annam, **Vietnam**, the Kon Tum plateau of Vietnam and south-east **Laos**, and Hainan island, south-east mainland **China** (Robson 2000). There are recent records from Hainan (King and Liao 1989, P. Davidson and D. Farrow verbally 1999), several localities on the Da Lat plateau, where surveys since 1990 have recorded small numbers (Eames *et al.* 1992, Robson *et al.* 1993a,b), and the Xe Sap region in Laos (Timmins and Vongkhamheng 1996, Showler *et al.* 1998b). It occurs in primary and logged evergreen forest at 900–2,500 m (Robson 2000). The major threat throughout is deforestation, e.g. forest cover on Hainan fell from an estimated 8,630 km^2 (25.7%) in 1949 to c.2,420 km^2 (7.2%) in 1991 and much of the Da Lat plateau is being cleared by shifting cultivation, charcoal burning and fuelwood collection, factors that are intensifying after government resettlement programmes (Eames 1995a, Stattersfield *et al.* 1998). It occurs in several protected areas such as Bawangling and Jiangfengling Nature Reserves, Hainan, Chu Yang Sin and Ngoc Linh Nature Reserves, Vietnam, and Xe Sap NBCA, Laos. **Criteria nearly met:** C1; C2a.

SCARLET-BREASTED FLOWERPECKER *Prionochilus thoracicus* occurs in the Sundaic lowlands of peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan and north Sumatra (including the Lingga archipelago and Belitung), **Indonesia** and **Brunei**, where it is widespread and fairly common in forest (including swamp and heath forest) to 1,280 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson

2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia, but the species's use of montane and secondary forests (Smythies 1981) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

BROWN-BACKED FLOWERPECKER *Dicaeum everetti* occurs in the lowlands of Sarawak and Peninsular **Malaysia**, Kalimantan, the Natuna Islands and Riau archipelago, **Indonesia** and **Brunei**, in forest (including swamp forest) and gardens to 1,100 m (Smythies 1981, Mann 1987, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's use of submontane and secondary forests, gardens and coffee plantations (MacKinnon and Phillipps 1993) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

WHISKERED FLOWERPECKER *Dicaeum proprium* is endemic to Mindanao, **Philippines**, where it is found in forest, edge and second growth above 800–900 m at Mt Mayo (Ripley and Rabor 1966); Mainit, Manticao, Misamis Oriental (female in DMNH; also Dickinson *et al.* 1991); Mt Kitanglad on the lower slopes of Mt Nangkabulos (Gonzalez and Mallari 1993, Heaney *et al.* 1993); Mt Piapayungan at Saronayan, Lumba-Bayabao (Dickinson *et al.* 1991; female in AMNH); Mt Apo (Fisher mss, Dickinson *et al.* 1991, Jensen and Hornskov 1992, Robson and Davidson 1995); Lake Sebu (P. A. J. Morris, N. Higgins and several other observers *in litt*. 1990s); Mt Matutum at Tupi (three specimens in DMNH, FMNH); and Mt Sugarloaf at Burakan Hill (male in DMNH; also Dickinson *et al.* 1991). Although considered an uncommon, low-density species (Dickinson *et al.* 1991), and hence as threatened (Collar *et al.* 1994), recent evidence of its occurrence on Mt Kitanglad and on Mt Apo suggests that it is locally common (T. H. Fisher verbally 1997). **Criterion nearly met:** B1+2c.

WHITE-THROATED FLOWERPECKER *Dicaeum vincens* is found mainly in the lowland rainforests of the wet zone in south-west **Sri Lanka** (records from 95 localities: BirdLife database), at altitudes of up to c.900 m, although there have also been a few records in the drier forests in the intermediate zone and reports from higher altitudes (Legge 1880, Whistler 1944, IUCN/WCMC 1997, *Ceylon Bird Club Newsletter passim*). It was recently considered to be "common" in the low country wet zone (Kotagama and Fernando 1994), it was found to be "locally abundant" during surveys in 1997 (Jones *et al.* 1998), and during a major survey of over 200 forest sites in 1991–1996, it was recorded in 60 forests (IUCN/WCMC 1997). Nevertheless, suitable forest habitat is now highly fragmented within its small range (Hoffmann 1998) and it has presumably declined in numbers since it seems to be dependent on forest, even though it ranges outside it (Jones *et al.* 1998). However, it is likely that it currently numbers at least several tens of thousands of individuals. **Criteria nearly met:** B1+2a,b,c,d,e.

FLAME-CROWNED FLOWERPECKER *Dicaeum anthonyi* is endemic to the **Philippines**, where it occupies mossy forest above 800 m on Luzon at Mt Polis, Mt Puguis, Mt Tabuan and Dipalayag (race *anthonyi*), and on Mindanao at Mt Pasian, Daggayan, Mt Kitanglad, Mt Kampalili, Mt McKinley and Mt Apo (race *kampalili*) and at Mt Malindang (race *masawan*); it is in relatively secure habitat but appears to be a naturally low-density and hence very uncommon species (Morioka and Sakane 1979, Dickinson *et al*. 1991, Jensen and Hornskov 1992, Poulsen 1995). **Criteria nearly met:** B1+2c; C1; C2a.

RED-THROATED SUNBIRD *Anthreptes rhodolaema* is restricted to the Sundaic lowlands of southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sabah, Sarawak and Peninsular

Malaysia, Singapore, (extinct: Lim Kim Seng *in litt*. 2001), Kalimantan and Sumatra, **Indonesia**, **Brunei** (uncommon) and Palawan, **Philippines**, in broadleaf evergreen forest to 790 m (Smythies 1981, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's use of submontane and secondary forests and forest edge (Lekagul and Round 1991, Robson 2000) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

GREY-HOODED SUNBIRD *Aethopyga primigenius* is endemic to Mindanao, **Philippines**, where it is fairly common to common above 1,000 m in forest and forest edge including on Mt Hilong-hilong (race *diuatae*), Mt Kitanglad, Civolig, Daggayan, Mt Lamut, Mt Apo and Lake Sebu (nominate race); although its area of occupancy must be very small, its habitat is apparently relatively secure (Dickinson *et al.* 1991, R. J. Timmins *in litt.* 1994). **Criterion nearly met:** B1+2c.

LINA'S SUNBIRD *Aethopyga linaraborae* is a recently described endemic to Mindanao, **Philippines**, where it is currently known from (and is common in) mossy forest above 1,200 m on Mts Mayo, Puting Bato (Tagub) and Pasian in the eastern provinces of Davao del Norte and Davao Oriental; assuming that it occurs in all available remaining habitat above 1,200 m, its total range comprises 770 km^2, but it is unlikely to be seriously threatened at present or in the near future, occurring in rugged and inaccessible mountains that contain few commercial tree species and generally are too steep for agriculture (Kennedy *et al.* 1997). **Criterion nearly met:** B1+2c.

APO SUNBIRD *Aethopyga boltoni* is endemic to Mindanao, **Philippines**, where it is fairly common generally above 1,500 m in forest including on Mt Malindang (race *malindangensis*), Mt Kitanglad, Mt Apo, Mt Pasian and Lake Sebu (nominate *boltoni*), and Mt Busa and Mt Matutum (race *tibolii*), this latter taxon ranging on Busa from 1,300 m upwards and on Matutum from 820 m upwards; although its area of occupancy must be very small, its habitat is apparently relatively secure (Dickinson *et al.* 1991, Kennedy *et al*. 1997; also Jensen and Hornskov 1992, R. J. Timmins *in litt.* 1994, B. Gee *in litt.* 1997). It was recently found on Mt Parker (R. S. Kennedy *in litt.* 1998). **Criterion nearly met:** B1+2c.

JAVAN WHITE-EYE *Zosterops flavus* is restricted to the Javan Coastal Zone Endemic Bird Area and the Bornean Coastal Zone Secondary Area, occurring in mangroves, coastal scrub and relict coastal forest on Java and southern Kalimantan, **Indonesia**, with one record from Sarawak, **Malaysia** (Stattersfield *et al.* 1998; also MacKinnon and Phillipps 1993). The population on Java is apparently small (Allport and Milton 1988) but recent records from Kalimantan suggest the species might be locally common in mangrove on the south coast (Holmes and Burton 1987). **Criteria nearly met:** B1+2a,b,c,d,e.

PEARL-BELLIED WHITE-EYE *Zosterops grayi* is restricted to the Banda Sea Islands Endemic Bird Area, **Indonesia**, where it inhabits forest on Kai Kecil only (White and Bruce 1986, Stattersfield *et al.* 1998). It is moderately common (Coates and Bishop 1997), but it may be vulnerable to forest loss within its small range. **Criterion nearly met:** B1+2c.

GOLDEN-BELLIED WHITE-EYE *Zosterops uropygialis* is restricted to the Banda Sea Islands Endemic Bird Area, **Indonesia**, where it inhabits forest and cleared land with scattered trees on Kai Besar and Tual only (White and Bruce 1986, Stattersfield *et al.* 1998). It is moderately common (Coates and Bishop 1997), but it may be vulnerable to forest loss within its small range. **Criterion nearly met:** B1+2c.

BIAK WHITE-EYE *Zosterops mysorensis* is endemic to the twin islands of Biak-Supiori off Papua (formerly Irian Jaya), **Indonesia** (Beehler *et al*. 1986), where it occurs in forest up to 675 m, including forest edge, secondary forest, and even low secondary growth (Bishop 1982, K. D. Bishop *in litt*. 2000). It has been described as "abundant in small flocks" of 3–12 birds (Bishop 1982), but other recent visits have recorded it in small numbers only (N. Bostock *in litt.* 1993, Eastwood 1996, B. M. Beehler and SvB *in litt*. 2000, K. D. Bishop *in litt*. 2000,) or failed to find it altogether (Gibbs 1993, M. van Beirs *in litt*. 2000). On Biak, forest is under heavy threat from logging and subsistence farming, but there appear to be large areas of forest remaining in interior Supiori (Bishop 1982, K. D. Bishop *in litt*. 1996, D. A. Holmes *in litt.* 2000). Its reported tolerance of degraded forest, and its occurrence in the 110 km^2 Biak-Utara protected area (B. M. Beehler and SvB *in litt*. 2000), suggest that it is not threatened. **Criteria nearly met:** C1; C2a.

AMBON YELLOW WHITE-EYE *Zosterops kuehni* is restricted to the Seram Endemic Bird Area, **Indonesia**, where it inhabits lowland forest, scrub and gardens at 0–500 m (Stattersfield *et al*. 1998), or remnant patches of secondary forest and woodland and lightly wooded cultivation up to 300 m (Coates and Bishop 1997); wooded habitats are diminishing on this densely populated and relatively small island (Collar and Andrew 1988). **Criterion nearly met:** C1.

SPOT-BREASTED WHITE-EYE *Heleia muelleri* is restricted to the Timor and Wetar Endemic Bird Area, **Indonesia**, where it occurs on Timor at 0–1,300 m in monsoon forest (Stattersfield *et al*. 1998). It is uncommon and local, occasionally moderately common (Coates and Bishop 1997), and appears to favour lower-lying areas with closed-canopy forest (Collar and Andrew 1988, P. Andrew and R. Noske *in litt*. 2000), which are constantly diminishing (see Threats under Wetar Ground-dove *Gallicolumba hoedtii*). **Criteria nearly met:** A1c; A2c; C2b.

CRIMSON-HOODED MYZOMELA *Myzomela kuehni* is restricted to the Timor and Wetar Endemic Bird Area, **Indonesia**, where it occurs only on Wetar in lowland monsoon forest and gardens, with a recent observation in a mosaic of coastal scrub, overgrown cultivation and tall secondary woodland (Stattersfield *et al*. 1998; also Coates and Bishop 1997). It remains little known and apparently scarce (Coates and Bishop 1997) and, although extensive forest still remains on the island, there is now some logging occurring and there have been plans (recently set aside) to relocate Indonesian East Timorese there (see Threats under Wetar Ground-dove *Gallicolumba hoedtii*). **Criteria nearly met:** C1; C2b.

BLACK-CHESTED HONEYEATER *Lichmera notabilis* is restricted to the Timor and Wetar Endemic Bird Area, **Indonesia**, where it occurs only on Wetar in lowland monsoon forest and gardens, with a recent observation in a mosaic of coastal scrub, overgrown cultivation and tall secondary woodland (Stattersfield *et al*. 1998; also Coates and Bishop 1997). It remains little known and apparently uncommon (Coates and Bishop 1997) and, although extensive forest still remains on the island, there is now some logging occurring and there have been plans (recently set aside) to relocate Indonesian East Timorese there (see Threats under Wetar Ground-dove *Gallicolumba hoedtii*). **Criteria nearly met:** C1; C2b.

BRASS'S FRIARBIRD *Philemon brassi* is endemic to northern Papua (formerly Irian Jaya), **Indonesia**, where it was discovered in 1939 on a single lagoon on the Idenburg river which may not have been visited subsequently (Rand 1940, Beehler 1985, J. M. Diamond *in litt* 1987), and has recently been found along the lower Mamberamo river (B. M. Beehler *in litt*. 1990, Stattersfield *et al*. 1998), the Tirawiwa and Logari rivers, 250 km from the Idenburg site (Mack and Alonso 2000), and probably on the Rouffaer river (K. D. Bishop *in litt*.

2000). On the Idenburg river it was common in small parties in flooded cane grass and dense second growth around a lagoon at c.50 m (Rand 1940), but on the Tirawiwa and Logari rivers it was locally common in trees beside the rivers and other disturbed areas at 80–275 m (Mack and Alonso 2000). It may be threatened by various large-scale timber and agricultural schemes and a proposed dam on the Mamberamo gorge (Sujatnika *et al.* 1995), but all the known populations are currently safe, much lowland forest and floodplains of the Mamberamo and Idenburg rivers is encompassed within the c.10,000 km^2 Foja Nature Reserve (Stattersfield *et al.* 1998), and the region remains largely inaccessible and undisturbed. **Criteria nearly met:** B1+2b,c; C2a; D2.

TIBETAN BUNTING *Emberiza koslowi* is restricted to the eastern part of the Qinghai–Tibetan Plateau, in eastern Tibet and southern Qinghai, mainland **China**, where several large valleys cut into the plateau, including those of the Tongtian river (the upper reaches of the Chang Jiang or Yangtze), the Lancang Jiang (the upper Mekong) and the Nu Jiang (the upper Salween) (Cheng Tso-hsin 1987, Stattersfield *et al.* 1998). It is known by just a few scattered records in this inaccessible and poorly known area, but it appears to be genuinely rather scarce and localised (Robson 1986, P. Alström *in litt.* 1993, Olsson 1995), in barren areas, juniper and rhododendron scrub on valley slopes above the treeline at c.3,600–4,600 m (Olsson 1995, Stattersfield *et al.* 1998). It does not appear to be immediately threatened by habitat loss (P. Alström *in litt.* 1993). **Criteria nearly met:** C1; C2a.

OCHRE-RUMPED BUNTING *Emberiza yessoensis* breeds in wetlands with tall grass and scrub in Primorye in extreme south-east **Russia**, Honshu, Kyushu and formerly Hokkaido, **Japan**, and Heilongjiang in north-east mainland **China**, and it is a passage and/or winter visitor to **North Korea** (where it is also likely to breed), **South Korea** and the coast of eastern China; it is considered to be uncommon or rare in all parts of its range (Austin 1948, Dement'ev and Gladkov 1951–1954, Gore and Won 1971, Cheng Tso-hsin 1987, Brazil 1991). It is presumably declining because of the loss and degradation of wetland habitat within its breeding range (see Scott 1989, MacKinnon *et al.* 1996, Threats under Red-crowned Crane *Grus japonensis*). **Criterion nearly met:** C1.

VIETNAM GREENFINCH *Carduelis monguilloti* is endemic to the Da Lat plateau of south Annam, **Vietnam**, where it is locally common in open pine forest from 1,050–1,900 m (Robson 2000). Much of the Da Lat plateau is being deforested through clearance for shifting agriculture, charcoal production and logging (Eames 1995a). However, this species may have actually benefited from the increase in *Pinus kesiya* pine forest (as its growth is stimulated by fire clearance) and is therefore perhaps secure (Eames 1995a). Most pine forest is nevertheless allocated to production forest where logging is permitted and the potential for rapid clearance of this habitat remains (Stattersfield *et al.* 1998). **Criteria nearly met:** A2b,c; B1+2a,b,c,d,e.

RED-EARED PARROTFINCH *Erythrura coloria* is endemic to Mindanao in the **Philippines**, where until very recently it was documented as from only two or three localities (in Dickinson *et al.* 1991, Collar *et al.* 1994) but is now known from as many as seven, as follows: Mt Hilong-hilong at Siwod (breeding male in USNM); Mt Pasian, Bislig (female in CMNH); Mt Puting Bato, Davao del Norte (male in CMNH); Mt Kitanglad at Malaybalay (Ripley and Rabor 1961, Ziswiler *et al.* 1972), at Mt Kaatoan (12 specimens in AMNH, BMNH, CM, DMNH, FMNH), at Mt Nangkabulos (Gonzalez and Mallari 1993, Heaney *et al.* 1993), and at Kinubalan (CMNH register data), with unspecified sublocalities (Lambert 1993c, Evans *et al.* 1993a, many observers *in litt.*); Mt Piapayungan at Siwagat (male in UPLB); Mt Apo (Dickinson *et al.* 1991; recent specimens in CMNH); Mt Parker, T'boli (seven specimens in CMNH, PNM); Mt Three Kings (male in PNM); and Mt Busa on the south slope at

Binati (including one record from Mt Cabaay) (11 specimens in CMNH). The species inhabits forest understorey and edge, second growth and grassy clearings over 1,000 m, habitats which are under relatively low pressure, and it is described as moderately common but very local (Ziswiler *et al.* 1972, Dickinson *et al.* 1991), although common inside and near forest, probably being present on every mountain in central Mindanao (T. H. Fisher verbally 1997). It is very unobtrusive and secretive, and its high-pitched call (typical of the genus: Clement *et al.* 1993) is easily overlooked (P. A. J. Morris *in litt.* 1996); on voice, however, it appears to be not uncommon at Kitanglad (Evans *et al.* 1993a). Unlike the Green-faced Parrotfinch *E. viridifacies*, it does not appear to be strongly allied to bamboo (D. Allen verbally 1997), and it is therefore less irruptive, unpredictable and inherently vulnerable than that species. **Criterion nearly met:** B1+2c.

BLACK MUNIA *Lonchura stygia* is known from a small area of the Trans-Fly region of New Guinea, from Mandum (Papua, formerly Irian Jaya, **Indonesia**) to Lake Daviumbu (**Papua New Guinea**); it is reportedly locally common in flocks of up to 20 birds, inhabiting savannas, marshes and riverine grasses, often on floating islands, but sometimes visiting rice crops (Beehler *et al.* 1986, Coates 1990, Gregory 1995b). It is, however, less common than the largely sympatric Grey-crowned Munia *L. nevermanni* and only one was seen in four months of fieldwork in its Papuan range (K. D. Bishop *in litt.* 1987). It may be threatened as a result of destruction of reedbeds by introduced rusa deer *Cervus timorensis* and by encroachment of woodland on grasslands, possibly promoted by the activities of pigs (although woodlands might also represent suitable habitat for this species) (N. Stronach *in litt.* 1993, 1994). In the dry season, birds concentrate around drinking pools and are susceptible to trapping for the Indonesian cagebird trade; 250 were being exported from Merauke airport in August 1993 (N. Bostock *in litt.* 1993). **Criteria nearly met:** A1c,d; A2c,d; C1.

ASIAN GOLDEN WEAVER *Ploceus hypoxanthus* occurs in **Myanmar** (scarce to locally common in the centre, south, south-west and Tenasserim: Robson 2000), **Thailand** (local and uncommon in the north and centre: Lekagul and Round 1991, Robson 2000), **Laos** (local in the south: Round 1999, Robson 2000), **Vietnam** (scarce in south Annam and Cochinchina: Robson 2000), **Cambodia** (scarce to locally common: Robson 2000) and **Indonesia** (rare resident in Sumatra, previously widespread but now local and uncommon on Java, restricted to the west: MacKinnon and Phillips 1993). Inhabits marshes, grassland, reeds and rice paddies, always close to water in the lowlands where large flocks sometimes form around colonies (MacKinnon and Phillipps 1993, Robson 2000). It is threatened by conversion of wetland habitat to agriculture and also by persecution: it is commonly trapped (particularly in Java) and colonies are robbed and destroyed (JAT). **Criteria nearly met:** A1b,d; A2b,d; C1.

TANIMBAR STARLING *Aplonis crassa* is restricted to the Banda Sea Islands Endemic Bird Area, **Indonesia**, where it inhabits forest and mangroves on Tanimbar only (Stattersfield *et al.* 1998). It is common within its small range, but there is significant logging in the south of Yamdena (Bishop and Brickle 1998; see Threats under Lesser Masked-owl *Tyto sororcula*), so the species is likely to be in steady decline. **Criterion nearly met:** B1+2c.

YELLOW-EYED STARLING *Aplonis mystacea* has a scattered range across New Guinea (Papua, formerly Irian Jaya, **Indonesia** and **Papua New Guinea**) but is easily overlooked and may prove to be more widespread (Beehler *et al.* 1986, Coates 1990, Burrows 1993, Beehler and Bino 1995). It is a generally scarce and probably nomadic species of lowland forest, usually in riverine or alluvial lowlands, but also in hills up to 580 m (Beehler and Bino 1995). The one known breeding colony (of c.200 birds) was in a sparsely populated area, but the tree was cut down to eat the eggs and chicks (Safford 1996). This species may have a naturally

small and scattered population threatened locally by logging and hunting. **Criteria nearly met:** C1; C2a.

HELMETED MYNA *Basilornis galeatus* is restricted to the Banggai (Peleng, Banggai) and Sula Islands (Taliabu, Seho and Mangole) Endemic Bird Area, **Indonesia**, where it inhabits lowland and montane forest and tall trees in reedswamps at 0–1,000 m (Stattersfield *et al*. 1998; also Coates and Bishop 1997). While it may be relatively secure in montane regions and it occurs in the Banggai Islands in secondary forest (Indrawan *et al*. 1997b), in the Sula lowlands it is uncommon and increasingly vulnerable to the loss of the least disturbed forest, its preferred habitat (Davidson *et al*. 1995, Stones *et al*. 1997). **Criteria nearly met:** C1; C2a.

APO MYNA *Basilornis miranda* is endemic to Mindanao, **Philippines**, where it is common within a highly circumscribed area of occupancy above 1,250 m in forest and forest edge, even in cut-over areas, including at Daggayan, Mt Kitanglad and Mt Apo; thus despite a highly restricted range, its numbers are apparently relatively secure and stable (Dickinson *et al.* 1991, N. J. Redman verbally 1993). **Criterion nearly met:** B1+2c.

BARE-EYED MYNA *Streptocitta albertinae* is restricted to the Banggai and Sula Islands Endemic Bird Area (Taliabu and Mangole in the Sulas only), **Indonesia**, where it uses tall trees in lowland forest and agricultural land at 0–250 m (Stattersfield *et al*. 1998). It appears to be more tolerant of degraded forest than Helmeted Myna *Basilornis galeatus*, but it is uncommon and has a smaller total and narrower elevation range (Davidson *et al*. 1995, Stones *et al*. 1997). **Criteria nearly met:** C1; C2a.

SRI LANKA MYNA *Gracula ptilogenys* is endemic to the wet zone of **Sri Lanka**, where it is common to very common in lowlands and hills wherever forest persists (Kotogama and Fernando 1994, Grimmett *et al.* 1998, Jones *et al.* 1998). It prefers natural forest and well-wooded country, although visits gardens and plantations if forest is nearby and appears to tolerate habitat degradation (Grimmett *et al.* 1998, Jones *et al.* 1998). Forest on the island has suffered rapid degradation and fragmentation in the past decades through excessive gathering of fuelwood, clearance for permanent agriculture, shifting cultivation, fire, urbanisation and logging (Stattersfield *et al.* 1998). Closed-canopy forest is estimated to have declined from 29,000 km^2 (44% of the island's area) in 1956 to 12,260 km^2 in 1983 (Collins *et al.* 1991). **Criteria nearly met:** B1+2a,b,c,d,e.

DARK-THROATED ORIOLE *Oriolus xanthonotus* occurs in the Sundaic lowlands, from southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct), Kalimantan, Sumatra (including Mentawai and Bangka islands) and Java, **Indonesia**, **Brunei** and the Calamian islands and Palawan, south-west **Philippines**, where it is fairly common and widespread in evergreen forest (including swamp forest) to 1,220 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's use of submontane and secondary forests (MacKinnon and Phillipps 1993) and forest edge (Robson 2000) implies that it is not unduly threatened. **Criteria nearly met:** A1c; A2c.

BLACK ORIOLE *Oriolus hosii* is restricted to the Bornean Mountains Endemic Bird Area, in montane forest at 900–2,000 m (Stattersfield *et al*. 1998). It is apparently endemic to Sarawak, **Malaysia**, with records from Murud, Derian, Dulit, Batu Patap, Mt Kalulong and

Mt Mulu (Sharpe 1893–1894, Banks 1937b, Fogden 1965). It is specialised to moss transitional forest around 1,100–1,200 m on the sandy north side of the Dulit range (Harrisson 1933, Smythies 1981) and seems likely to be declining in some areas as habitat loss encroaches into montane areas (see Threats under Mountain Serpent-eagle *Spilornis kinabaluensis*). **Criterion nearly met:** C1.

WETAR FIGBIRD *Sphecotheres hypoleucus* is restricted to the Timor and Wetar Endemic Bird Area, **Indonesia**, where it occurs only on Wetar in lowland monsoon forest and scrub, with a recent observation in coastal secondary monsoon woodland, disturbed habitat and lightly wooded scrub (Stattersfield *et al*. 1998; also Coates and Bishop 1997). It remains little known but probably moderately common (Coates and Bishop 1997); however, although extensive forest still remains on the island, there is now some logging occurring and there have been plans (recently set aside) to relocate Indonesian East Timorese there (see Threats under Wetar Ground-dove *Gallicolumba hoedtii*). **Criteria nearly met:** C1; C2b.

SUMATRAN DRONGO *Dicrurus sumatranus* has been described as restricted to the Sumatra and Peninsular Malaysia Endemic Bird Area, occurring only in Sumatra (with race *viridinitens* on the Mentawai Islands), **Indonesia**, in lowland, hill and lower montane forest up to 800 m, and perhaps to 1,500 m (Stattersfield *et al.* 1998; also van Marle and Voous 1988). However, it is also described as a bird exclusively of lowland Sumatra (Sibley and Monroe 1990), and indeed it has been recorded (rarely) in Way Kambas National Park (highest point 16 m) (Parrott and Andrew 1996), but appears to be commoner at mid-altitudes where it is a fairly common member of mixed species foraging parties (P. A. J. Morris *in litt*. 2000). Loss of lowland forest in Sumatra has been extensive (see Threats under Crestless Fireback *Lophura erythrophthalma* and Rueck's Blue-flycatcher *Cyornis rucki*) and this species may be compromised as a consequence. **Criteria nearly met:** A1c; A2c.

ANDAMAN DRONGO *Dicrurus andamanensis* is endemic to the Andaman archipelago, **India**, and Coco Island, **Myanmar**, where it is a common resident of forests (Davidar *et al.* 1996, Grimmett *et al.* 1998). Although its range is very small, forested habitat is relatively intact on the Andamans and insufficiently disturbed or fragmented to be of immediate concern. However, there are signs that pressure on forests is increasing in the Andamans through increasing human populations and consequent conversion of habitat to cultivation, grazing, increased logging and development (Whitaker 1985, Curson, 1989, Pande *et al.* 1991, Sinha 1992, Stattersfield *et al.* 1998). **Criteria nearly met:** B1+2a,b,c,d,e.

ARCHBOLD'S BOWERBIRD *Archboldia papuensis* is very patchily distributed in the Central Ranges of New Guinea, from the Weyland mountains of Papua (formerly Irian Jaya), **Indonesia**, to the highlands provinces of **Papua New Guinea**, with as few as seven confirmed localities in this wide range (Frith *et al.* 1995). It is locally moderately common, inhabiting forests, often with *Pandanus*, on high frost-disturbed plateaus from 2,300–3,660 m, rarely as low as 1,750 m (Beehler *et al.* 1986, Coates 1990); however, subpopulation sizes and inter-site dispersal is unknown, and it may be at risk of population fragmentation (A. Mack *in litt.* 1999). It is threatened by logging on two mountains in its tiny range of only c.800 km^2 in Papua New Guinea (Beehler 1985, Collar 1986) but the larger Papuan populations are judged to be safe. ***Taxonomy*** Although the eastern race *sanfordi* has been treated as a full species, the Tomba Bowerbird (e.g. Collar 1986), scrutiny of the evidence suggests it is better treated as conspecific with the western race *papuensis* (Frith *et al.* 1995). **Criteria nearly met:** B1+2c,e; C2a.

YELLOW-BREASTED BIRD-OF-PARADISE *Loboparadisea sericea* is patchily distributed along the Central Ranges of New Guinea (Papua, formerly Irian Jaya, **Indonesia**, and **Papua**

New Guinea). It is locally common, for instance, in primary montane forest 1,400–2,000 m on Crater mountain (A. Mack *in litt.* 1999), but often uncommon or absent from seemingly appropriate habitats for unknown reasons, and may be split into isolated subpopulations. However, it may be overlooked because of its unobtrusive habits and rugged, rarely visited habitat. It inhabits montane forest, rarely visiting secondary habitats, from 600 to 2,000 m but mostly above 1,200 m (K. D. Bishop *in litt.*1994, Frith and Beehler 1998). Although locally threatened in sites such as the Ok Tedi mine (P. Gregory *in litt.* 1994), the majority of its range is safe from logging and large-scale habitat degradation (K. D. Bishop *in litt.*1994, Frith and Beehler 1998). **Criteria nearly met:** C1; C2a.

LONG-TAILED PARADIGALLA *Paradigalla carunculata* is endemic to the Arfak mountains in the Vogelkop Peninsula of Papua (formerly Irian Jaya), **Indonesia**, where there are few records but it is probably not uncommon (Gibbs 1993, Poulsen and Frolander 1994, Frith and Beehler 1998). An undescribed *Paradigalla* taxon discovered recently in the nearby Fakfak mountains, and possibly occurring on other mountain ranges in the Vogelkop, may be this species, although it shows some features of Short-tailed Paradigalla *P. brevicauda* (Gibbs 1994, Frith and Beehler 1998). It inhabits montane forest and forest edge from 1,400 to 2,100 m (Frith and Beehler 1998). The extensive forests within its small range remain largely undisturbed owing to their geographical isolation and the low density and traditional lifestyle of the human population, although deforestation is occurring in the hills (Sujatnika *et al*. 1995). There is one large protected area in the Arfak mountains, which has been proposed for extension (Sujatnika *et al*. 1995). This species may be declining locally through forest loss but is probably secure in most of its range. **Criterion nearly met:** C2a.

PALE-BILLED SICKLEBILL *Epimachus bruijnii* is a poorly known species which ranges along the north New Guinea coast from the south-east coast of Geelvink Bay, Papua (formerly Irian Jaya), **Indonesia**, east to Vanimo, just across the border into **Papua New Guinea**. It is widespread and usually common within this fairly small range, with one male ranging over 15 ha in a week-long study. It is also relatively common in selectively logged forest, but most records are from forests below 180 m (Beehler and Beehler 1986, N. Bostock *in litt.* 1994, Frith and Beehler 1998) which are under pressure for development and timber schemes (Sujatnika *et al.* 1995). **Criterion nearly met:** C1.

WILSON'S BIRD-OF-PARADISE *Cicinnurus respublica* is endemic to the Papuan islands of Waigeo and Batanta off north-west Papua (formerly Irian Jaya), **Indonesia**, where it is reported to be "frequent" in hill forest, generally above 300 m, although it is often heard even at low altitude (Beehler *et al*. 1986, Gibbs 1993, Poulsen and Frolander 1994, Eastwood 1996). Waigeo's rugged relief and lack of infrastructure suggest that there may be no serious immediate threats to its forests, and the Pulau Waigeo nature reserve, established in the late 1980s, covers 1,530 km^2, but there are reports that it may be substantially reduced in size (Holmes 1989, Dekker and McGowan 1995). Selective logging has been reported in the north of Waigeo, the south-east corner of the island was ravaged by fire in 1982, and there are concerns over a cobalt mining concession on the island (Dekker and Argeloo 1993, WWF-IUCN 1994–1995, Dekker and McGowan 1995, Sujatnika *et al*. 1995). Logging on Batanta (where the only protected area is only 100 km^2) is resulting in major habitat degradation but this species appears to be moderately common in logged forest and it is probably safe at higher altitudes (Frith and Beehler 1998). As well as declining slowly through habitat loss, it is hunted in some areas for skins (D. Gibbs *in litt*. 2000). **Criterion nearly met:** C1.

RED BIRD-OF-PARADISE *Paradisea rubra* is endemic to the West Papuan islands of Waigeo, Batanta, Gemien and Saonek off north-west Papua (formerly Irian Jaya), **Indonesia**,

where it is common in forest and forest edge in the lowlands and hills, up to 600 m (Beehler *et al*. 1986, Gibbs 1993, Poulsen and Frolander 1994, Eastwood 1996). Waigeo's rugged relief and lack of infrastructure suggest there may be no serious immediate threats to its forests, and the Pulau Waigeo nature reserve, established in the late 1980s, covers 1,530 km^2, but there are reports that it may be substantially reduced in size (Holmes 1989, Dekker and McGowan 1995). Selective logging has been reported in the north of Waigeo, the south-east corner of the island was ravaged by fire in 1982, and there are concerns over a cobalt mining concession on the island (Dekker and Argeloo 1993, WWF-IUCN 1994–1995, Dekker and McGowan 1995, Sujatnika *et al*. 1995). Logging on Batanta (where the only protected area is only 100 km^2) is resulting in major habitat loss. As well as possibly declining from habitat degradation, it is hunted locally for skins and possibly cagebirds (Frith and Beehler 1998, D. Gibbs *in litt*. 2000). **Criterion nearly met:** C1.

CRESTED JAY *Platylophus galericulatus* is restricted to the Sundaic lowlands, from southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Kalimantan, Sumatra and Java, **Indonesia** and **Brunei**, where it is generally uncommon in evergreen forest to 1,500 m (Smythies 1981, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but the species's use of submontane forest (MacKinnon and Phillipps 1993) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

BLACK MAGPIE *Platysmurus leucopterus* occurs in the Sundaic lowlands, from southern Tenasserim, **Myanmar**, peninsular **Thailand**, Sabah, Sarawak and Peninsular **Malaysia**, Singapore (extinct: Lim Kim Seng *in litt*. 2001), Kalimantan and Sumatra (including Bangka) and Java, **Indonesia** and **Brunei** (widespread but uncommon), in evergreen forest and mangrove to 800 m (Smythies 1981, Mann 1987, Lekagul and Round 1991, MacKinnon and Phillipps 1993, Robson 2000). Forest destruction in the Sundaic lowlands of Indonesia has been so extensive that all primary formations are expected to disappear by 2010, and the situation is little different in Thailand and Malaysia (see Threats under Crestless Fireback *Lophura erythrophthalma*), but this bird's use of hill forests (MacKinnon and Phillipps 1993) and forest edge (Robson 2000) implies that it is not immediately threatened. **Criteria nearly met:** A1c; A2c.

ANDAMAN TREEPIE *Dendrocitta bayleyi* is endemic to the Andaman archipelago, **India**, where it is usually found in pairs or parties of up to 20 birds, or in mixed flocks in tall trees in dense broadleaved evergreen forest (Grimmett *et al.* 1998). It is uncommon (Davidar *et al.* 1996) to locally fairly common (Grimmett *et al.* 1998), and although habitat on the Andamans remains relatively intact, there are indications that an increase in human populations and habitat loss is occurring in the archipelago suggesting that the very small range of this species might rapidly shrink and fragment (Whitaker 1985, Curson, 1989, Pande *et al.* 1991, Sinha 1992, Stattersfield *et al.* 1998). **Criteria nearly met:** B1+2a,b,c,d,e.

HOODED TREEPIE *Crypsirina cucullata* is endemic to the dry zone of central **Myanmar**, on the plains of the Irrawaddy and Sittang rivers, where it is found in dry dipterocarp forest, dry thornscrub, secondary growth and the edge of agricultural land in the lowlands to 1,000 m (King *et al.* 1975, Smythies 1986, Robson 2000). It was formerly common (Smythies 1986), but the forests of the Irrawaddy plains are now almost entirely cleared for agriculture (Collins *et al.* 1991, Stattersfield *et al.* 1998) and a sharp decline is thought to have taken place (Collar

et al. 1994). However, recent surveys have found it locally common in large areas of suitable habitat (Robson *et al.* 1998, Robson 2000). **Criteria nearly met:** C1; C2a.

XINJIANG GROUND-JAY *Podoces biddulphi* is known from Xinjiang, western mainland **China**, where it is found in sandy desert, scrub and desert poplar in the Taklimakan Desert (Cheng Tso-hsin 1987, Grimmett 1991), with a recent sight record well to the east of this area, near Golmud in Qinghai (Turton and Speight 1986). It was described as being common in 1929–1930 (Ludlow and Kinnear 1933–1934), but scarce and difficult to locate in the same areas in 1988, and it may be declining with the degradation of desert habitats through the intensive grazing of goats and camels, extraction of fuelwood and the conversion of huge areas to irrigated land (Grimmett 1991, Grimmett and Taylor 1992). However, it has recently been found to be widespread and locally common in the interior of the Taklimakan Desert (Ma Ming 1998). **Criteria nearly met:** C1; C2a.

BROWN-HEADED CROW *Corvus fuscicapillus* is endemic to eastern **Indonesia**, where its known distribution is highly fragmented, presumably related to some unknown habitat specialisation. There are records from the lower Mamberamo river and Nimbokrang (near Jayapura) in northern Papua (formerly Irian Jaya), where it may prove to be more widespread, Waigeo and Gemien in the West Papuan islands, and the Aru islands (Beehler *et al*. 1986, Gibbs 1993, Diamond and Bishop 1994, Eastwood 1996). It is widespread, but in low numbers, on the Aru islands (Diamond and Bishop 1994), and quite common at Nimbokrang (Gibbs 1993). It inhabits primary forest, but also mangroves and occasionally second growth, but rarely open habitats and never on the coast or outlying islands, in lowlands and hills up to 500 m (Beehler *et al*. 1986, Diamond and Bishop 1994). Forest within its range is threatened in places by logging, a cobalt mining concession on Waigeo, and a dam proposed across the Mamberamo river, although much of the forest remains intact and relatively secure, and includes some protected areas (WWF-IUCN 1994–1995, Dekker and McGowan 1995, Sujatnika *et al*. 1995, K. D. Bishop *in litt*. 1996). Although the paucity of records suggests that this species may be rare and locally declining, it is judged to be safe in the large areas of forest without any immediate threats within its range. **Criteria nearly met:** C1; C2a.